

1 **Machine learning reveals biological activities as the dominant**
2 **factor in controlling deoxygenation in the South Yellow Sea**

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21 bloom; Random Forest

22 **Abstract:**

23 Dissolved oxygen (DO) is a crucial element for both biotic and abiotic processes
24 in marine ecosystems, but has declined globally in recent decades. Therefore, there is
25 an urgent need for solid large-scale and continuous estimation of DO concentration in
26 vital ecosystems, such as coastal areas. A random forest (RF) model for DO in South
27 Yellow Sea (SYS) was developed by integrating satellite data and simulation data
28 during 2011–2019. The root mean squared error (RMSE) for the training and test sets
29 were 0.514 mg/L and 0.732 mg/L, respectively. Spatiotemporal distributions of DO of
30 multiple layers in the study area during 2011–2019 were very well reproduced by the
31 RF model and showed a slight decline trend in most SYS areas, while more intense
32 decline occurred in the deep central SYS. The analysis of the mechanisms of DO
33 decline in the South Yellow Sea cold water mass (SYSCWM), located in the deep
34 central SYS, indicates that the deoxygenation here is largely due to biological activities.
35 This finding may have implications for studies on drivers of deoxygenation in coastal
36 areas. Furthermore, integrating satellite data with machine learning models can offer a
37 powerful approach to capturing the continuous spatiotemporal characteristics of ocean
38 parameters over large spatial scales.

39 **1. Introduction**

40 Dissolved oxygen (DO) concentration in seawater is a crucial parameter that
41 impacts both biotic and abiotic processes. It influences the growth, reproduction, and
42 feeding of organisms (Batziakas et al., 2020; Flint et al., 2015; Fock and Czudaj, 2019),

43 while also playing a crucial role in regulating biogeochemical cycles involving carbon,
44 nitrogen, and many other elements (Banks et al., 2012; Codispoti et al., 2001; Levin,
45 2018; Lonborg et al., 2020; Mathew et al., 2022). However, global warming has led to
46 the heating of seawater and stronger stratification; in the meantime, rising nutrient loads
47 resulted in increased productivity in the upper water column. All of this caused
48 increased oxygen consumption, leading to “deoxygenation” in the ocean (Breitburg et
49 al., 2018; Fennel and Testa, 2019; Keeling and Garcia, 2002). The DO content in the
50 global ocean has declined by over 2% since 1960 (Schmidtko et al., 2017), with coastal
51 waters experiencing worse deoxygenation than the open ocean in recent decades
52 (Gilbert et al., 2010). This decline in DO at the coastal water poses a potential threat to
53 marine ecosystem health, which could ultimately impact not only the well-being of
54 marine life but also the livelihoods of approximately 10-12% of the global population
55 who rely directly on marine resources for their survival and economic activities
56 (Wenning, 2020). Despite growing attention to the issue, there is a strong lack of
57 continuous dataset to help fully understand its underlying mechanisms and quantify its
58 declining rates, especially in the marginal seas.

59 Traditionally, DO concentrations have been measured using titration facilitated by
60 cruises (Bushinsky et al., 2016; Edwards et al., 2010). However, this approach is labor-
61 intensive and costly, and is also subject to weather conditions. Even with improved
62 accessibility, Argo floats face challenges in achieving large-scale, continuous
63 observations. In this context, numerical simulations can be valuable, especially when a

64 significant amount of *in-situ* data have been collected. At present, numerical simulation
65 methods used for DO estimation include process-based (mechanistic) models and data-
66 driven models. Process-based models, such as high-resolution numerical simulation
67 models, can effectively reproduce real-world scenarios through extensive calculations
68 across the entire spatial extent of the study area (Scully, 2013; Xu et al., 2011). Even
69 though such models may provide higher accuracy, they demand extensive
70 multidisciplinary knowledge of the specific study area and are not easily adjustable. By
71 contrast, data-driven models have emerged as a promising approach. They demonstrate
72 significant potential for high efficiency in various fields (Beyan and Browman, 2020;
73 Chen et al., 2020; Goldstein et al., 2019; Malde et al., 2020; Pahlevan et al., 2022;
74 Rubbens et al., 2023; Xiao et al., 2019). When combined with satellite data at high
75 spatiotemporal resolution, these models have been proven effective in modeling DO
76 concentrations (Sadaiappan et al., 2023). For example, Guo et al. (2021) used support
77 vector regression (SVR) to predict DO concentration at the surface of lakes. Li et al.
78 (2023b) employed multiple machine learning models to reproduce hypoxia events in
79 the bottom of the Gulf of Mexico. All of this shows that machine learning performs
80 well at DO simulation. If machine learning methods can be integrated with some
81 marginal seas where DO is at risk of declining, it may offer enlightening information.

82 The South Yellow Sea (SYS) is a significant marginal sea in the western Pacific
83 Ocean, providing crucial ecological and economic services (Barbier et al., 2011; Dong,
84 2019; Hou et al., 2020; Long et al., 2023; Yu et al., 2022). However, it is experiencing

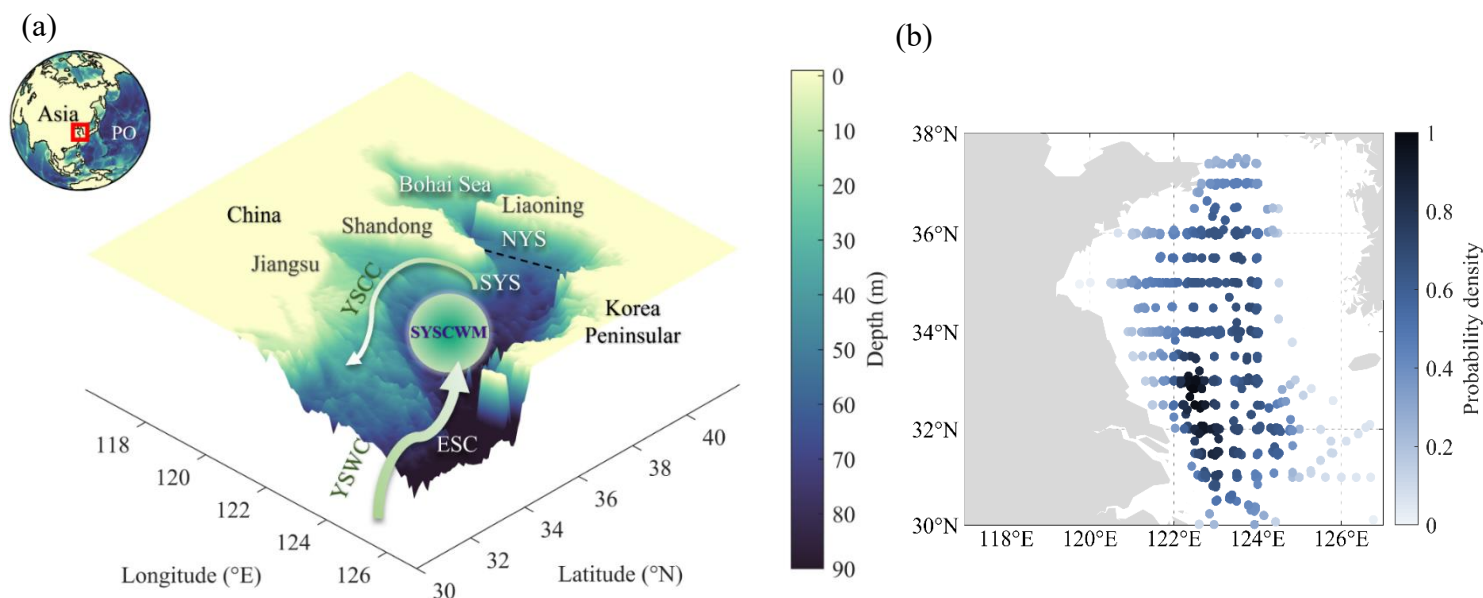
85 local deoxygenation (Lin et al., 2005; Wei et al., 2021), and this has a significant impact
86 on the marine ecosystem. The South Yellow Sea Cold Water Mass (SYSCWM) in the
87 SYS plays a vital role in its hydrodynamics and significantly affects its primary
88 production (Guo et al., 2020a; Li et al., 2019), maintaining high DO concentration (> 6
89 mg/L) year around (Xin et al., 2013; Zhang et al., 2008). Although recent reports
90 suggest warming in the SYSCWM (Yang et al., 2023), the trend in DO concentration
91 remains uncertain. Moreover, it is unknown whether deoxygenation is taking place in
92 the entire SYS and what its driving mechanisms. Existing studies on deoxygenation
93 mechanisms often focus on solubility changes, with less emphasis on the role of
94 biological processes (Oschlies et al., 2018). Thus, we aim to gather large-scale DO data
95 alongside biomass data from satellite observations to explore the mechanisms involved.
96 Using machine learning to generate such DO data is a suitable choice.

97 Therefore, we developed a machine learning model for reproducing DO
98 concentration in the SYS during 2011–2019. We evaluated four models and 15 sets of
99 inputs, identified the optimal model, and applied it to generate a high-resolution (1/12°)
100 multi-layer map of DO concentration in the SYS. Through analyzing the model outputs,
101 this study aims to determine the spatial and temporal distributions of DO in the SYS,
102 assess whether DO levels are declining, especially in the SYSCWM region, and
103 understand the roles of warming and biological activities in the DO change. This study
104 may have implications for studies on drivers of deoxygenation in the SYS and other
105 coastal areas.

