

**Phytoplankton and microzooplankton population dynamics along the western area  
from the North Pacific to the Bering Sea in summer**

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## Abstract

Phytoplankton growth and microzooplankton grazing are two critical processes in marine food webs, but they remain understudied in the vast area of the subarctic western Pacific and the Bering Sea. In this study, we measured the phytoplankton growth and microzooplankton grazing rates via the dilution technique to demonstrate their spatial patterns and investigate underlying mechanisms driving the planktonic food web dynamics in these less-explored regions. Our results showed that the phytoplankton growth in these regions was determined by nutrient availability and temperature. In the high nutrient, low chlorophyll regions, iron availability was the primary factor limiting phytoplankton growth. In contrast, phytoplankton growth in the Gulf of Anadyr and Kamchatka Strait was mainly limited by inorganic nitrogen exhausted by the summer blooms. Also, we found that microzooplankton grazing rate was affected by both temperature and prey availability, highlighting the positive effect of temperature. Strong top-down control on phytoplankton by microzooplankton was observed in the Gulf of Anadyr and Kamchatka Strait, indicating a active microbial food web with high turnover rates. In contrast, the decoupling of phytoplankton growth and microzooplankton grazing in the HNLC regions illustrates a weak role of microzooplankton in the marine food web. These results indicated different food web structures in the areas with and without riverine iron input. By revealing the roles of temperature and nutrient or prey availability in regulating the spatial variability of plankton rates, we expect that the plankton will respond differently to ocean warming between the HNLC and coastal regions of the western subarctic Pacific due to different nutrient conditions. Our study contributes to understanding how marine plankton will respond to the projected ocean warming and sea ice declines at high latitudes.

## Introduction

Marine phytoplankton contribute about half of global primary production with less than 1% of the Earth's photosynthetic biomass (Field et al. 1998). The fates of marine primary production are critical to carbon export and global biogeochemical cycling, among which grazing by microzooplankton is a major pathway (Steinberg and Landry 2017). Microzooplankton (<200  $\mu\text{m}$  in size), including ciliates, heterotrophic or mixotrophic dinoflagellates, and metazoan nauplii, are key trophic links of the marine food web and play essential roles in nutrient regeneration. They consume about 59–75% of daily primary production on a global scale (Calbet and Landry 2004). Nevertheless, more accurate quantification of the role of microzooplankton in the food web entails more empirical estimates of microzooplankton grazing activities in various spaces and times (Schmoker et al. 2013). For instance, microzooplankton grazing accounts for about 60% of primary production loss in high latitude seas where the primary production is relatively high (Calbet and Landry 2004). However, this estimate is primarily based on intensive studies in the eastern subarctic North Pacific (e.g., Gulf of Alaska: Strom et al. 2001; eastern Bering Sea: Sherr et al. 2013; Stoecker et al. 2014a; eastern North Pacific: Landry et al. 1993; McNair et al. 2021). There were few estimates in the vast area along the western North Pacific and the western Bering Sea, and the magnitude of microzooplankton grazing on phytoplankton in these regions remains unknown.

The subarctic Pacific is at the end of the global ocean conveyor belt (Broecker et al. 1991), and it is characterized as one of the high-nutrient, low-chlorophyll (HNLC) regions, where the surface macronutrient concentrations are persistently high while the phytoplankton biomass is relatively low (Boyd et al. 2004). The Bering Sea is connected to the subarctic Pacific through ocean circulations. The Alaskan Stream enters the Bering Sea through several Aleutian passes, such as Near Strait, and joins the anti-clockwise Bering Sea Gyre and the

Bering Slope Current. At the same time, an outflow with East Kamchatka Current through the Kamchatka Strait is balanced with the inflows (Aydin and Mueter 2007). As such, the oceanic part of the Bering Sea is hydrographically recognized as an entity with the subarctic Pacific and presents the characteristic of HNLC as well (Banse and English 1999; Nishioka et al. 2021). In the HNLC regions, iron (Fe) limitation is traditionally recognized as the major factor controlling low primary production (Miller et al. 1991). The top-down control from microzooplankton grazing on phytoplankton has been argued as another critical mechanism in maintaining the relatively low phytoplankton biomass (Landry et al. 1993; Strom et al. 2001). As the small phytoplankton are less affected by the Fe limitation, they could be under strong top-down control by microzooplankton herbivory, leading to low phytoplankton biomass in the eastern subarctic Pacific, where small phytoplankton are dominant (Miller et al. 1991).

In contrast to the HNLC basin, the coastal regions of the Bering Sea are highly productive, especially during spring and summer blooms (Banse and English 1999). Although the diatom-dominated primary production during blooms may be more efficiently grazed by mesozooplankton (Hunt Jr. et al. 2011), studies in the eastern Bering Sea have observed significant microzooplankton grazing on spring blooms (Sherr et al. 2013). Nevertheless, the magnitude of microzooplankton grazing in the western Bering Sea (e.g., the Gulf of Anadyr) remains unexplored, which could be different from the eastern area due to distinct hydro-physical processes (Khen et al. 2013). Therefore, we need to measure microzooplankton grazing rates in these regions to further understand the plankton dynamics at high latitudes.

In addition to the magnitude of microzooplankton grazing, understanding what factors drive the spatial and temporal variability of such food web processes is equally important, particularly in the rapidly changing ocean. Recent observations have documented a

significant increase in sea-surface temperature accompanied by a decrease in sea ice in the subarctic Pacific and the western Bering Sea during the past decades (Max et al. 2014; Frey et al. 2022). A recent study also found early retreat and delayed arrival of sea ice in the northern Bering Sea, resulting in warmer surface temperatures in summer (Stabeno et al. 2019). The substantial warming may impact the marine food web as temperature is crucial in determining metabolic rates. Some studies have pointed out that enhanced phytoplankton blooms at high latitudes occur partially because the low temperature constrains the zooplankton grazing activities (Rose and Caron 2007), which may open a ‘loophole’ for the explosion of phytoplankton (Irigoien et al. 2005). The ongoing warming may enhance the microzooplankton herbivory on phytoplankton (Chen et al. 2012), affecting the timing and extent of high-latitude phytoplankton blooms (Archer et al. 2000; Sherr et al. 2013). More empirical evidence is needed to understand the relationship between temperature and plankton rates at high latitudes. Moreover, other factors such as prey concentration and quality also play a role in determining the microzooplankton grazing activities (Hansen et al. 1997; Chen et al. 2012).

