

Highlights

Electrochemical sensing of SARS-CoV-2 amplicons with PCB electrodes

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- Adsorption of PCR amplicons electrochemically detected on low-cost PCB electrodes with methylene blue (MB)
- Amplicon length and MB concentration are both critical to achieving linear sensor performance
- Detection of SARS-CoV-2 nucleocapsid gene amplicons from 10 pg/ μ l (1.7 fM)
- N1 fragment successfully amplified and detected in wastewater spiked with SARS-CoV-2 control RNA

Electrochemical sensing of SARS-CoV-2 amplicons with PCB electrodes

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Abstract

We present a low-cost electrochemical DNA biosensor based on printed circuit board (PCB) electrodes for wastewater monitoring using portable PCR instruments, such as miniPCR®, without the requirement for qPCR reagents. PCB electrodes are attractive candidates for low-cost and sensitive DNA biosensors of relevance in a pandemic such as COVID-19, and facilitate the opportunity to map disease spread in Low-Middle Income Countries (LMICs) through monitoring of environmental samples such as wastewater.

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The biosensor reported in this work is capable of detecting PCR amplicons through the intercalation of methylene blue (MB) with DNA, which increases the voltammogram peak current at the redox potential of MB. We describe how these changes are likely to result from the adsorption of MB-DNA complex on the electrode surface. The electrodes are reusable, easy to clean, do not undergo any surface modification and represent a cost-effective solution with long shelf-life. We also explore the impact that MB concentration and DNA length have upon our biosensor performance and provide insights useful to other investigators in the field. The sensor reported here is capable of detecting SARS-CoV-2 nucleocapsid gene amplicons at concentrations as low as 10 pg/μl (approximately 1.7 fM) and can detect nucleotides amplified after 10 PCR cycles. Furthermore, using the PCB electrode and approaches described here, SARS-CoV-2 amplicons were detected in simulated wastewater sample, by spiking wastewater collected from a sewage treatment plant in Mumbai, India with SARS-CoV-2 RNA.

Keywords: Methylene blue, DNA biosensor, PCB electrode, SARS-CoV-2, Wastewater epidemiology

1. Introduction

The COVID-19 pandemic has negatively impacted economies and communities globally since its spread in early 2020. One of the most effective ways to minimize and control the spread of the SARS-CoV-2 virus has been to limit close human contact and therefore social activity [1, 2]. Whilst this approach has been highly effective in reducing levels of spread, the overall economic and societal impact is extremely high. Furthermore, these measures

are difficult to meaningfully implement in many low-middle income countries (LMICs) where large proportions of the population are dependent upon informal labour for income and survival [3]. Another challenge in many LMICs is that the advanced laboratory facilities and skilled workforce required to achieve mass human testing cannot be implemented in the short timescales required to support contact tracing and thus reduce spread whilst allowing restrictions on person to person economic activities to take place. An alternative to molecular testing is the use of rapid, lateral flow immunochemistry testing, but these tests have lower sensitivity and specificity than gold standard qPCR testing [4, 5]. On the other hand, testing of wastewater for the presence of SARS-CoV-2 nucleic acid could be used as a surveillance tool to identify areas where the case numbers are likely to be increasing and therefore allow more targeted action to be taken to limit viral spread in specific regions.

Wastewater epidemiology has been widely recognized and researched as a tool for surveillance and management of the spread of SARS-CoV-2 [6–14]. Over the last year, several groups have used gold standard qPCR-based techniques in order to pilot surveillance methods globally including in India, Italy, and North America [15–18]. One potential problem is that qPCR still requires expensive laboratory infrastructure and skilled scientists or technicians to complete the assay. Furthermore, if resources are limited, testing of human samples would most likely take precedence over wastewater epidemiology surveillance. Therefore, lower cost, alternative approaches are required in order to support wastewater surveillance for SARS-CoV-2 in LMICs, thus supporting widespread adoption of a further tool in the fight against the

33 virus. One way to support this would be through the use of low-cost nucleic
34 acid biosensors. In addition to serving as tools for data collection from popu-
35 lations which lack adequate access to healthcare and diagnostic testing, such
36 technologies could potentially also serve as more accurate indicators of the
37 scale of infection by including asymptomatic cases and recovered individuals
38 not accounted in clinical tests, that continue to shed virions in feces.