106 **2. Data and Methodology**

107 **2.1 Study area**

108 The study area includes the SYS and parts of the north East China Sea, specifically
109 ranging from the northern boundary of the SYS (marked by a black dashed line in [Fig.](#)
110 [1a](#)) to 30°N. The SYS is a semi-enclosed marginal sea bordered by the North Yellow
111 Sea to the north and the East China Sea to the south, constituting a part of the Northwest
112 Pacific Ocean. It has an average depth of 44 m ([Liu et al., 2009](#)). The coastal areas of
113 Jiangsu and Shandong provinces are of great significance for coastal aquaculture, while
114 most of the SYS shows potential for deep-sea aquaculture ([Hou et al., 2020](#); [Yu et al.,](#)
115 [2022](#)). The prevailing ocean currents in this region follows a counterclockwise pattern.
116 There is the strong northwestward YSWC in the central area and southward YSCC
117 along the western coasts ([Naimie et al., 2001](#)). This circulation pattern results in
118 relatively high temperature in the central SYS during winter. In the deep central SYS,
119 the SYSCWM undergoes seasonal formation, weakening, and disappearance from
120 summer to winter.



121 **Fig. 1** (a) Geographic location and topography of the study area (NYS: North Yellow Sea; ECS:
122 East China Sea; YSWC: Yellow Sea Warm Current; YSCC: Yellow Sea Coastal Current) and (b)
123 Distribution of *in-situ* DO samples by horizontal location.

124 2.2 Data

125 2.2.1 In-situ data

126 The *in-situ* DO concentration data (3214 data points) used for modeling were
127 compiled from previous studies, including some from the National Earth System
128 Science Data Center (<http://www.geodata.cn/index.html>). Details about the data source
129 can be found in [Table S1](#). Data samples show a higher frequency in Jun and Aug
130 (usually when DO is low) as shown in [Fig. S1a](#). Spatially, the distribution of samples
131 adequately covers the entire study area ([Fig. 1b](#)) and both above and below the mixed
132 layer depth (MLD) ([Fig. S1b](#)). The climatological seawater temperature at the vertical
133 profile relating to the DO in Section 4 was obtained from the Regional Climatological
134 Dataset of East Asia (RCDEA; [7](https://www.ncei.noaa.gov/products/regional-ocean-</p></div><div data-bbox=)

135 climatologies).

136 **2.2.2 Satellite data**

137 The 8-day sea surface temperature (SST) satellite data at a resolution of 4 km
138 during 2011–2019 were extracted from level-3 product of MODIS Aqua Ocean Color
139 Data, produced by NASA's Earth Observing System Data and Information System
140 (EOSDIS; <https://oceancolor.gsfc.nasa.gov/>). The 8-day sea surface Chlorophyll-*a*
141 (Chl-*a*) concentration data at 4 km resolution were obtained from Ocean Color-CCI
142 (OC-CCI), which provides Chl-*a* multi-satellite fusion products from a blended
143 combination of OCI, OCI2, OC2 and OCx algorithms (<https://www.oceancolour.org/>).

144 To enhance the model's accuracy and validity, we conducted Data Interpolating
145 Empirical Orthogonal Functions (DINEOF) interpolation (Li et al., 2023a; Niu et al.,
146 2021) to fill in missing pixels (caused by weather or technology) and aligned the DO
147 samples with corresponding satellite observations from the same week (8 days). The
148 coefficient of variation (CV) of the 3*3 window was restricted to within 20% to filter
149 sensor and algorithm noises. Data that not meet this criterion were excluded (about
150 3.76%). Additional details about the data can be found in Table 1.

151 **Table 1** Data sources used in this study.

Variables	Data Product	Temporal Resolution	Spatial Resolution
<i>In-situ</i> DO	(Present study)	/	/
Modeling DO	GOBH	Daily	1/4°
SST	MODIS-Aqua	8-day	4km

Seawater temperature	RCDEA	Monthly climatological	1/10°
	GLORYS	Monthly	1/12°
Chl- <i>a</i>	OC-CCI	8-day	4 km
MLD	GLBu0.08	8-day	1/12°

152 **2.2.3 Simulated data acquisition and processing**

153 8-day composite MLD data at a resolution of 1/12° were derived from GLBu0.08
154 from the Hybrid Coordinate Ocean Model (HYCOM; <https://www.hycom.org/>). The
155 simulated DO concentration data were obtained from the Global Ocean
156 Biogeochemistry Hindcast (GOBH) model, produced by the Copernicus Marine
157 Environment Monitoring Service (CMEMS; <http://marine.copernicus.eu/>) and were
158 used for comparison with our model. Monthly three-dimensional seawater temperature
159 and salinity data were obtained from the Global Ocean Physics Reanalysis (GLORYS)
160 product from the CMEMS and were used to define the SYSCWM region. This region
161 was defined as areas within the 10 °C isotherm at the bottom layer for a given month.

162 **2.3 Models**

163 The model selection and development procedures are illustrated in Fig. 2a, with
164 more detailed steps are shown in Fig. 2b. The model selection included 15*4*50
165 independent simulations. The number 15 represents the number of input variable sets.
166 Each set consists of SST, water depth, MLD, Chl-*a*, and time (coded by an integer). In
167 every set, only the Chl-*a* varies while the other variables stay unchanged. We focused
168 on the cumulative effect of surface Chl-*a* prior to *in-situ* DO. Subsequently, Chl-*a* from

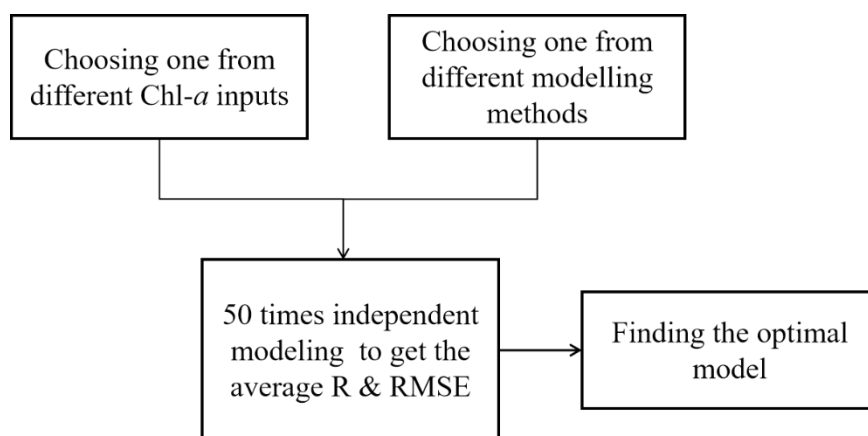
169 0 to 14 weeks before *in-situ* DO was extracted. Then, 15 means of Chl-*a* averaged over
170 specific weeks were imputed for selection, as Chl-*a* has a delayed effect on DO with
171 different time lags (Zheng and DiGiacomo, 2020). The number 4 represents four
172 modeling methods, i.e., RF, SVR, Generalized Regression Neural Network (GRNN)
173 and Stepwise Linear Regression (SR). These models have been demonstrated to be
174 effective in predicting DO in multiple previous studies (Guo et al., 2021; Heddam, 2014;
175 Ji et al., 2017; Li et al., 2023b; Valera et al., 2020). Detailed information about the
176 structure and principles of the four models can be found in the [supplementary materials](#).

177 For each of the 15*4 choices, a model was developed and repeated 50 times. In
178 each iteration, input and output variables were randomly divided into a training set (70%
179 of data points) and a test set (30% of data points) using a hierarchical random sampling
180 method. The data in the training set were used for modeling, while the data in the test
181 set were employed for evaluation. The average root mean square error (RMSE, Eq. (S1))
182 and average correlation coefficient (R, Eq. (1)) of the test set from 50 simulations were
183 calculated and recorded.