In this study, we measured the phytoplankton growth rate and microzooplankton grazing rate via the dilution approach (Landry and Hassett 1982) along the western area of the subarctic North Pacific and the Bering Sea in the summer of 2018. The study region spans a latitude of 10° and is scarcely studied on plankton dynamics. We aim to 1) quantify the magnitude of microzooplankton grazing rate in such less studied regions; 2) demonstrate the spatial patterns of phytoplankton growth rate and microzooplankton grazing rate; 3) identify the determining factors driving the spatial variability of these planktonic food web processes.

## **Materials and Methods**

### ***Study area and environmental parameters measurement***

The study was conducted in the western subarctic Pacific Ocean and the western Bering Sea during the cruise aboard the R/V *Professor Multanovskiy* (belonging to the Far Eastern Hydrometeorological Research Institute, Russia) in summer from July to September 2018. The total 33 stations where we conducted dilution experiments were roughly divided into three water masses according to their locations (Fig. 1). As the Aleutian Arc is the boundary between the North Pacific and the Bering Sea, we treated the stations off the Pacific side of the Kamchatka Peninsula and the Kamchatka Strait as a group of western North Pacific (WNP). We also grouped the stations located in the western Bering Sea basin (WBS) and the stations located in the Gulf of Anadyr (GoA).

Hydrographic data, including temperature and salinity, were obtained at each station by an SBE-911 plus CTD sensor (Sea-Bird Electronics). The incident photosynthetic active radiation (PAR) was measured continuously by an LI-190SB PAR sensor (LI-COR Inc.). The inorganic macronutrient concentrations, including DIN (dissolved inorganic nitrogen, i.e.,  $\text{NO}_3^- + \text{NO}_2^-$  and  $\text{NH}_4^+$ ),  $\text{PO}_4^{3-}$  (P), and  $\text{Si}(\text{OH})_4$  (Si), were measured using an autoanalyzer (QuAatro, BL TEC Co. Ltd.) which were quality-controlled using KANSO reference material (KANSO Co.). The samples for determining dissolved Fe concentration were collected and measured by an FIA chemiluminescence detection system. The details in the sampling and measurement of macronutrients and dissolved Fe were in Nishioka et al. (2020, 2021).

#### ***Phytoplankton and microzooplankton biomass and community measurements***

The phytoplankton biomass was represented by chlorophyll *a* (Chl *a*) concentration, and the size structure of the algal community was characterized by size-fractionated Chl *a* analysis. Aliquots of 250 mL or 500 mL seawater were sequentially filtered through 20  $\mu\text{m}$  polycarbonate membrane filters (GVS Corporation) and GF/F glass fiber filters under a low vacuum. The phytoplankton retained on 20  $\mu\text{m}$  and GF/F filters (nominal pore size: 0.7  $\mu\text{m}$ )

were defined as micro and nano+pico phytoplankton, respectively (Sieburth et al. 1978). All filters were soaked in 6 mL N, N-dimethylformamide (DMF) at -20°C in the darkness for 24h, and the Chl *a* concentration was measured onboard by a Turner Design fluorometer (10-AU, Turner Designs; Welschmeyer 1994; Kobari et al. 2007).

The biomass and community composition of microzooplankton were measured from the Lugol's-preserved samples. Aliquots of 250 mL seawater were preserved with 6 mL acidic Lugol's solution. The samples were stored in amber bottles at room temperature onboard and 4 °C in the laboratory. When analyzing, ten to twenty mL samples were settled by Utermöhl chambers for more than 24 hours and observed entirely using an inverted microscope (Olympus IX51) at 200× magnification. The coastal samples from high Chl *a* environments were diluted (3 to 5 fold) before the settling procedure. All ciliates (aloricate-ciliate and Tintinnids), heterotrophic and mixotrophic dinoflagellates, and metazoan nauplii (<200 µm) were enumerated and taken pictures for dimensions measurements. The cell volumes were calculated by the width and length measured by SPOT software (version3.5) after assigning appropriate geometrical shapes. The biovolumes were then converted into carbon content (µg C L<sup>-1</sup>) by the empirical conversion factors of each microzooplankton category (Table S1).

#### ***Dilution experimental setup***

Dilution experiments were conducted at each station to estimate the phytoplankton growth rate and mortality rate due to microzooplankton grazing (Landry and Hassett 1982). Seawater samples were collected using acid-cleaned 12 L X-Niskin sampling bottles attached to an SBE-911 plus or SBE-32 rosette multi-sampler system (CTD-CMS system; Sea Bird Electronics, Inc.) or an acid-washed plastic bucket. The particle-free seawater was prepared by filtering the natural seawater through a 0.2 µm filter capsule (Pall Corporation) by gravity and added to the clean polycarbonate bottles (1.2 L) to the pre-designed volumes. The bottles

were filled to the full capacity with natural seawater pre-screened with a 200  $\mu\text{m}$  mesh to remove mesozooplankton. Five dilution treatments with the percentages of natural seawater (i.e., the dilution factors) as 10, 25, 50, 75, and 100% were set up at each station. 10% and 100% of these treatments were established in duplicated bottles, while others were in one bottle. Inorganic nutrients were added to all bottles to ensure that nutrients did not limit phytoplankton growth. The final nutrient concentrations in the bottles were 18  $\mu\text{mol L}^{-1}$   $\text{NaNO}_3$ , 2  $\mu\text{mol L}^{-1}$   $\text{NH}_4\text{Cl}$ , 20  $\mu\text{mol L}^{-1}$   $\text{Na}_2\text{SiO}_3$ , and 1.25  $\mu\text{mol L}^{-1}$   $\text{KH}_2\text{PO}_4$ . Another two bottles filled with pre-screened natural seawater without nutrient addition were prepared as controls. All bottles were tightly capped and incubated for 24 h in an on-deck incubator. The light intensity was simulated by covering bottles with neutral screens, and the temperature was controlled by a thermo-controller (EYELA CTP-3000). The subsamples for determining size-fractionated Chl *a* were taken from the initial natural seawater and each bottle after incubating and measured as described above. Before each experiment, all polycarbonate bottles, carboys, tubing, and filter capsules were acid-washed using 10% HCl and rinsed with distilled water, followed by *in situ* seawater.