39 The field of biosensors for nucleic acid and in particular DNA is well es-
40 tablished (see [19] for a detailed review). Methylene Blue (MB) is a redox
41 mediator that can be transduced with an electrochemical sensor. The inter-
42 calating properties of MB with DNA have resulted in its widespread use as a
43 redox mediator in a range of electrochemical DNA sensor approaches [20–24].
44 These systems conventionally employ electrodes to immobilize the MB-DNA
45 complex [24–28] or probe the DNA concentration in homogenous solution di-
46 rectly [29–33]. A common feature of all these studies is the prevention of DNA
47 adsorption onto the electrode surface. In the case of gold electrodes, this
48 has been achieved using an alkanethiol layer, such as 6-mercapto-1-hexanol
49 (MCH) [24, 34]. However, electrode modifications such as this can lead to
50 more onerous electrode storage requirements prior to use. Also, the use of
51 modification processes increases the complexity and thus production costs
52 of the electrode. An alternative approach, is to use the ability of DNA to
53 adsorb to the electrode surface as the basis for identification of PCR product.

54 The use of printed circuit boards (PCBs) conventionally used in the elec-
55 tronics manufacturing industry has been explored for Micro Total Analysis
56 Systems (μ TAS) platforms since early-1990s, however the focus for a large
57 part of the previous three decades has been on demonstration of proof-of-

58 concept PCB based lab-on-chips, and not particularly on integration strate-
 59 gies and manufacturability. In recent times the cost of μ TAS platforms has
 60 emerged as a significant technology driver, and therefore lab-on-PCB plat-
 61 forms are seeing renewed interest in the broad biosensors community [35–42].
 62 Despite this, the technology is not as commonly adopted as screen printed
 63 electrodes, mainly due to challenges associated with impact of gold surface
 64 roughness on biorecognition element immobilization and the need for ad-
 65 ditional processing steps to ensure reliability due to the use of copper in
 66 PCBs, which is susceptible to corrosion and hampers electrochemical analy-
 67 sis [43, 44]. With suitable development, PCB electrode based electrochemical
 68 DNA sensors could be of great relevance in the COVID-19 pandemic [45] e.g.
 69 for mapping disease spread through monitoring of wastewater. In this paper,
 70 we build on the benefits of PCB based technologies and present a low-cost
 71 PCB electrode based electrochemical sensor for detection of SARS-CoV-2
 72 PCR amplicons. The transduction mechanism is based on intercalation of
 73 MB with DNA, and change in current through voltammetry measurements
 74 resulting from adsorption of MB-DNA complex on the electrode surface. The
 75 PCB electrodes are manufactured using the conventional electroless nickel
 76 immersion gold (ENIG) process and do not undergo any surface modifica-
 77 tion. The electrodes thus present a cost-effective solution with long shelf-life,
 78 as they do not require any specific low-temperature storage conditions for
 79 normal operation, unlike electrodes with DNA immobilization or hybridiza-
 80 tion. The system design and approach can thus be implemented at very
 81 low cost with minimal resources, and provides a basis for a future low cost
 82 SARS-CoV-2 wastewater surveillance system (as illustrated in Figure 1(a)).