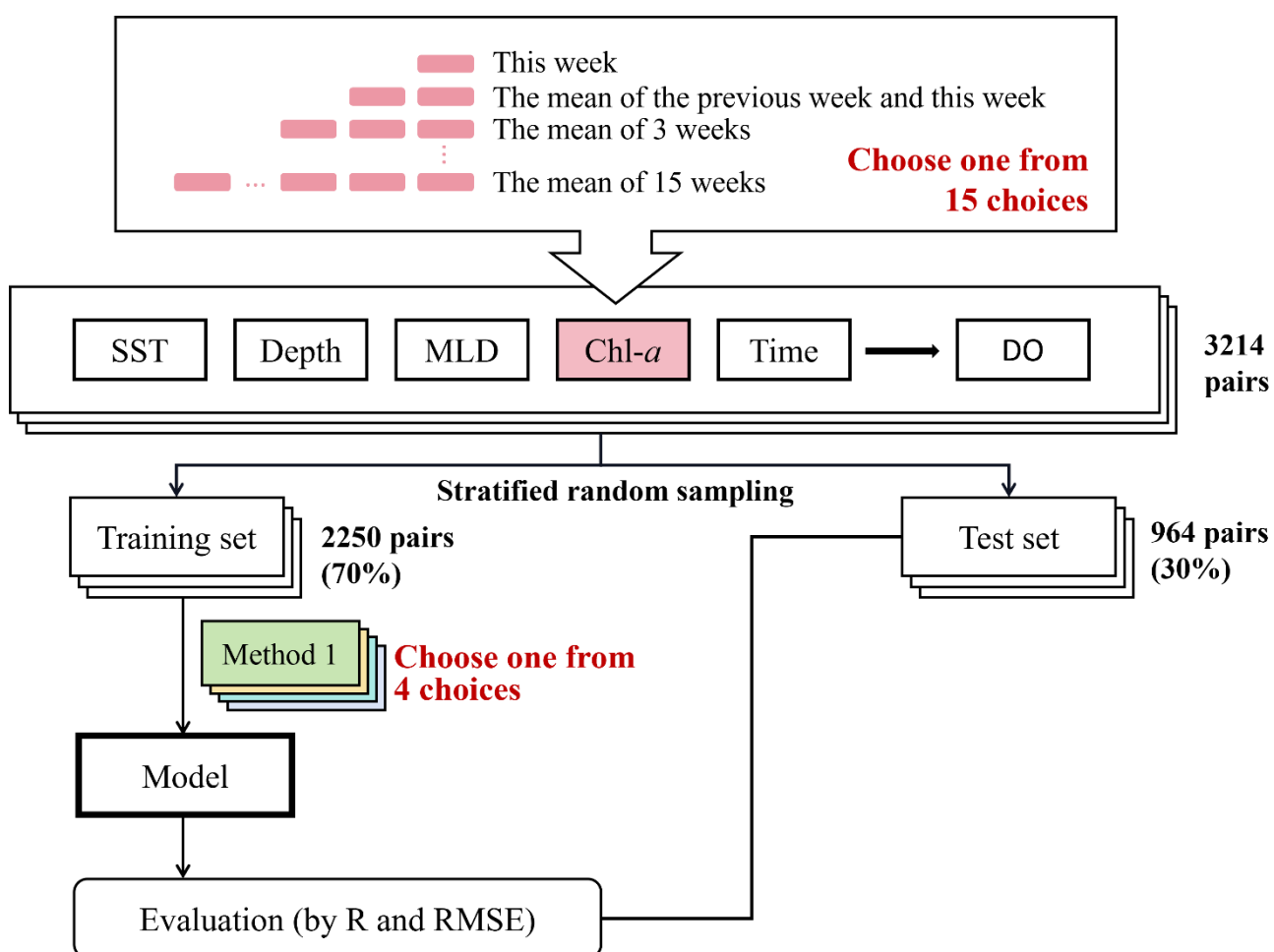
$$184 \quad R = \frac{\sum_{i=1}^N (y_i - \bar{y}_i) * (\hat{y}_i - \bar{\hat{y}}_i)}{\sqrt{\sum_{i=1}^N (y_i - \bar{y}_i)^2 * (\hat{y}_i - \bar{\hat{y}}_i)^2}} \quad (1)$$

185 Where N represents the number of the match-up pairs, y_i , \bar{y}_i , \hat{y}_i , and $\bar{\hat{y}}_i$ are the
186 *in-situ*, mean *in-situ*, simulated, and mean simulated DO concentrations, respectively.

(a)



(b)

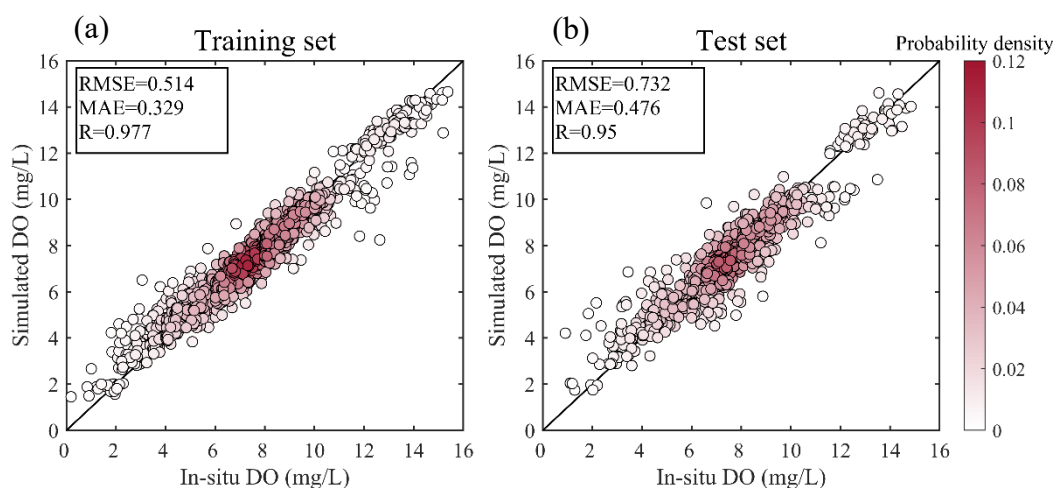


187 **Fig. 2** Sketch diagrams showing (a) overall procedures of model selection and (b) the detailed
 188 procedures of modeling.

189 3. Results

190 3.1 Performance of models

191 The RF model, which used the mean of 8 weeks' Chl-*a* as one of the input variables,
192 produces the lowest RMSE and highest R values. This indicates its superior
193 performance compared to the other models (Fig. S2). The results predicted by the RF
194 model closely match the *in-situ* DO observations (Fig.S2b), confirming its effectiveness.
195 Hence, we selected the RF model with the 8 weeks mean Chl-*a* concentration as one
196 input variable as our final model. However, when comparing its performance on the
197 test set with training dataset, we observe slight overfitting of the RF model to the
198 training data (Fig. 3). In Fig. 3b, some outliers appear in the low DO segment (< 4mg/L),
199 indicating that the model overestimates DO in this range. Despite of this defect, the
200 model demonstrates reliable performance in both the training and test sets, with RMSE
201 < 0. 8 mg/L, $R > 0. 95$, and mean absolute error (Eq. S2) < 0. 5 mg/L. We also compared
202 the model performances above and below the MLD. Both regions exhibit a similar
203 ability to the overall model (Fig. S3), indicating consistent performance across different



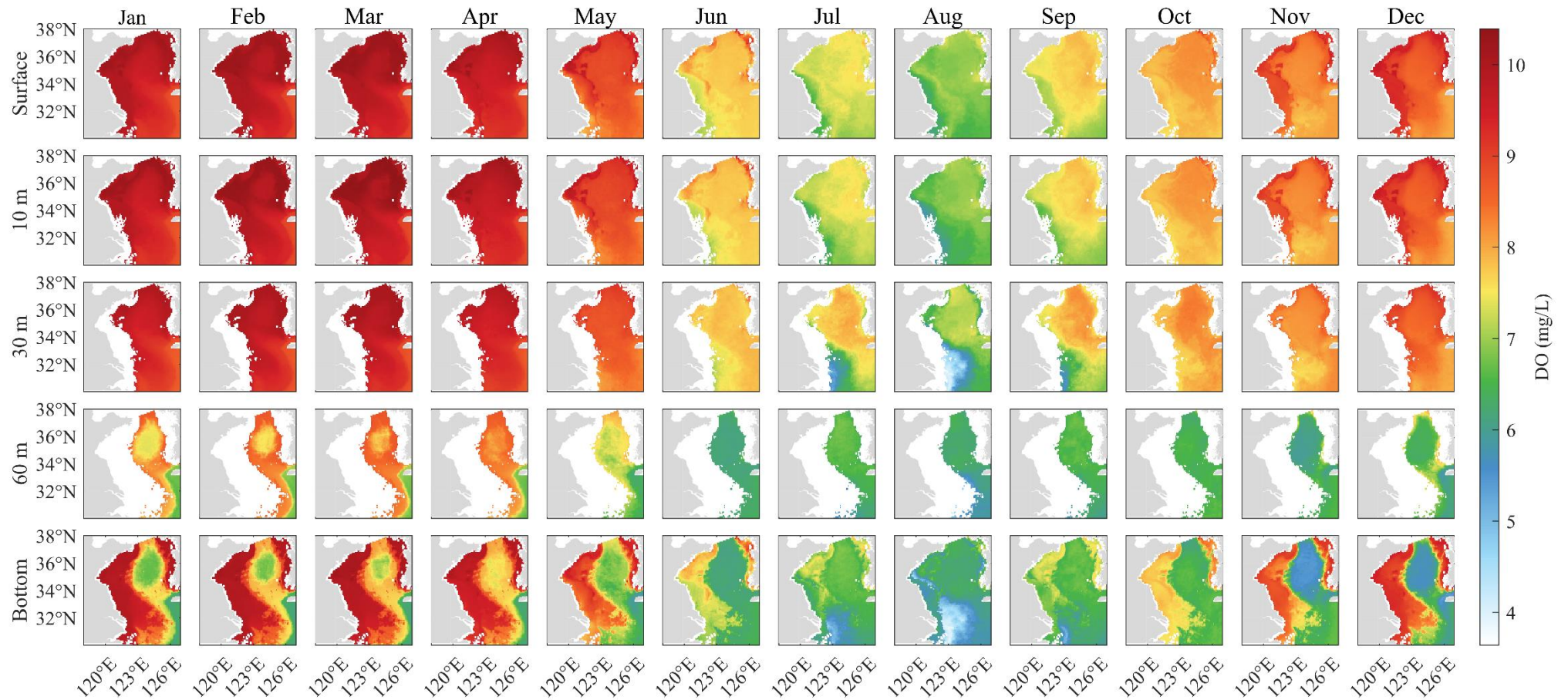
204 water depths.

205 **Fig. 3** Scatter plots of results of the RF model alongside the *in-situ* data. (a) Training set, (b) Test
206 set.