#### ***Phytoplankton growth rate and grazing mortality rate estimate***

The phytoplankton in each bottle was assumed to grow exponentially with the net growth rate as  $k \text{ (d}^{-1}\text{)} = (1/t) \ln(P_t/dP_0)$ , where  $P_0$  and  $P_t$  were the phytoplankton biomass represented by Chl *a* concentration before and after incubation,  $d$  is the dilution factor (the percentages of natural seawater) of each bottle,  $P_0$  was obtained by multiplying  $d$  with initial Chl *a* concentration of *in situ* unfiltered seawater,  $t$  is the incubation time (1 day). The growth rate with nutrient enrichment ( $\mu_n$ ;  $\text{d}^{-1}$ ) and grazing mortality rate ( $m$ ;  $\text{d}^{-1}$ ) of phytoplankton were derived from the linear regression of net growth rate  $k$  against dilution factor  $d$ . The instantaneous phytoplankton growth rate ( $\mu_0$ ;  $\text{d}^{-1}$ ) was calculated as the sum of the mortality



rate and the net growth rate in the control bottle without nutrient enrichment (i.e.,  $\mu_0 = m + k_{control}$ ).

When saturated grazing was observed (i.e., the net growth rates of least-diluted bottles were leveling off; Gallegos 1989), the  $\mu_n$  and  $m$  were estimated following Chen et al. (2009). Briefly, the Chl *a* concentrations between initial natural seawater ( $P_0$ ) and the least diluted bottle after incubation ( $P_I$ ) were compared to judge whether the saturated grazing occurred under *in situ* conditions or arose from the nutrient enrichments. If  $P_I > P_0$ , this means that phytoplankton do not reach the level for grazing saturation under *in situ* conditions, and the  $m$  equals the slope of the regression plots within the non-saturation range. By contrast,  $P_I < P_0$  means that saturated grazing has already occurred under *in situ* conditions. In this case, the  $m$  was assumed to be the same in the treatments with and without nutrient enrichments and calculated as  $m = \mu_n - k_{n100\%}$ , where  $k_{n100\%}$  is the net growth rate of 100% bottles with nutrient enrichments, and  $\mu_n$  was calculated from the intercept of the regression curve within the non-saturation range. For the experiments with a positive slope of linear regression curve (negative grazing rate) that was not significantly different from 0, the  $m$  was set to 0, and the  $\mu_n$  and  $\mu_0$  were calculated as the average  $k_{n100\%}$  and  $k_{control}$ , respectively (Chen et al. 2009).

The phytoplankton nutrient limitation index was calculated as  $\mu_0/\mu_n$ , which indicates whether the *in situ* nutrients limited the phytoplankton growth (Landry et al. 1995). The lower index indicates more severe nutrient limitation on phytoplankton growth. The impact of microzooplankton grazing on phytoplankton as the percentage of primary production consumed was calculated as %PP consumed =  $m/\mu_0 \times 100\%$  (Calbet and Landry 2004).

### **Statistical analysis**

To investigate the potential factors affecting the phytoplankton growth and microzooplankton grazing rates, Spearman's rank-order correlation analysis was applied to examine the relationships among physical (temperature, salinity, and PAR), chemical (DIN,

P, Si, and DIN:P), and biological (log-transformed Chl *a*, size-fractionated Chl *a*, log-transformed microzooplankton biomass,  $\mu_n$ ,  $\mu_0$ ,  $m$ , biomass-specific grazing rate  $m/B_z$ , and the growth and grazing mortality rates of two size fractions) parameters. Chl *a* concentration and microzooplankton biomass were log-transformed to ensure a normal distribution. To further explore whether microzooplankton grazing and biomass-specific grazing rate are affected by Chl *a* and temperature, we also used multiple linear regression with these two factors as explanatory variables.

### ***Compiling the data of dilution experiments in the subarctic North Pacific and the Bering Sea***

To further understand the plankton population dynamic in the subarctic North Pacific and the Bering Sea, we searched the published data of dilution experiments conducted in the surface water of these regions and compiled a dataset on plankton rates and the corresponding environmental parameters (e.g., temperature, NO<sub>3</sub> concentration, and PAR). We then compared our study regions with other regions based on this dataset (Table S3). In addition, we used the General Additive Models (GAMs, Wood 2006), which employed nonparametric smooth functions to describe the effects of environmental factors, to explore the underlying mechanisms determining phytoplankton growth and microzooplankton grazing at high latitudes (Supplementary Information).

All analyses were implemented using R in version 4.1.2 (R Core Team, 2021). The GAMs were conducted via R function “*gam*” in the package “*mgcv*” (Wood 2006). The multiple linear regression was performed with the R function “*lm*”, and the partial regression plots were created by “*avPlots*” in the package “*car*” (Fox and Weisberg, 2019). The figures were created in R using “*ggplot2*” (Wickham 2016) and “*oceanmap*” (Bauer 2020) packages and the Ocean Data View (Schlitzer 2013).

## Results

We have conducted the experiments from the area east of Kamchatka Peninsula to the Gulf of Anadyr (Fig. 1). The hydrological and biogeochemical properties varied across the three regions (i.e., WNP, WBS, and GoA; Fig. 2), which influences the population dynamics of phytoplankton and microzooplankton in these regions (Figs. 3, 4).