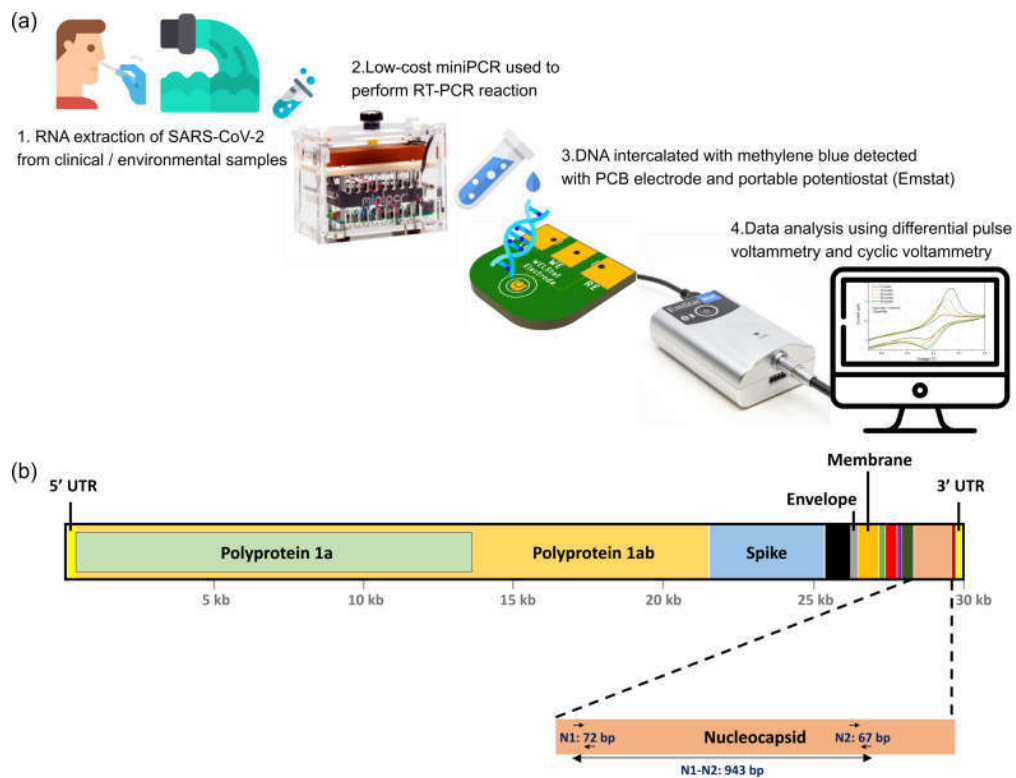


Figure 1: (a) SARS-CoV-2 nucleic acid detection workflow. RNA extracted from wastewater is amplified using a low cost thermo cycler and then placed onto PCB electrodes. (b) Illustration of major protein coding genes of the SARS-CoV-2 coronavirus genome. The 943 bp N1-N2 fragment produced using N1 forward and N2 reverse primers used in this present study is also shown.

83 2. Materials and methods

84 2.1. Preparation of target amplicons

85 Control plasmid containing complete nucleocapsid (N) gene from SARS-
86 CoV-2 isolate Wuhan-Hu-1 (GenBank: NC 045512.2; 2019-nCoV_N_Positive
87 Control from IDT, supplied as 2×10^6 copies/ μ l) was used as template (10^2 -
88 10^6 copies per reaction) for amplifying the target fragment using mini8 PCR
89 machine (miniPCR®) with the US CDC-recommended N1 and N2 forward
90 and reverse primers (IDT). Fragments amplified included the N1 fragment (72
91 bp), and N1-N2 fragment (943 bp, using N1 forward and N2 reverse primers),
92 as illustrated in Figure 1(b). In some instances, PCR reactions were stopped
93 after fewer cycles (5, 10, 20, and 30 cycles) followed by a final extension of
94 10 min at 72 °C. The concentration of the PCR-amplified products was es-
95 timated using a Nanodrop spectrophotometer. For some measurements, the
96 products were diluted (in Milli-Q water) to final concentrations of 10 pg/ μ l,
97 20 pg/ μ l, 50 pg/ μ l, 100 pg/ μ l, 150 pg/ μ l and 200 pg/ μ l. The template concen-
98 tration chosen for amplification is 10^5 copies/ μ l unless otherwise mentioned.
99 For measurements performed with DNA-MB complex, desired concentration
100 of MB (5 μ M, 10 μ M, 50 μ M or 100 μ M) were added to the PCR product.
101 The DNA-MB samples were incubated at room temperature for 1 h before
102 performing measurements, to allow sufficient time for intercalation [21].

103 2.1.1. Generation of amplicons using different template concentrations

104 PCR using different template concentrations was performed to evaluate
105 lower limit of sensing using PCB electrodes. 2019-nCoV_N_Positive Control
106 from IDT was used as a template (10^2 - 10^5 copies per reaction) for amplifying

107 the target fragment using the mini8 thermal cycler (40 cycles) with CDC-
108 recommended N1 forward and N2 reverse primers (IDT). The amplicons were
109 visualized by agarose gel electrophoresis and ethidium bromide staining.