207 **3.2 Variation of Monthly DO during 2011–2019**

208 The monthly climatological DO derived from the model outputs during 2011–
209 2019 reveals that DO generally decreased from the inshore to the center of the SYS and
210 from the upper layer to the deeper layer in most months (Fig. 4). However, above the
211 10 m layer from Jun to Sep, DO increases from the inshore to the center of the SYS. In
212 addition, both the surface and the 10 m layers show prominent bands of high DO
213 inshore in most months, especially from Apr to Jun. Furthermore, from Jun to Aug, an
214 area with DO higher than both the upper and deeper layers is displayed at the 30 m
215 layer in the center of the SYS.

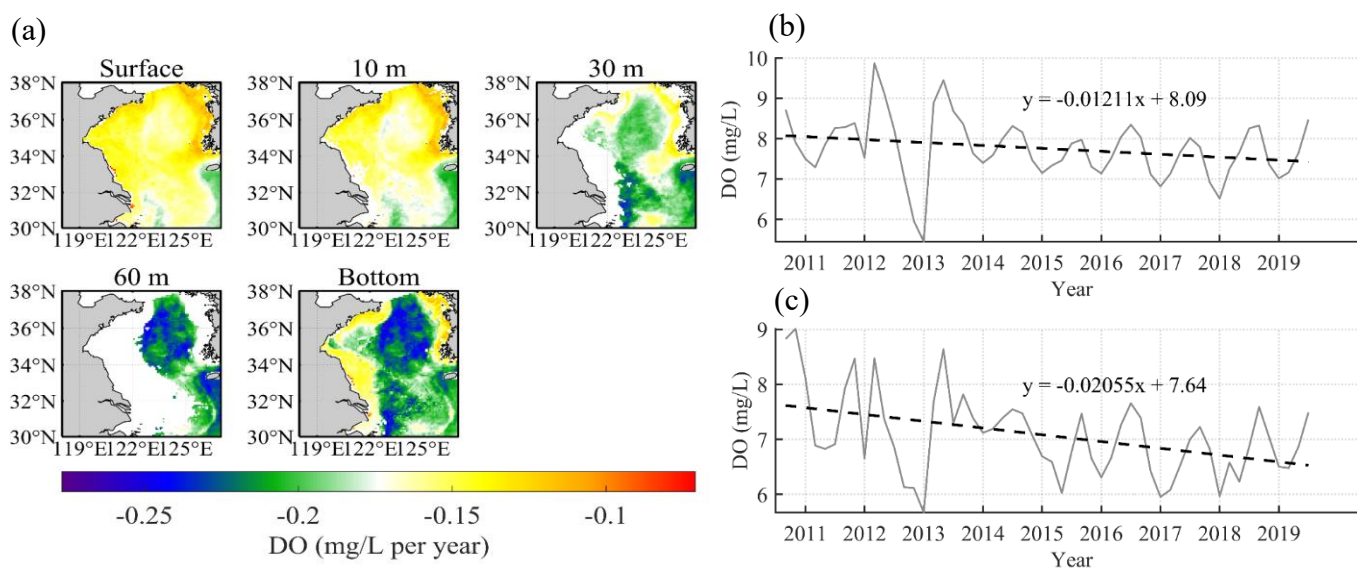
216 The DO in the SYS also exhibits significant seasonal variations. Generally, DO
217 exhibits a gradual decline from late winter (Jan, Feb) to midsummer (Aug), followed
218 by a subsequent increase after early autumn (Sep, Oct). The lowest DO levels are
219 observed in August, while the highest levels occur in Mar. However, the recovery of
220 DO does not take place during autumn (Sep, Oct, and Nov) in the center of the SYS
221 (SYSCWM region) at a depth of 60 m.



222 **Fig. 4** Simulated monthly climatological DO concentrations from Jan 2011–Dec 2019. The surface and bottom layers are defined as the layer with a depth of 1 m and
223 the depth of the seabed topography, respectively.

224 3.3 Long-term patterns of DO during 2011–2019

225 The long-term temporal variations of DO across different layers were analyzed
226 during 2011–2019. By using the slope coefficient of a linear regression model as an
227 indicator of the change extent, the spatial trends in DO were determined. During 2011–
228 2019, DO in the SYS show a slight downward trend, ranging from 0.0712 to 0.277
229 mg/L per year. A more profound decrease is observed at deeper layers compared to the
230 surface (Fig. 5a). Particularly noteworthy is the significant downward trend below the
231 60 m layer of the center of the SYS, where the seasonal SYSCWM typically forms. We
232 further divided the water column above and below the MLD in the SYSCWM region
233 using the simulated MLD data (the well-known SYSCWM is typically found below the
234 MLD in this division). Then, the DO trends above and below the MLD in the SYSCWM
235 region were analyzed, revealing a consistent decline in both regions, with a more
236 pronounced decline in DO below the MLD (Fig. 5b). This suggests the possibility of
237 different regulating mechanisms.



238 **Fig. 5** Long-term trend of DO. (a) Spatial patterns and (b, c) temporal evolutions of DO
239 concentrations in the SYSCWM region (b) Above the MLD, (c) Below the MLD (the DO
240 concentration from the model, with a vertical resolution of 1 m, was monthly averaged above and
241 below the MLD). The period shown in (b, c) only includes the DO of Jun to Nov each year. During
242 this period, a closure of the SYSCWM region can be identified in most of the years. In (b) and (c),
243 missing values were replaced by interpolated values.

244 **4. Discussion**

245 **4.1 Mechanisms underlying the spatiotemporal patterns of DO**

246 We observe close relationship between SST and DO, as evident from [Figs. 4 and](#)
247 [S4](#). DO shows a gradual decline from late winter (Jan, Feb) to midsummer, followed
248 by an increase after early autumn (Sep, Oct). Meanwhile, SST increases from early
249 spring (Mar) and declines after autumn (Sep). This negative correlation is mainly
250 explained by the lower DO solubility in warmer seawater. In addition, enhanced
251 stratification caused by high seawater temperature at the surface can also limit
252 ventilation into deep water, leading to depletion of DO at the bottom. In Jan and Feb,
253 the dominant current patterns of YSCC and YSWC result in high seawater temperature
254 in the central region and low temperatures inshore. The spatial patterns of DO at surface,
255 10 m and 30 m layers in these months closely match the patterns of SST ([Fig. S4](#)), with
256 high DO inshore and low DO in the center of the SYS. This indicates that DO in these
257 months is primarily driven by seawater temperature. This is likely due to the well-mixed
258 water column in winter ([Fig. S5](#)), leading to sufficient replenishment of DO, thus

259 following the oxygen solubility (Wei et al., 2021). From Jun to Sep, there is a reversal
260 in the trend in the upper layers, with lower DO inshore. This can be attributed to the
261 shallow water depth in the Subei Shoal, resulting in elevated seawater temperatures.
262 The strong correlation of SST and DO is further supported by the fact that SST emerges
263 as a significant factor among all variables considered in our model (Fig. S6).

264 DO concentrations are significantly influenced by biological activities such as
265 photosynthesis and respiration. The presence of a high DO band inshore at the surface
266 and 10 m layers is evident from Apr to Jun. This can be attributed to phytoplankton
267 blooms, which is supported by the frequent observation of green tides in Subei Shoal
268 from May to Jul (Hu et al., 2010; Wei et al., 2018). In early spring, phytoplankton
269 growth is stimulated by favorable seawater temperature and light. At the same time,
270 nutrient supplies are enhanced by stronger vertical mixing. This leads to subsequent
271 DO production through photosynthesis (Niu et al., 2021). During summer, the
272 maximum DO is found in the middle depth of the SYS, particularly from Jul to Sep as
273 shown in Fig. 4. The maximum DO concentration is typically found at around 30 m in
274 the center of the SYS during summer (Wei et al., 2010). At this depth, cold water exists
275 due to strong stratification (Fig. S7), which prevents mixing with the adjacent upper
276 and lower layers. In addition, higher Chl-*a* levels are also found at this depth, suggesting
277 that phytoplankton may contribute to the high DO levels in this region (Zheng et al.,
278 2006). However, this phenomenon gradually diminished as the MLD became deeper
279 (Fig. S5). A sustained low DO concentration in the deep central SYS is observed under

280 the 60 m layer in the following months. This can be attributed to the persistence of the
281 stratification until Nov (Fig. S7). In the absence of sufficient ventilation, DO in deep
282 water is depleted by the remineralization of sinking organic particles.

283 **4.2 Mechanisms of DO decline in the SYSCWM**

284 Even though our results suggest that seasonal DO variations are highly correlated
285 with seasonal seawater temperature changes in Section 4.1, the long-term downward
286 trend in the center of the SYS is not solely caused by rising temperature. Thus, we used
287 a computational method to separate the effect of biological activities and warming on
288 deoxygenation. The details of this method are described in the [supplementary materials](#).

289 In the SYSCWM region, either above or below the MLD, the primary factor
290 accounting for decline in DO is the biological activity (Table S2). These could be
291 interpreted as either a decline in oxygen production or an increase in oxygen
292 consumption. Above the MLD in the water column, DO often appears oversaturated as
293 phytoplankton thrives under optimal temperature and light conditions (Wei et al., 2021).
294 This suggests that the decline in DO above the MLD can be attributed to a decrease in
295 oxygen production during photosynthesis. Below the MLD in the water column, the
296 decline in DO may be attributed to the more intense oxygen consumption of sinking
297 organic particles through respiration. To see if phytoplankton biomass is changing as
298 expected, the trends of sea surface Chl-*a* at the annual and monthly scales were plotted
299 respectively (Fig. 6, S8).