### *Western North Pacific (WNP)*

In the WNP, the inorganic macronutrient concentrations, including DIN, P, and Si, at the stations located off the Pacific side of the Kamchatka Peninsula (St. C04, B03, and B06) were relatively high, while the dissolved Fe was extremely low (Fig. 2, Table S2). At these stations, the average Chl *a* concentration was  $0.78 \pm 0.37 \mu\text{g L}^{-1}$  (mean  $\pm$  sd, the same below unless otherwise indicated). The phytoplankton community was dominated by nano+pico phytoplankton (Fig. 3). By contrast, the macronutrients were low in the Kamchatka Strait, whereas the dissolved Fe was higher (ranges = 0.06 – 0.57 nM) than that at other stations in the WNP (ranges = 0.05 – 0.28 nM, Fig. 2). The Chl *a* concentration was also higher in the Strait ( $1.37 \pm 0.85 \mu\text{g L}^{-1}$ ), with the size community dominated by nano+pico phytoplankton (Fig. 3). The high dissolved Fe and Chl *a* concentrations in the Strait could result from the freshwater inputs indicated by the slightly lower sea surface salinity (Fig. 2b).

The average phytoplankton growth rate in the WNP was  $0.30 \pm 0.23 \text{ d}^{-1}$  ranging from 0.04 to  $0.74 \text{ d}^{-1}$ , which was insignificantly different from those of micro- and nano+pico phytoplankton ( $F_{2,17} = 0.06$ ;  $p = 0.94$ ; Figs. 4, 5a). The growth rate was significantly stimulated by macronutrient enrichments, especially at the stations in the Kamchatka Strait, with an average macronutrient-enriched growth rate of  $0.78 \pm 0.44 \text{ d}^{-1}$  ( $F_{1,12} = 6.52$ ;  $p < 0.05$ ; Figs. 4, 5a).

Microzooplankton biomass ranged from  $15 \mu\text{g L}^{-1}$  to  $60 \mu\text{g L}^{-1}$ , with more contribution by heterotrophic or mixotrophic dinoflagellates (Fig. 3). No significant difference was

observed among the grazing mortality rate of the total phytoplankton community and the two size groups ( $F_{2,17} = 0.06$ ;  $p = 0.84$ ; Figs. 4, 5d). The average  $82\% \pm 73\%$  (0~221%) primary production was consumed by microzooplankton in this area, where a high %PP consumed was observed in the Kamchatka Strait (Table 1).

### ***Western Bering Sea basin (WBS)***

The WBS includes the Kamchatka basin and Aleutian Basin, where the hydrological and nutrient conditions were relatively stable and similar to the Kamchatka basin on the Pacific side (Fig. 2). The average DIN, P, and Si concentrations ( $9.2 \mu\text{M}$ ,  $1.0 \mu\text{M}$ , and  $25.5 \mu\text{M}$ , respectively) were between those found for WNP and GoA, and the dissolved Fe concentration was very low (Fig. 2). Such conditions led to low Chl *a* concentration ( $0.8 \pm 0.27 \mu\text{g L}^{-1}$ ) with major contributions by nano+pico phytoplankton ( $79.4\% \pm 6.6\%$ , Fig. 3). Based on the results of pigment analysis in the same expedition, the haptophytes, pelagophytes, and chlorophytes were the major groups of the phytoplankton community (Fig. 3 in Waga et al. 2022). The growth rates of total phytoplankton and the two size groups ( $> 20 \mu\text{m}$  and  $< 20 \mu\text{m}$ ) were not different from each other with average values of  $0.30 \pm 0.15 \text{ d}^{-1}$ ,  $0.34 \pm 0.21 \text{ d}^{-1}$ , and  $0.26 \pm 0.18 \text{ d}^{-1}$ , respectively ( $F_{2,74} = 1.9$ ;  $p = 0.16$ ; Figs. 4, 5b). Adding macronutrients did not boost their growth rate ( $F_{1,74} = 3.324$ ;  $p = 0.07$ ; Figs. 4, 5b). As such, the  $\mu_o/\mu_n$  was about one at most stations (Fig. S1), indicating that the phytoplankton growth was not limited by the macronutrient in this area.

Microzooplankton biomass ranged from 14 to  $93 \mu\text{g C L}^{-1}$  (Fig. 3), and their average grazing rates on the total phytoplankton community were  $0.07 \pm 0.07 \text{ d}^{-1}$ , which were relatively low compared with WNP and GoA (Figs. 4, 5d). At some stations, no microzooplankton grazing were observed (Fig. 4, Table 1). This low grazing pressure was more frequently observed on micro-phytoplankton in this area (Table 1). The percentage of

primary production consumed by microzooplankton (0 to 86%) was also lower than WNP and GoA (Table 1).

### ***Gulf of Anadyr (GoA)***

The hydrological properties and nutrient conditions in the GoA were very different from the other two regions. The temperature in the GoA ( $8.9 \pm 0.8$  °C) was lower than that in the WNP and WBS (Fig. 2a). The sea surface salinity varied from 28.5 at St. I01 to 32.7 at St. H03, signifying freshwater inputs to the Gulf (Fig. 2b). In contrast to WNP and WBS, the inorganic macronutrient concentrations were extremely low in the GoA, where the DIN and P were depleted at some stations, such as H03 and H05 (Fig. 2c, d). The Si was also low with an average concentration of  $5.9 \mu\text{M}$  (Fig. 2e). In contrast, the dissolved Fe concentration was relatively high, with an average concentration of  $3.03 \pm 1.59 \text{ nM}$  (Fig. 2f). Under such conditions, the Chl *a* concentration was also high ( $2.84 \pm 2.23 \mu\text{gL}^{-1}$ ), ranging from  $0.46 \mu\text{gL}^{-1}$  at St. I07 to  $7.7 \mu\text{gL}^{-1}$  at St. H01 (Fig. 3). The nano+pico phytoplankton were the dominant size fractions at most stations in the GoA ( $74.8\% \pm 26\%$ ), whereas the micro-phytoplankton dominated the phytoplankton communities at St. H01 (67.8%) and I03 (76.9%). At these two stations, abundant diatoms (mainly *Pseudo-nitzschia*) were observed (Fig. S3). The microzooplankton biomass ranged from 20.4 to  $103 \mu\text{gL}^{-1}$ , with the highest value at St. H01 (Fig. 3).