110 *2.1.2. Preparation of intermediate PCR amplicons*

111 2019-nCoV_N_Positive Control diluted to 10^5 copies/ μ l is used as template
112 for PCR to amplify the 943 bp N product. The PCR reactions are stopped
113 after 5, 10, 20, 30 and 40 cycles. Products obtained with <40 cycles undergo
114 a final extension of 10 min at 72 °C in dry bath. The final amplified product
115 (40 cycles) is left in mini8 for final extension. To verify the presence of PCR
116 products, a 4 μ l volume of each sample was electrophoresed on a 1 % agarose
117 gel.

118 *2.1.3. Generation of simulated wastewater sample for detection of SARS-* 119 *CoV-2 amplicons*

120 200 ml of grab wastewater samples were collected from a sewage treat-
121 ment plant in Mumbai, India. The samples were heat-inactivated by incu-
122 bating at 60 °C for 90 min, and the virus particles were concentrated using
123 aluminum hydroxide adsorption method reported by Randazzo et al. [46].
124 Briefly, sample pH was adjusted to 6.0 before the addition of AlCl_3 solu-
125 tion. The pH was readjusted to 6.0 and the samples were incubated on a
126 shaker at 150 rpm for 15 min at room temperature followed by centrifugation
127 at 1700 g for 20 min to concentrate the viruses. The pellet was resuspended
128 in 3 % beef extract and mixed slowly using a shaker at 150 rpm for 10 min at
129 room temperature. The samples were then centrifuged at 1900 g for 30 min,
130 and the pellet was resuspended in 1 ml of PBS. Total viral RNA was ex-

131 tracted from 150 μ l of the resuspended pellet using the AllPrep PowerViral
 132 DNA/RNA kit (Qiagen, Germany) as per the manufacturer’s instructions.
 133 RNA eluted in nuclease-free water was stored as aliquots at -80°C . The
 134 presence of SARS-CoV-2 RNA was checked by performing RT-PCR with
 135 CDC-recommended N1 and N2 primers (IDT). As the sample did not con-
 136 tain any SARS-CoV-2 RNA, a simulated sample was generated by mixing the
 137 isolated total RNA with approximately 4.73×10^3 copies of the SARS-CoV-2
 138 genomic RNA (VR-1986DTM, ATCCTM). The RNA was reverse transcribed
 139 using a cDNA kit (Genetix Biotech Asia Pvt. Ltd., India) to amplify the
 140 72 bp long N1 fragment.

141 *2.2. Electrochemical biosensor*

142 The PCB electrodes (Figure 2(a)) were designed using Autodesk EA-
 143 GLE software and manufactured in conventional ENIG plating process (Cir-
 144 cuit Systems (India) Ltd.). The manufacturing cost per electrode is ap-
 145 proximately USD \$0.55 (i.e. approximately INR ₹40). The electrodes were
 146 cleaned with an isopropyl alcohol (IPA) soaked lint-free wipe before each
 147 measurement. During each measurement, 5 μ l of the sample to be tested
 148 was dispensed on the electrode. The electrode was then cleaned with a
 149 lint-free wipe dampened with IPA before dispensing the next sample. The
 150 voltammetry measurements were performed using Palmsens EmStat3 Blue
 151 potentiostat. The potentiostat configuration and data acquisition are per-
 152 formed using PSTrace software. Differential pulse voltammetry (DPV) mea-
 153 surements were performed using following settings: equilibration time = 8 s,
 154 voltage step = 3 mV, pulse voltage = 25 mV, pulse duration = 50 ms and
 155 scan rate = 20 mV/s. The potential (voltage) range is set as desired for each

156 measurement. Cyclic voltammetry (CV) measurements were performed at
157 various scan rates using following settings: equilibration time = 8 s and volt-
158 age step = 3 mV. The voltage range and scan rate were set as desired for
159 each measurement. The equilibration time is the time during which the first
160 potential of the measurement is applied to the electrochemical cell without
161 recording the current. This is done in order to exclude initial capacitive cur-
162 rent from interfering with the Faradaic current to be measured. Peak heights
163 for DPV peak current and cathodic peak current in CV voltammograms were
164 obtained using PSTrace software. The values for peak height obtained from
165 PSTrace software were used for preparation of graphs wherever peak current
166 values are discussed.