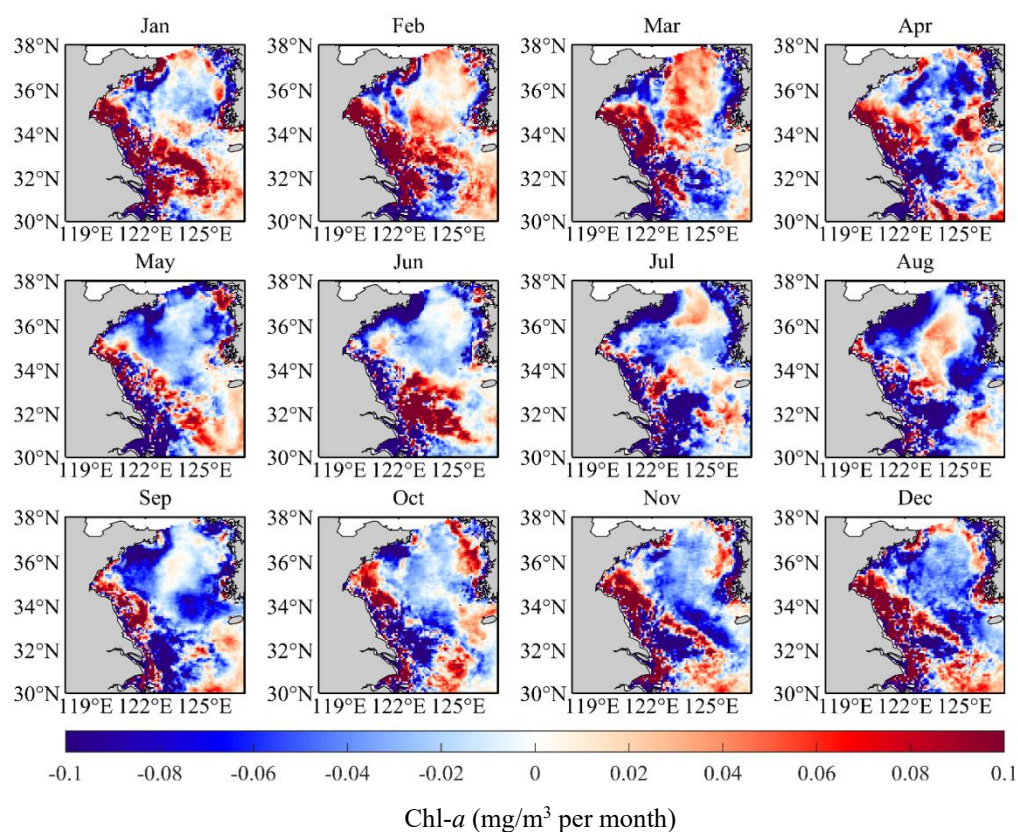
300 After examining the trend of the sea surface Chl-*a*, a declining trend was found

301 during 2011–2019 in the central SYS both in annual scale and most month (Fig. S8),
302 which aligns with our inferences above the MLD. But this trend seems to contradict the
303 decline in DO below the MLD. However, from the monthly Chl-*a* trend, we find that it
304 has an upward trend in spring. Fig. 6 shows a clear increasing trend in Chl-*a* in the
305 center of the SYS (the SYSCWM region) in Feb and Mar, especially in Mar. This
306 indicates that particles formed from a luxuriant phytoplankton bloom in early spring
307 may sink and consume DO below the MLD, thereby reducing DO in the deep water in
308 next several months. Some studies show a similar phenomenon that organic particles
309 produced from the phytoplankton blooms in spring may exacerbate depletion of DO at
310 the bottom, and the oxygen consuming effect may continue from spring into summer
311 (Zheng and DiGiacomo, 2020). Notably, there is also a significant increase in Chl-*a* in
312 some months near the Subei shoal (e.g. June). Although the increase in Chl-*a* in this
313 part may have an effect on the DO below it and, therefore, on the DO in SYSCWM,
314 this effect may be limited by the presence of a seawater temperature front that surrounds

315 the SYSCWM region (Chen, 2009; Hickox et al., 2000).

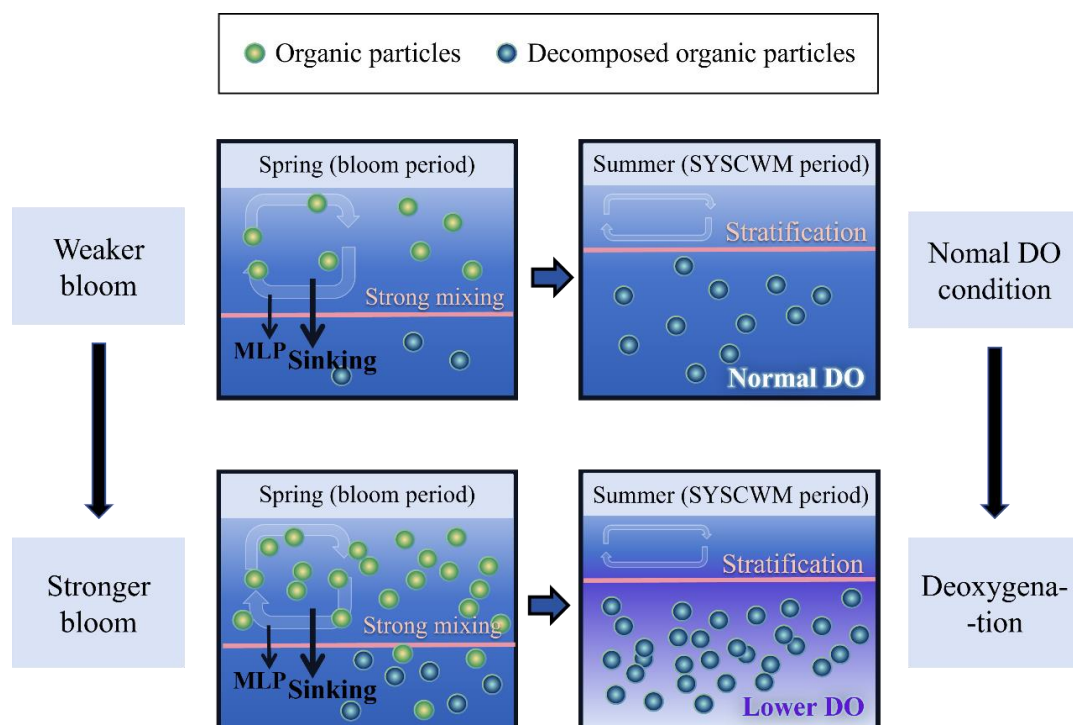
316 **Fig. 6** Long-term monthly variations of Chl-*a* concentration during 2011–2019.

317 To investigate whether the biological contribution to the decline in DO in the
318 SYSCWM (i.e., below the MLD in the SYSCWM region) can be attributed to the
319 stronger early spring bloom, we plotted the Chl-*a* in Mar and the DO in Aug in the
320 SYSCWM region (below the MLD). These months were chosen because they denoted
321 the increase in phytoplankton bloom in early spring and the peak phase of the
322 SYSCWM, respectively. The correlation between them is -0.772 ($P < 0.05$), indicating
323 a plausible link between stronger early spring phytoplankton blooms and reduced DO
324 in the SYSCWM (Fig. S9). We also calculated the correlation between DO in Jun in the
325 SYSCWM and that in Aug ($P < 0.05$). The strong correlation suggests that the low DO



326 in Aug may be due to that in Jun. Similarly, we investigated whether the Chl-*a* in the
327 southern part of the study area in Jun was related to the DO of the SYSCWM. The
328 results showed that there was no significant correlation between them ($p > 0.05$),
329 suggesting the effect of the seawater temperature fronts as a barriers. Collectively, our
330 hypothesis that the organic particles formed in spring may account for the decline in
331 DO in the SYSCWM remains valid. The proposed mechanism is visually represented
332 in [Fig. 7](#).

333 In addition, physical processes also play a role in this process. For instance, the
334 process illustrated in [Fig. 7](#) may be augmented by the mixed layer pump (MLP)
335 observed in many studies. That is, in regions with significant seasonal variation in the
336 MLD, surface organic particles are pushed into deeper waters by intensified wind
337 mixing during the colder seasons ([Lacour et al., 2019](#); [Xing et al., 2020](#)). Stronger
338 stratification due to warming can reduce nutrient supply to the surface, subsequently
339 decreasing primary production ([Dave and Lozier, 2013](#)). It can also limit the extent of
340 ventilation from the surface to the deep water. In essence, warming and biological
341 activities can sometimes interact. Aside from directly affecting the DO content entering
342 the ocean (through air-sea exchange), warming can also change DO content by
343 influencing ocean hydrology which then regulates the biological activities.



344 **Fig.7** Schematic of the mechanisms of deoxygenation explained by biological activities in the
 345 SYSCWM. In spring, stronger phytoplankton blooms formed more sinking particles, while the
 346 mixing gradually weakened and the water column gradually stratified in the following months. Then
 347 these particles continued to sink and gradually consumed more DO below the MLD through
 348 respiration. In summer, or the SYSCWM period, lower DO is present and it is difficult to be
 349 replenished from the upper layers of the stratified water column.

350 In addition to the overall downward trend, there are some details worth our
 351 attention in Fig.5, such as DO in the year 2013 is considerably lower than any other
 352 years. After plotting the interannual variation of SST in Aug and Chl-*a* in Feb and Mar,
 353 we found that 2013 had a higher SST than most years and the highest spring Chl-*a* (Fig.
 354 S10) High seawater temperatures and strong stratification, combined with enough
 355 sinking particles produced by phytoplankton, contributed to the low DO in summer

356 2013.