In these macronutrient-depleted regions, the growth rates were negative at most stations except H01, I03, and I05 (Fig. 4). Adding macronutrients significantly drove the growth rates to increase to positive values with the average macronutrient-enriched growth rate of  $0.28 \pm 0.14 \text{ d}^{-1}$  ( $F_{1,18} = 11.74$ ,  $p < 0.01$ ; Fig. 5c). The  $\mu_0/\mu_n$  was less than 0.5 (Fig. S1), signifying a severe nutrient limitation in this region. The average grazing mortality rate of the total phytoplankton community was  $0.20 \pm 0.27 \text{ d}^{-1}$ , which was also not different from the grazing mortality rate of the micro and nano+pico phytoplankton ( $F_{2,27} = 0.16$ ,  $p = 0.85$ ; Fig.

5d). In the GoA, the percentage of primary production consumed by microzooplankton was not calculated at the stations where the phytoplankton growth rates were negative, but for the stations H03, I03, and I05, the values were 183%, 103% and 119%, respectively (Table 1).

### ***Potential environmental variables affecting plankton population dynamics***

The growth rates of the phytoplankton community and nano+pico size group were strongly positively correlated with the macronutrients but negatively correlated with the dissolved Fe (Fig. 6), as the growth rate was low in the macronutrient-depleted but Fe-rich GoA (Fig. 4). The growth rates were also positively correlated with temperature (Fig. 6). While the micro-phytoplankton growth rate was positively correlated with macronutrients but not temperature (Fig. 6). No significant relationship was observed between phytoplankton growth and grazing mortality rate (Fig. 6).

Microzooplankton grazing rate was positively correlated with Chl *a* and microzooplankton biomass (Fig. 6). No significant correlations were observed between grazing rates and physical and chemical parameters, including temperature (Fig. 6). However, the relationship between temperature and grazing rate could be masked by the correlation with Chl *a*. We then conducted multiple linear regression analyses that used both Chl *a* and temperature as explanatory variables to predict the grazing rates. The effect of temperature on grazing rates was significant ( $t = 2.14$ ,  $p < 0.05$  for phytoplankton community;  $t = 3.40$ ,  $p < 0.01$  for nano+pico size group; Fig. 7), suggesting that microzooplankton grazing rate increased with increasing temperature in the study regions. Considering the effect of grazer biomass, we calculated the biomass-specific grazing rate (normalized to microzooplankton biomass,  $m/B_z$ ) and conducted the multiple linear regression analysis. The result showed that  $m/B_z$  also increased with increasing Chl *a* and temperature (Fig. 7).

Based on the GAMs analysis on the compiled dataset of the plankton rates in the subarctic North Pacific and the Bering Sea, including the Chukchi sea, the phytoplankton

growth rate was affected by temperature ( $p < 0.05$ ; Fig. 8),  $\text{NO}_3$  concentration ( $p < 0.001$ ; Fig. 8), and  $PAR$  ( $p < 0.001$ ; Fig. 8), which explained 23.6% variability of the growth rate. The temperature explained the 9.09 % variability of microzooplankton grazing rate ( $p < 0.001$ ; Fig. 8), whereas the effect of Chl  $a$  on microzooplankton grazing rate was not significant ( $p = 0.09$ ; Fig. 8).

## Discussion

The subarctic Pacific and the Bering Sea cover a large area at high latitudes in the northern hemisphere, and their offshore region is one of the HNLC regions. This region is undergoing substantial warming and changes in winter sea ice concentration and is sensitive to climate changes (Stabeno et al. 2019; Max et al. 2014). Such changes could strongly impact the planktonic food web structure and function, which, however, remains understudied in this region. Our study filled gaps of *in situ* data on phytoplankton growth and grazing mortality rates in the less-explored western regions of the subarctic Pacific and the Bering Sea and revealed their distinct spatial patterns. Moreover, we compiled the corresponding data in the subarctic Pacific and the Bering Sea, including the Chukchi Sea, to further identify the pivotal environmental variables driving the spatial variability of plankton population dynamics, which is conducive to our understanding of the impact of environmental changes on planktonic ecosystems at high latitudes.

### ***Phytoplankton growth rates: effects of nutrient availability and temperature***

The phytoplankton growth rate varied among stations, with a sharp contrast observed between two different types of environments, i.e., Fe-limited HNLC regions and Fe-rich Gulf and Strait regions. This contrast was shaped by the impact of nutrient availability.

In the HNLC areas, including the offshore stations in the WNP (St. C04, B03, and B06) and WBS, the phytoplankton growth rate was not limited by macronutrients as the

macronutrients enrichment did not significantly promote their growth (Figs. 4, 5). The phytoplankton growth rate in the offshore stations of WNP was slightly higher than that in the eastern North Pacific (Rivkin et al. 1999). No published data on phytoplankton growth rates are available in the western Bering Sea, but extensive studies have been conducted in the eastern Bering Sea (Fig. 8), where no macronutrient limitation was observed for phytoplankton growth as we found in western regions (Liu et al. 2002; Olson and Strom 2002). These results collectively show that macronutrients were not limited for phytoplankton growth in the HNLC areas. However, dissolved Fe availability may limit the phytoplankton growth rate as it was extremely low in these regions (Fig. 2).