357 **4.3 Advantages and drawbacks of the study**

358 Based on extensive previous studies, our modelling study provides distribution
359 and trends of DO with higher spatial-temporal resolution, thus highlighting locations
360 of more intense deoxygenation. The spatial patterns of DO exhibited by our model
361 aligns closely with a substantial body of empirical studies (Guo et al., 2020b; Guo et
362 al., 2020c; Li et al., 2015; Lu et al., 2017; Luo et al., 2018; Qu et al., 2015; Xiong et
363 al., 2020; Zhu et al., 2017). Besides, the declining DO observed by our model in the
364 SYS is consistent with previous findings (Wei et al., 2021), which also show a more
365 significant decline in DO in the deeper layers of the SYS. This further substantiates
366 the reliability of modeling as an approach to comprehensively capture spatiotemporal
367 patterns despite limited *in-situ* data availability. Although global model data with
368 similar functions are available, our model exhibited higher accuracy and higher
369 resolution in comparison to the global model in our study area (Fig. S11). Moreover,
370 the error observed in our study is considered acceptable compared to similar studies,
371 which achieved an RMSE of approximately 1 mg/L (Guo et al., 2021; Kim et al., 2020;
372 Li et al., 2023b; Ross and Stock, 2019).

373 This study not only captures the spatial and temporal characteristics of DO, but
374 also demonstrates the feasibility of machine learning combined with satellite data to
375 predict the DO concentration in the SYS. This approach would reduce the cost of
376 instruments and labor. However, narrowing down the error caused by potential

377 inaccuracies in satellite data as well as the inherent limitations of the model itself
378 remains a challenging task. Our model may not fully capture the actual variations in
379 extremely low DO conditions (e.g., < 4 mg/L). This can be attributed to the involvement
380 of more intricate processes in DO dynamics in these cases, such as sediment
381 consumption and advection transport. There are also other challenges that remain to be
382 further addressed. For example, the precision of input data and the comprehensiveness
383 of *in-situ* data need improvement. In the future, we anticipate that more *in-situ* data will
384 be collected, satellite data will be accurately calibrated, and additional variables will be
385 incorporated to enhance the precision of the model. Moreover, due to the introduction
386 of the time variable into the model and its significant contribution to the result (Fig. S6),
387 the accuracy of dissolved oxygen simulation beyond the time range cannot be
388 guaranteed. It is expected that the model's time variable will be adjusted or deleted in
389 the future to enhance the usability of the model in forecasting. More discussions on the
390 time variable and the partial correlation curves (Fig. S12) can be found in the
391 supplementary material. Additionally, the potential of the proposed approach for
392 achieving similar performance in other waters remains to be tested.

393 **5. Conclusion**

394 In this study, an RF model was presented to efficiently and accurately simulate the
395 DO conditions of multiple layers in the SYS during 2011–2019. This provided a
396 comprehensive spatiotemporal distribution of DO within the SYS. The model has
397 achieved good accuracy and effectively captures a declining trend of DO in the SYS,

398 with a more significant decline at the deep layers compared to the shallower ones. The
399 decline in DO in the SYSCWM region above and below the MLD was regulated by
400 decreased surface Chl-*a* and stronger early spring phytoplankton blooms, respectively.
401 Although this study focuses on DO in the SYS, the proposed framework and
402 methodology can be readily applied to other coastal systems lacking sufficient *in-situ*
403 data. The findings of the model may serve as a valuable reference for high-resolution
404 modeling studies and other related investigations. Additionally, it can aid in the
405 development of appropriate frameworks for deep-sea aquaculture activities and cruise
406 survey planning.

407

408 **Data availability statement**

409 Data will be made available on request.

410

411 **Author Contribution Statement:**

412 Q.L. handled data curation, conceptualization, and wrote the original draft. C.L.
413 supervised, acquired funding, and reviewed/edited. Q.M. acquired funding and
414 reviewed/edited. B.S., H.Y., B.C., and W.L. reviewed/edited. X.C. curated data. W.N.
415 performed formal analysis. N.M. visualized the data.

416

417 **Declaration of competing interest**

418 The authors declare that they have no competing financial interests or personal

419 relationships that could have appeared to influence the work reported in this paper. All
420 authors have read and agree to the published version of the manuscript.

421

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427

428 **Reference**

- 429 Banks, J., Ross, D.J., Keough, M.J., 2012. Short-term (24 h) effects of mild and severe hypoxia (20% and
430 5% dissolved oxygen) on metal partitioning in highly contaminated estuarine sediments.
431 *Estuarine Coastal and Shelf Science*, 99: 121-131.
432 <http://dx.doi.org/10.1016/j.ecss.2011.12.025>
- 433 Barbier, E.B., Hacker, S.D., Kennedy, C., Koch, E.W., Stier, A.C., Silliman, B.R., 2011. The value of estuarine
434 and coastal ecosystem services. *Ecological Monographs*, 81(2): 169-193.
435 <http://dx.doi.org/10.1890/10-1510.1>
- 436 Batziakas, S., Frangoulis, C., Tsiola, A., Nikolioudakis, N., Tsagaraki, T.M., Somarakis, S., 2020. Hypoxia
437 changes the shape of the biomass size spectrum of planktonic communities: a case study in
438 the eastern Mediterranean (Elefsina Bay). *Journal of Plankton Research*, 42(6): 752-766.
439 <http://dx.doi.org/10.1093/plankt/fbaa055>
- 440 Beyan, C., Browman, H.I., 2020. Setting the stage for the machine intelligence era in marine science.
441 *Ices Journal of Marine Science*, 77(4): 1267-1273. <http://dx.doi.org/10.1093/icesjms/fsaa084>
- 442 Breitbart, D., Levin, L.A., Oschlies, A., Gregoire, M., Chavez, F.P., Conley, D.J., Garcon, V., Gilbert, D.,
443 Gutierrez, D., Isensee, K., Jacinto, G.S., Limburg, K.E., Montes, I., Naqvi, S.W.A., Pitcher, G.C.,
444 Rabalais, N.N., Roman, M.R., Rose, K.A., Seibel, B.A., Telszewski, M., Yasuhara, M., Zhang, J.,
445 2018. Declining oxygen in the global ocean and coastal waters. *Science*, 359(6371): 46-+.
446 <http://dx.doi.org/10.1126/science.aam7240>
- 447 Bushinsky, S.M., Emerson, S.R., Riser, S.C., Swift, D.D., 2016. Accurate oxygen measurements on
448 modified Argo floats using in situ air calibrations. *Limnology and Oceanography-Methods*,
449 14(8): 491-505. <http://dx.doi.org/10.1002/lom3.10107>
- 450 Chen, B., Liu, H., Xiao, W., Wang, L., Huang, B., 2020. A machine-learning approach to modeling
451 picophytoplankton abundances in the South China Sea. *Progress in Oceanography*, 189.

- 452 <http://dx.doi.org/10.1016/j.pocean.2020.102456>
- 453 Chen, C.-T.A., 2009. Chemical and physical fronts in the Bohai, Yellow and East China seas. *Journal of*
454 *Marine Systems*, 78(3): 394-410. <http://dx.doi.org/10.1016/j.jmarsys.2008.11.016>
- 455 Codispoti, L.A., Brandes, J.A., Christensen, J.P., Devol, A.H., Naqvi, S.W.A., Paerl, H.W., Yoshinari, T., 2001.
456 The oceanic fixed nitrogen and nitrous oxide budgets: Moving targets as we enter the
457 anthropocene? *Scientia Marina*, 65: 85-105. <http://dx.doi.org/10.3989/scimar.2001.65s285>
- 458 Dave, A.C., Lozier, M.S., 2013. Examining the global record of interannual variability in stratification and
459 marine productivity in the low-latitude and mid-latitude ocean. *Journal of Geophysical*
460 *Research: Oceans*, 118(6): 3114-3127. <http://dx.doi.org/https://doi.org/10.1002/jgrc.20224>
- 461 Dong, S., 2019. Researching Progresses and Prospects in Large Salmonidae Farming in Cold Water Mass
462 of Yellow Sea. *Periodical of Ocean University of China*, 49(3): 1-6.
- 463 Edwards, B., Murphy, D., Janzen, C., Larson, N., 2010. Calibration, Response, and Hysteresis in Deep-Sea
464 Dissolved Oxygen Measurements. *Journal of Atmospheric and Oceanic Technology*, 27(5): 920-
465 931. <http://dx.doi.org/10.1175/2009jtecho693.1>
- 466 Fennel, K., Testa, J.M., 2019. Biogeochemical Controls on Coastal Hypoxia. In: Carlson, C.A., Giovannoni,
467 S.J. (Eds.), *Annual Review of Marine Science*, Vol 11. *Annual Review of Marine Science*, pp. 105-
468 130. DOI:10.1146/annurev-marine-010318-095138
- 469 Flint, N., Crossland, M.R., Pearson, R.G., 2015. Sublethal effects of fluctuating hypoxia on juvenile
470 tropical Australian freshwater fish. *Marine and Freshwater Research*, 66(4): 293-304.
471 <http://dx.doi.org/10.1071/mf14120>
- 472 Fock, H.O., Czudaj, S., 2019. Size structure changes of mesopelagic fishes and community biomass size
473 spectra along a transect from the equator to the Bay of Biscay collected in 1966-1979 and 2014-
474 2015. *Ices Journal of Marine Science*, 76(3): 755-770.
475 <http://dx.doi.org/10.1093/icesjms/fsy068>
- 476 Gilbert, D., Rabalais, N.N., Diaz, R.J., Zhang, J., 2010. Evidence for greater oxygen decline rates in the
477 coastal ocean than in the open ocean. *Biogeosciences*, 7(7): 2283-2296.
478 <http://dx.doi.org/10.5194/bg-7-2283-2010>
- 479 Goldstein, E.B., Coco, G., Plant, N.G., 2019. A review of machine learning applications to coastal
480 sediment transport and morphodynamics. *Earth-Science Reviews*, 194: 97-108.
481 <http://dx.doi.org/10.1016/j.earscirev.2019.04.022>
- 482 Guo, C., Zhang, G., Sun, J., Leng, X., Xu, W., Wu, C., Li, X., Pujari, L., 2020a. Seasonal responses of nutrient
483 to hydrology and biology in the southern Yellow Sea. *Continental Shelf Research*, 206.
484 <http://dx.doi.org/10.1016/j.csr.2020.104207>
- 485 Guo, H., Huang, J.J., Zhu, X., Wang, B., Tian, S., Xu, W., Mai, Y., 2021. A generalized machine learning
486 approach for dissolved oxygen estimation at multiple spatiotemporal scales using remote
487 sensing. *Environ Pollut*, 288: 117734. <http://dx.doi.org/10.1016/j.envpol.2021.117734>
- 488 Guo, J., Yuan, H., Song, J., Li, X., Duan, L., 2020b. Hypoxia, acidification and nutrient accumulation in the
489 Yellow Sea Cold Water of the South Yellow Sea. *Sci Total Environ*, 745: 141050.
490 <http://dx.doi.org/10.1016/j.scitotenv.2020.141050>
- 491 Guo, X., Xu, B., Burnett, W.C., Wei, Q., Nan, H., Zhao, S., Charette, M.A., Lian, E., Chen, G., Yu, Z., 2020c.
492 Does submarine groundwater discharge contribute to summer hypoxia in the Changjiang
493 (Yangtze) River Estuary? *Sci Total Environ*, 719: 137450.
494 <http://dx.doi.org/10.1016/j.scitotenv.2020.137450>

- 495 Heddam, S., 2014. Generalized regression neural network-based approach for modelling hourly
496 dissolved oxygen concentration in the Upper Klamath River, Oregon, USA. *Environ Technol*,
497 35(13-16): 1650-7. <http://dx.doi.org/10.1080/09593330.2013.878396>
- 498 Hickox, R., Belkin, I., Cornillon, P., Shan, Z., 2000. Climatology and seasonal variability of ocean fronts in
499 the East China, Yellow and Bohai seas from satellite SST data. *Geophysical Research Letters*,
500 27(18): 2945-2948. <http://dx.doi.org/10.1029/1999gl011223>
- 501 Hou, J., Zhou, W., Wang, L., Fan, W., Yuan, Z., 2020. Spatial analysis of the potential of deep-sea
502 aquaculture in China. *Resources Science*, 42(7): 1325-1337.
- 503 Hu, C., Li, D., Chen, C., Ge, J., Muller-Karger, F.E., Liu, J., Yu, F., He, M.-X., 2010. On the recurrent *Ulva*
504 *prolifera* blooms in the Yellow Sea and East China Sea. *Journal of Geophysical Research-Oceans*,
505 115. <http://dx.doi.org/10.1029/2009jc005561>
- 506 Ji, X., Shang, X., Dahlgren, R.A., Zhang, M., 2017. Prediction of dissolved oxygen concentration in hypoxic
507 river systems using support vector machine: a case study of Wen-Rui Tang River, China. *Environ*
508 *Sci Pollut Res Int*, 24(19): 16062-16076. <http://dx.doi.org/10.1007/s11356-017-9243-7>
- 509 Keeling, R.F., Garcia, H.E., 2002. The change in oceanic O-2 inventory associated with recent global
510 warming. *Proceedings of the National Academy of Sciences of the United States of America*,
511 99(12): 7848-7853. <http://dx.doi.org/10.1073/pnas.122154899>
- 512 Kim, Y.H., Son, S., Kim, H.C., Kim, B., Park, Y.G., Nam, J., Ryu, J., 2020. Application of satellite remote
513 sensing in monitoring dissolved oxygen variabilities: A case study for coastal waters in Korea.
514 *Environ Int*, 134: 105301. <http://dx.doi.org/10.1016/j.envint.2019.105301>
- 515 Lacour, L., Briggs, N., Claustre, H., Ardyna, M., Dall'Olmo, G., 2019. The Intraseasonal Dynamics of the
516 Mixed Layer Pump in the Subpolar North Atlantic Ocean: A Biogeochemical-Argo Float
517 Approach. *Global Biogeochemical Cycles*, 33(3): 266-281.
518 <http://dx.doi.org/10.1029/2018gb005997>
- 519 Levin, L.A., 2018. Manifestation, Drivers, and Emergence of Open Ocean Deoxygenation. In: Carlson,
520 C.A., Giovannoni, S.J. (Eds.), *Annual Review of Marine Science*, Vol 10. *Annual Review of Marine*
521 *Science*, pp. 229-260. DOI:10.1146/annurev-marine-121916-063359
- 522 Li, H.-M., Zhang, C.-S., Han, X.-R., Shi, X.-Y., 2015. Changes in concentrations of oxygen, dissolved
523 nitrogen, phosphate, and silicate in the southern Yellow Sea, 1980–2012: Sources and seaward
524 gradients. *Estuarine, Coastal and Shelf Science*, 163: 44-55.
525 <http://dx.doi.org/10.1016/j.ecss.2014.12.013>
- 526 Li, W., Liu, C., Zhai, W., Liu, H., Ma, W., 2023a. Remote sensing and machine learning method to support
527 sea surface pCO₂ estimation in the Yellow Sea. *Frontiers in Marine Science*, 10.
528 <http://dx.doi.org/10.3389/fmars.2023.1181095>
- 529 Li, W., Wang, Z., Huang, H., 2019. Relationship between the southern Yellow Sea Cold Water Mass and
530 the distribution and composition of suspended particulate matter in summer and autumn
531 seasons. *Journal of Sea Research*, 154. <http://dx.doi.org/10.1016/j.seares.2019.101812>
- 532 Li, Y., Robinson, S.V.J., Nguyen, L.H., Liu, J., 2023b. Satellite prediction of coastal hypoxia in the northern
533 Gulf of Mexico. *Remote Sensing of Environment*, 284.
534 <http://dx.doi.org/10.1016/j.rse.2022.113346>
- 535 Lin, C., Ning, X., Su, J., Lin, Y., Xu, B., 2005. Environmental changes and the responses of the ecosystems
536 of the Yellow Sea during 1976-2000. *Journal of Marine Systems*, 55(3-4): 223-234.
537 <http://dx.doi.org/10.1016/j.jmarsys.2004.08.001>

- 538 Liu, Z., Wei, H., Lozovatsky, I.D., Fernando, H.J.S., 2009. Late summer stratification, internal waves, and
539 turbulence in the Yellow Sea. *Journal of Marine Systems*, 77(4): 459-472.
540 <http://dx.doi.org/10.1016/j.jmarsys.2008.11.001>
- 541 Lonborg, C., Carreira, C., Jickells, T., Anton Alvarez-Salgado, X., 2020. Impacts of Global Change on Ocean
542 Dissolved Organic Carbon (DOC) Cycling. *Frontiers in Marine Science*, 7.
543 <http://dx.doi.org/10.3389/fmars.2020.00466>
- 544 Long, L., Liu, H., Cui, M., Zhang, C., Liu, C., 2023. Offshore aquaculture in China. *Reviews in Aquaculture*.
545 <http://dx.doi.org/10.1111/raq.12837>
- 546 Lu, W., Xiang, X., Yang, L., Xu, Y., Li, X., Liu, S., 2017. The temporal-spatial distribution and changes of
547 dissolved oxygen in the Changjiang Estuary and its adjacent waters for the last 50 a. *Acta*
548 *Oceanologica Sinica*, 36(5): 90-98. <http://dx.doi.org/10.1007/s13131-017-1063-6>
- 549 Luo, X., Wei, H., Fan, R., Liu, Z., Zhao, L., Lu, Y., 2018. On influencing factors of hypoxia in waters adjacent
550 to the Changjiang estuary. *Continental Shelf Research*, 152: 1-13.
551 <http://dx.doi.org/10.1016/j.csr.2017.10.004>
- 552 Malde, K., Handegard, N.O., Eikvil, L., Salberg, A.-B., 2020. Machine intelligence and the data-driven
553 future of marine science. *Ices Journal of Marine Science*, 77(4): 1274-1285.
554 <http://dx.doi.org/10.1093/icesjms/fsz057>
- 555 Mathew, D., Gireeshkumar, T.R., Shameem, K., Furtado, C.M., Arya, K.S., Udayakrishnan, P.B.,
556 Balachandran, K.K., 2022. Dynamics of trace metals in sediments of a seasonally hypoxic
557 coastal zone in the southeastern Arabian Sea. *Oceanologia*, 64(4): 735-748.
558 <http://dx.doi.org/10.1016/j.oceano.2022.06.007>
- 559 Naimie, C.E., Blain, C.A., Lynch, D.R., 2001. Seasonal mean circulation in the Yellow Sea - a model-
560 generated climatology. *Continental Shelf Research*, 21(6-7): 667-695.
- 561 Niu, Y., Liu, C., Lu, X., Zhu, L., Sun, Q., Wang, S., 2021. Phytoplankton blooms and its influencing
562 environmental factors in the southern Yellow Sea. *Regional Studies in Marine Science*, 47.
563 <http://dx.doi.org/10.1016/j.rsma.2021.101916>
- 564 Oschlies, A., Brandt, P., Stramma, L., Schmidtke, S., 2018. Drivers and mechanisms of ocean
565 deoxygenation. *Nature Geoscience*, 11(7): 467-473. [http://dx.doi.org/10.1038/s41561-018-](http://dx.doi.org/10.1038/s41561-018-0152-2)
566 [0152-2](http://dx.doi.org/10.1038/s41561-018-0152-2)
- 567 Pahlevan, N., Smith, B., Alikas, K., Anstee, J., Barbosa, C., Binding, C., Bresciani, M., Cremella, B., Giardino,
568 C., Gurlin, D., Fernandez, V., Jamet, C., Kangro, K., Lehmann, M.K., Loisel, H., Matsushita, B., Ha,
569 N., Olmanson, L., Potvin, G., Simis, S.G.H., VanderWoude, A., Vantrepotte, V., Ruiz-Verdu, A.,
570 2022. Simultaneous retrieval of selected optical water quality indicators from Landsat-8,
571 Sentinel-2, and Sentinel-3. *Remote Sensing of Environment*, 270.
572 <http://dx.doi.org/10.1016/j.rse.2021.112860>
- 573 Qu, B., Song, J., Yuan, H., Li, X., Li, N., Duan, L., Chen, X., Lu, X., 2015. Summer carbonate chemistry
574 dynamics in the Southern Yellow Sea and the East China Sea: Regional variations and controls.
575 *Continental Shelf Research*, 111: 250-261. <http://dx.doi.org/10.1016/j.csr.2015.08.017>
- 576 Ross, A.C., Stock, C.A., 2019. An assessment of the predictability of column minimum dissolved oxygen
577 concentrations in Chesapeake Bay using a machine learning model. *Estuarine, Coastal and*
578 *Shelf Science*, 221: 53-65. <http://dx.doi.org/10.1016/j.ecss.2019.03.007>
- 579 Rubbens, P., Brodie, S., Cordier, T., Destro Barcellos, D., Devos, P., Fernandes-Salvador, J.A., Fincham, J.I.,
580 Gomes, A., Handegard, N.O., Howell, K., Jamet, C., Kartveit, K.H., Moustahfid, H., Parcerisas, C.,