In contrast, the phytoplankton assemblages in the Kamchatka Strait and GoA suffered from severe macronutrient limitations in August. The growth rates were negative at most stations in the GoA and dramatically boosted by the enrichment of macronutrients (Figs. 4, 5). As the macronutrients, especially Si, were deprived, whereas the biomass (i.e., Chl *a*) was very high, this area was experiencing the summer diatom bloom at its late or post phases. At St. H01 and I03, the bloom was in the late phase as we detected positive growth rates and observed highly abundant diatoms (mainly *Pseudo-nitzschia*) (Fig. S3). Conversely, at other stations, the diatoms should have bloomed and probably been grazed by microzooplankton as the bloom proceeded. In this post-bloom period, the phytoplankton community has been reshaped, in which diatoms comprised a small proportion but haptophytes and cryptophytes were the major phytoplankton groups (the results from the same expedition; Fig. 3 in Waga et al. 2022). The blooms in the Kamchatka Strait were initiated by the Fe-rich freshwater input from the Kamchatka Peninsula to the Strait, and its influence expanded to adjacent coastal waters (Nishioka et al. 2021). In the GoA, the mixing of macronutrient-rich oceanic waters and the Fe-rich freshwater from the Anadyr River boosted the growth of phytoplankton and



fostered the spring-summer bloom (Sorokin, 1999), fueling the higher trophic levels of the food web.

Our study captured the dynamic features of plankton in the late- or post-phase of the bloom initiated by Fe input, as well as the features in the relatively stable HNLC waters where Fe was limited. These results revealed that the effect of macronutrients on phytoplankton growth in our study regions was mediated by Fe availability. Throughout the subarctic North Pacific, the Bering Sea, and the Chukchi Sea, the effect of macronutrients on phytoplankton growth was also significant, whereas the growth rate tended to level off at high nitrogen concentrations (Fig. 8), indicating that the effect of Fe availability overrides macronutrients in the HNLC regions. As such, both macronutrients and micronutrients are critical drivers of phytoplankton growth at high latitudes.

In addition to nutrient availability, temperature also influences phytoplankton growth. In our study region, temperature was significantly positively correlated with phytoplankton growth rate (Fig. 6). The results of the compiled dataset further confirmed the significant effect of temperature on phytoplankton growth (Fig. 8). Consequently, the low temperature contributed to the low phytoplankton growth rate in the GoA. When adding macronutrients to satisfy the phytoplankton growth, the macronutrient-enriched phytoplankton growth rates in the GoA were still much lower than that in the Kamchatka Strait where the temperature was higher (Figs. 4, 5). Despite the low growth rate, massive algal blooms frequently occur in the GoA partially because the low-temperature constraint on the zooplankton grazers might be more severe (Rose and Caron 2007), which opens a ‘loophole’ for the explosion of phytoplankton (Irigoien et al. 2005). Nevertheless, more investigation is still needed to understand the role of temperature in shaping algal blooms and plankton dynamics at high latitudes.

***Microzooplankton grazing rate: bottom-up controls of prey availability and temperature***

The microzooplankton grazing rates in the WNP ( $0.18 \pm 0.12 \text{ d}^{-1}$ ) were almost identical to the previous estimates in the same regions (Liu et al. 2002) and similar to those in the eastern North Pacific (Rivkin et al. 1999; McNair et al. 2021). In the WBS, the grazing rate ( $0.07 \pm 0.07 \text{ d}^{-1}$ ) was lower than in the WNP and the southeastern Bering Sea (Olson and Strom 2002). The lower grazing rate in the WBS can be attributed to the low Chl *a* concentration and possibly poor food quality stemming from the limited Fe concentration in the HNLC regions (Strom and Fredrickson 2008). Conversely, higher microzooplankton grazing rates were observed at high Chl *a* stations of GoA and Kamchatka Strait. Therefore, the prey availability represented by Chl *a*, along with food quality, should be regarded as the critical factors that determine microzooplankton grazing activities (Fig. 6).

In addition, the prey availability affects the grazing activities of individual grazers as the  $m/B_z$  positively correlated with Chl *a* concentration (Fig. 6). The  $m/B_z$  can be considered as the per capita clearance rate of grazers, which usually decreases with increasing prey concentration based on the functional responses, especially the Ivlev and Holling II functions (Holling 1959; Ivlev 1975; Chen et al. 2012). However, the  $m/B_z$  increased with increasing Chl *a* in our study regions (Fig. 7). This pattern can be explained by the increasing part of the Holling III function that is unimodal with a peak at low prey concentrations (Liu et al. 2021). Furthermore, it is also likely due to a shift of community composition towards more active grazers when Chl *a* increased. When expanding the dataset to the whole subarctic North Pacific, the Bering Sea, and the Chukchi Sea, the effect of Chl *a* concentration on grazing rate was insignificant (Fig. 8), which was also observed in other (sub)arctic waters (e.g., Lawrence and Mender-Deuer 2012). The results imply that other factors (e.g., prey quality and grazer selectivity), in addition to prey concentration, may affect the microzooplankton grazing rates at high latitudes.

Temperature is another vital factor affecting microzooplankton grazing (Chen et al. 2012; Liu et al. 2019). However, some studies found no relationship or an inverse relationship between temperature and microzooplankton grazing (Verity 1986; Menden-Deuer et al. 2018; Lavrentyev et al. 2019; Marrec et al. 2021). However, the temperature effect could be confounded by prey concentration or predator biomass with the consequence of no or negative correlations (e.g., Fig. 6). In our study, we used multiple linear regression analysis to consider both effects of temperature and prey concentration and found a positive correlation between temperature and microzooplankton community herbivory as well as the individual clearance rate (i.e.,  $m/B_z$ ; Fig. 7). Likewise, temperature positively influenced the microzooplankton grazing rate in the subarctic and polar regions based on GAM including the effect of prey concentration (Fig. 8). As such, in the warmer ocean, the grazing activities of microzooplankton on primary production will be enhanced by warming with more energy and material transferred to the higher trophic levels. It has been found that the temperature sensitivity of microzooplankton growth and grazing is much higher than phytoplankton (Chen et al. 2012). Rose and Caron (2007) attributed the algae blooms in cold waters to the stronger low-temperature constraint of microzooplankton over phytoplankton. Similarly, rising temperature enhances microzooplankton herbivory to a greater extent than phytoplankton growth elevates, which may accelerate the blooming progress and change the food web dynamics in high-latitude oceans. Nevertheless, the effect of temperature on predator-prey interaction and food web dynamics is more complex than the clear temperature dependence of plankton (Synodinos et al. 2021). For example, although the temperature sensitivity of per capita grazing rate of microzooplankton can be greater than that of phytoplankton growth rate, the biomass of microzooplankton will also depend on temperature, which may actually lead to a temperature independent grazing rate of the whole

microzooplankton community. More investigations are imperative to unveil the impact of projected ocean warming on the high-latitude marine food web and ecosystems.