- 581 Politikos, D., Sauzede, R., Sokolova, M., Uusitalo, L., Van den Bulcke, L., van Helmond, A.T.M.,
582 Watson, J.T., Welch, H., Beltran-Perez, O., Chaffron, S., Greenberg, D.S., Kuehn, B., Kiko, R., Lo,
583 M., Lopes, R.M., Moeller, K.O., Michaels, W., Pala, A., Romagnan, J.-B., Schuchert, P., Seydi, V.,
584 Villasante, S., Malde, K., Irisson, J.-O., 2023. Machine learning in marine ecology: an overview
585 of techniques and applications. *Ices Journal of Marine Science*, 80(7): 1829-1853.
586 <http://dx.doi.org/10.1093/icesjms/fsad100>
- 587 Sadaippan, B., Balakrishnan, P., Vishal, C.R., Vijayan, N.T., Subramanian, M., Gauns, M.U., 2023.
588 Applications of Machine Learning in Chemical and Biological Oceanography. *ACS Omega*, 8(18):
589 15831-15853. <http://dx.doi.org/10.1021/acsomega.2c06441>
- 590 Schmidtko, S., Stramma, L., Visbeck, M., 2017. Decline in global oceanic oxygen content during the past
591 five decades. *Nature*, 542(7641): 335-339. <http://dx.doi.org/10.1038/nature21399>
- 592 Scully, M.E., 2013. Physical controls on hypoxia in Chesapeake Bay: A numerical modeling study. *Journal*
593 *of Geophysical Research-Oceans*, 118(3): 1239-1256. <http://dx.doi.org/10.1002/jgrc.20138>
- 594 Valera, M., Walter, R.K., Bailey, B.A., Castillo, J.E., 2020. Machine Learning Based Predictions of Dissolved
595 Oxygen in a Small Coastal Embayment. *Journal of Marine Science and Engineering*, 8(12).
596 <http://dx.doi.org/10.3390/jmse8121007>
- 597 Wei, Q., Wang, B., Yao, Q., Fu, M., Sun, J., Xu, B., Yu, Z., 2018. Hydro-biogeochemical processes and their
598 implications for *Ulva prolifera* blooms and expansion in the world's largest green tide
599 occurrence region (Yellow Sea, China). *Science of the Total Environment*, 645: 257-266.
600 <http://dx.doi.org/10.1016/j.scitotenv.2018.07.067>
- 601 Wei, Q., Wei, X., Zhan, R., Liu, L., Zang, J., 2010. Distribution of dissolved oxygen and influence factor in
602 west of Southern Yellow Sea in summer. *Marine Environmental Science*, 29(6): 808-814.
- 603 Wei, Q., Xue, L., Yao, Q., Wang, B., Yu, Z., 2021. Oxygen decline in a temperate marginal sea: Contribution
604 of warming and eutrophication. *Sci Total Environ*, 757: 143227.
605 <http://dx.doi.org/10.1016/j.scitotenv.2020.143227>
- 606 Wenning, R., 2020. THE STATE OF WORLD FISHERIES AND AQUACULTURE (SOFIA) 2020 REPORT.
607 *Integrated Environmental Assessment and Management*, 16(5): 800-801.
- 608 Xiao, C., Chen, N., Hu, C., Wang, K., Gong, J., Chen, Z., 2019. Short and mid-term sea surface temperature
609 prediction using time-series satellite data and LSTM-AdaBoost combination approach. *Remote*
610 *Sensing of Environment*, 233. <http://dx.doi.org/10.1016/j.rse.2019.111358>
- 611 Xin, M., Wang, B., Ma, D., 2013. Chemicohydrographic Characteristics Along the Vertical Section in the
612 Yellow Sea. *Advances in Marine Science*, 31(3): 377-390.
- 613 Xing, X., Wells, M.L., Chen, S., Lin, S., Chai, F., 2020. Enhanced Winter Carbon Export Observed by BGC -
614 Argo in the Northwest Pacific Ocean. *Geophysical Research Letters*, 47(22).
615 <http://dx.doi.org/10.1029/2020gl089847>
- 616 Xiong, T.-q., Wei, Q.-s., Zhai, W.-d., Li, C.-l., Wang, S.-y., Zhang, Y.-x., Liu, S.-j., Yu, S.-q., 2020. Comparing
617 Subsurface Seasonal Deoxygenation and Acidification in the Yellow Sea and Northern East
618 China Sea Along the North-to-South Latitude Gradient. *Frontiers in Marine Science*, 7.
619 <http://dx.doi.org/10.3389/fmars.2020.00686>
- 620 Xu, J., Long, W., Wiggert, J.D., Lanerolle, L.W.J., Brown, C.W., Murtugudde, R., Hood, R.R., 2011. Climate
621 Forcing and Salinity Variability in Chesapeake Bay, USA. *Estuaries and Coasts*, 35(1): 237-261.
622 <http://dx.doi.org/10.1007/s12237-011-9423-5>
- 623 Yang, J., Liu, C., Sun, Q., Zhai, L., Sun, Q., Li, S., Ai, L., Li, X., 2023. Interannual Variability and Long-Term

-
- 624 Trends in Intensity of the Yellow Sea Cold Water Mass during 1993–2019. *Journal of Marine*
625 *Science and Engineering*, 11(10). <http://dx.doi.org/10.3390/jmse11101888>
- 626 Yu, S.-E., Dong, S.-L., Zhang, Z.-X., Zhang, Y.-Y., Sara, G., Wang, J., Dong, Y.-W., 2022. Mapping the potential
627 for offshore aquaculture of salmonids in the Yellow Sea. *Marine Life Science & Technology*, 4(3):
628 329-342. <http://dx.doi.org/10.1007/s42995-022-00141-2>
- 629 Zhang, S.W., Wang, Q.Y., Lue, Y., Cui, H., Yuan, Y.L., 2008. Observation of the seasonal evolution of the
630 Yellow Sea Cold Water Mass in 1996-1998. *Continental Shelf Research*, 28(3): 442-457.
631 <http://dx.doi.org/10.1016/j.csr.2007.10.002>
- 632 Zheng, G., DiGiacomo, P.M., 2020. Linkages Between Phytoplankton and Bottom Oxygen in the
633 Chesapeake Bay. *Journal of Geophysical Research: Oceans*, 125(2).
634 <http://dx.doi.org/10.1029/2019jc015650>
- 635 Zheng, G., Song, J., Dai, J., Wang, Y., 2006. Distributions of chlorophyll A and carbon fixed strength of
636 phytoplankton in autumn of the southern Huanghai Sea waters. *Acta Oceanologica Sinica*,
637 25(3): 68-81.
- 638 Zhu, Z.Y., Wu, H., Liu, S.M., Wu, Y., Huang, D.J., Zhang, J., Zhang, G.S., 2017. Hypoxia off the Changjiang
639 (Yangtze River) estuary and in the adjacent East China Sea: Quantitative approaches to
640 estimating the tidal impact and nutrient regeneration. *Mar Pollut Bull*, 125(1-2): 103-114.
641 <http://dx.doi.org/10.1016/j.marpolbul.2017.07.029>
- 642