***The balance between phytoplankton growth and microzooplankton grazing with implications of food web structure***

The grazing activities of microzooplankton were more active with greater grazing pressure on phytoplankton in the Fe-rich Kamchatka Strait and GoA, which was assessed by the  $m/\mu_0$  calculated as %PP consumed (Calbet and Landry 2004). The low (<100%) and high (>100%) %PP consumed indicate the decoupling of growth and grazing with weak and strong grazing pressure from microzooplankton, respectively. In the GoA, although the negative growth rates at most stations make the estimation of %PP consumed difficult, the average %PP consumed of the three stations with positive growth rates (H05, I03, I05) was high (Fig. 8), indicating a strong grazing pressure on phytoplankton from microzooplankton. At these stations, the phytoplankton communities were dominated by large phytoplankton (> 20  $\mu\text{m}$ , mainly diatoms; Figs. 3; S3), which had been thought to be primarily consumed by mesozooplankton such as copepods. Nevertheless, our results confirmed that the grazing impacts from microzooplankton were as or more significant as mesozooplankton on large diatoms. It has been increasingly recognized that heterotrophic dinoflagellates were voracious predators on diatoms (Sherr and Sherr 2007). Ciliates have been found to prey on large-sized diatoms in the eastern Bering Sea (Sherr et al. 2013). Therefore, both heterotrophic dinoflagellates and ciliates are important predators on large phytoplankton such as diatoms. In addition, the top-down control of microzooplankton grazing on phytoplankton was also substantial in the Kamchatka Strait, where the average %PP consumed was 139% (Fig. 8). Collectively, the Fe-rich freshwater input drove the rapid growth of phytoplankton, followed by vigorous grazing activities, which made the planktonic food web more active.

By contrast, the %PP consumed at the offshore stations of WNP and WBS was relatively low, with an average of 40% and 35%, respectively, indicating a decoupling between phytoplankton growth and microzooplankton grazing and relatively weak grazing pressure from microzooplankton (Fig. S2). Such decoupling has also been observed in other HNLC regions (e.g., eastern Bering Sea basin; Fig. 8). In these regions, phytoplankton growth rate exceeded microzooplankton grazing, whereas the phytoplankton biomass did not accumulate as Chl *a* concentration was low. One possible explanation is a possible trophic cascade arising from the exclusion of higher trophic levels in the incubation bottles. The microzooplankton were released from grazing pressure by mesozooplankton that were removed artificially, which in turn enhanced microzooplankton predation on nanoflagellates and relieved the grazing of heterotrophic nanoflagellates on small phytoplankton (Marrec et al. 2021). In the current study, we conducted the same treatment at all stations, however, the decoupling between growth and grazing in the HNLC regions suggested that the trophic cascade was more severe and the protistan predatory chains could be longer. The role of heterotrophic nanoflagellates in consuming small phytoplankton could be more significant in the Fe-limited HNLC regions than in the Fe-rich Strait and Gulf.

In addition, the mixotrophs in plankton communities may also cause the growth-grazing imbalance. The mixotrophs may use their inherent or acquired phototrophy to supplement the energy demands when suffering from low food concentration and poor food quality, which reduces their grazing activities and results in low microzooplankton grazing rates (Stoecker et al. 2017). It has been found that the Chl *a* of mixotrophic ciliates was sometimes higher than 50% of total Chl *a* in the eastern Bering Sea, implying a significant role of mixotrophs in plankton communities at high latitudes (Stoecker et al. 2014b). Nevertheless, it remains a challenge to quantify the proportion of mixotrophs in the plankton community and their impact on the grazing rate estimates (Stoecker et al. 2017).

The balance between phytoplankton growth and microzooplankton grazing illustrates the role of microzooplankton in marine food webs with implications for the planktonic food web structure (Calbet and Landry 2004). In the more stable and Fe-limited HNLC regions, the decoupling between growth rate and grazing rate with low %PP consumed suggests a more complicated food web structure involving nanoflagellates and mixotrophs. By contrast, in the GoA and Kamchatka Strait, the Fe input via freshwater discharges not only triggered the rapid growth of phytoplankton but also enhanced the grazing activities of microzooplankton, driving the microbial food web more active with high turnover rates. Microzooplankton in these regions is a crucial vehicle for transferring organic matter and energy to the higher trophic levels. In fact, the strong top-down control of microzooplankton was observed more frequently in the coastal regions with high phytoplankton biomass at high latitudes (Fig. 8). For instance, high %PP consumed values in the southeastern Bering Sea occurred during phytoplankton blooms (Strom et al. 2008). As such, in these regions, the role of microzooplankton is more significant, with more contributions to the energy transfer from primary production to higher trophic levels.

### **Concluding remarks and perspectives**

Our results revealed the variability in spatial patterns of phytoplankton growth and grazing mortality rates and the potential mechanisms driving such spatial variability in the less-explored western area of the subarctic Pacific and the Bering Sea. In the HNLC regions, the phytoplankton growth rate was relatively low and presumably limited by Fe availability, while the mortality due to microzooplankton grazing was also low, resulting in weak top-down controls on phytoplankton. By contrast, in the Gulf of Anadyr and Kamchatka Strait, where freshwater input brought a large amount of Fe, phytoplankton growth was limited by macronutrients (mainly N limitation) under the late-bloom or past-bloom conditions, together with strong top-down control by microzooplankton grazing. Collectively, in high-latitude

waters, phytoplankton growth was determined by temperature, nutrient availability, and light intensity. Particularly, in the HNLC regions, the effect of macronutrients was mediated by Fe availability, and the blooms could be stimulated by Fe input, accomplished by more active microzooplankton herbivory. Through multiple linear regression analysis, we also identified the effect of temperature and prey availability on microzooplankton grazing rate at high latitudes. These insights into the crucial environmental variables driving the variability of plankton population dynamics advance our understanding of how plankton will respond to environmental changes.

The subarctic and polar regions are already under substantial warming. In the HNLC regions, warming enhances the stratification of water column (Behrenfeld et al. 2006). Without an additional Fe source, phytoplankton community will grow slowly and be dominated by small phytoplankton, along with less active microzooplankton grazing, which will result in a reduction of carbon sequestration capacity and fisheries production. While in the Gulfs and Straits with Fe-rich freshwater input, the early retreat of sea ice and warmer sea surface temperature may bring forward phytoplankton blooms and enhance microzooplankton herbivory, leading to an increase in energy transfer to higher trophic levels, shifts in food web structures, and accelerating carbon cycling in high-latitude waters.

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## Figure legends

**Fig. 1** Map of experimental stations. The rectangles show three study regions (from below to top): Western North Pacific (WNP), Western Bering Sea Basin (WBS), and the Gulf of Anadyr (GoA).

**Fig. 2** Spatial distributions of (a) sea surface temperature (SST, °C); (b) sea surface salinity; (c) total dissolved inorganic nitrogen concentration (DIN, including  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and  $\text{NH}_4^+$ ,  $\mu\text{mol L}^{-1}$ ); (d)  $\text{PO}_4^{3-}$  concentration (P,  $\mu\text{mol L}^{-1}$ ); (e)  $\text{Si(OH)}_4$  concentration ( $\mu\text{mol L}^{-1}$ ); (f) dissolved Fe concentration at the depth of 10 m (D-Fe,  $\text{nmol L}^{-1}$ ).

**Fig. 3** The spatial patterns of plankton in the study regions: (a) Chl *a* concentration ( $\mu\text{g L}^{-1}$ ); (b) microzooplankton abundance ( $\text{cell ml}^{-1}$ ); and (c) microzooplankton biomass ( $\mu\text{g C L}^{-1}$ ). The community composition of (d) phytoplankton; (e) microzooplankton abundance, and (f) microzooplankton biomass. The phytoplankton community includes microphytoplankton (Chl *a* > 20  $\mu\text{m}$ ) and nano+pico phytoplankton (Chl *a* < 20  $\mu\text{m}$ ), and the microzooplankton consist of three categories: dinoflagellates, tintinnids, and aloricate-ciliates. The dinoflagellates include the heterotrophic and mixotrophic dinoflagellates.

**Fig. 4** Spatial variations of growth rate and mortality rate due to microzooplankton grazing of micro-phytoplankton (Chl *a* > 20  $\mu\text{m}$ ) (a, d), nano+phytoplankton (Chl *a* < 20  $\mu\text{m}$ ) (b, e), and total phytoplankton community (c, f).

**Fig. 5** Boxplots of nutrient enriched growth rate ( $\mu_n$ ,  $\text{d}^{-1}$ ), instantaneous growth rate ( $\mu_0$ ,  $\text{d}^{-1}$ ), of phytoplankton community and the two size groups (micro, nano+pico-phytoplankton) in the three regions: (a) Western North Pacific; (b) Western Bering Sea basin, and (c) Gulf of

Anadyr (GoA). (d) Boxplot of microzooplankton grazing rate ( $m$ ,  $d^{-1}$ ) on phytoplankton community and the two size groups in the three regions. The asterisk indicates whether the effect of adding macronutrients was significant (one-way ANOVA;  $*p < 0.05$ ;  $**p < 0.01$ ;  $***p < 0.001$ ).

**Fig. 6** Correlation matrix among physical (Temp, Sal, PAR), chemical (DIN, P, Si, DIN:P, DFe), and biological (log-transformed Chl  $a$ : lnChl $a$ , size-fractionated Chl  $a$ , log-transformed microzooplankton biomass: lnBz, phytoplankton growth rate:  $\mu_0$ , and microzooplankton grazing rates:  $m$ , and biomass-specific grazing rate  $m/B_z$ ) parameters. %Chl $a$  represents the percentage of size fractions (indicated by micro and nano+pico in parentheses) account for the total Chl  $a$  concentration. The colour and colour intensities indicate the direction (green = negative, purple = positive) and magnitude of Spearman's rank-order correlation coefficient, respectively. The asterisks showing the significant levels of the correlations ( $*p < 0.05$ ;  $**p < 0.01$ ;  $***p < 0.001$ ).

**Fig. 7** The partial relationship between microzooplankton grazing rate ( $m$ ) and Chl  $a$  concentration (a), and temperature (b). The black dots are the total phytoplankton, and the blue open dots are the nano+pico phytoplankton. The partial relationship between biomass-specific grazing rate ( $m/B_z$ ) and Chl  $a$  (c) and temperature (d).

**Fig. 8** (a) The percentage of primary production consumed by microzooplankton (%PP consumed) in the subarctic North Pacific, the Bering Sea, and the Chukchi Sea (45°N - 80 °N) based on a compiled dataset including the studies: [1] Landry et al. (1993); [2] Leising et al. (2005); [3] Liu et al. (2002); [4] Liu et al. (2009); [5] Menden-Deuer and Fredrickson (2010); [6] Olson and Strom (2002); [7] Rivkin et al. (1999); [8] Sherr et al. (2009); [9]

814 Strom and Welschmeyer (1991); [10] Strom et al. (2001); [11] Strom et al. (2007); [12]  
815 Strom et al. (2008); [13] Yang et al. (2014); [14] Sherr et al. (2013); [15] Connell et al.  
816 (2018); [16] this study. Partial effects of (b) temperature, (c)  $\ln(NO_3)$ , and (d) *PAR* on relative  
817 phytoplankton growth rate ( $\mu_0$ ). Partial effects of (e) temperature and (f)  $\ln(Chl)$  on relative  
818 microzooplankton grazing rate (*m*) based on the compile dataset. The black dots are the data  
819 from the abovementioned studies and the red dots are the data of this study. Shaded areas  
820 denote 95% confidential intervals.