167 3. Results and discussion

168 3.1. PCB electrodes can be reused multiple times after cleaning with IPA-wipe

169 The gold electrodes on the PCB were formed through an industry stan-
170 dard ENIG plating process, that results in a thin (<100 nm) gold layer on
171 top of a thicker nickel layer (the nickel layer is designed to prevent diffu-
172 sion of copper and gold layers). The nickel and copper layers underneath
173 the gold are susceptible to corrosion and therefore incompatible with con-
174 ventional gold electrode cleaning methods such as sulfuric acid potential cy-
175 cling or piranha clean [47]. Since the samples to be measured comprise of
176 organic compounds (MB and DNA), we clean the PCB electrodes with lint-
177 free wipe dampened with IPA. Repeat measurements of 100 μ M MB using
178 this approach show that consistent peak DPV currents are observed (Fig-
179 ure 2(b)). Following measurements with 100 μ M MB, no DPV peaks were

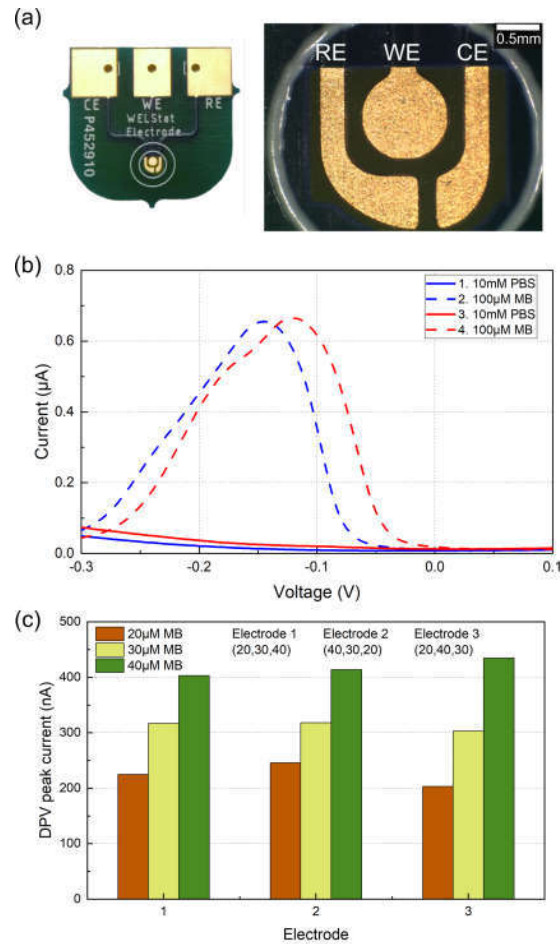


Figure 2: (a) Left - photograph of the PCB electrode used in this work. Right - optical micrograph of the PCB electrode showing reference electrode (RE), working electrode (WE), and counter electrode (CE), all formed with ENIG finish gold pads on the PCB. (b) The efficacy of cleaning process of the electrode is illustrated by measuring 10 mM PBS and 100 μ M methylene blue (MB) successively in the order indicated in the figure legend, by wiping the electrode with IPA dampened wipe after each measurement. The first PBS measurement is performed to ensure there is no contamination on the electrode. No DPV peaks are seen in the third measurement performed with PBS, indicating that any MB adsorbed on the electrode is wiped away. (c) The peak current is proportional to the concentration of MB independent of the order of addition of MB solutions with different concentration (20 μ M, 30 μ M and 40 μ M) on three electrodes (E1, E2, E3; order in which samples are dispensed is indicated in plot legend).

180 seen in the measurements performed with only PBS. This indicates that any
 181 residual MB on the electrode is wiped away through the cleaning process
 182 and provides confidence that the cleaning approach adopted does not nega-
 183 tively affect the thin gold layer on the electrode surface. DPV measurements
 184 were performed with various concentrations of MB (Figure 2(c)) on three
 185 electrodes, with samples dispensed sequentially in arbitrary order on each
 186 electrode ($20\text{ }\mu\text{M} \rightarrow 30\text{ }\mu\text{M} \rightarrow 40\text{ }\mu\text{M}$ on electrode 1, $40\text{ }\mu\text{M} \rightarrow 30\text{ }\mu\text{M} \rightarrow 20\text{ }\mu\text{M}$ on
 187 electrode 2, and $20\text{ }\mu\text{M} \rightarrow 40\text{ }\mu\text{M} \rightarrow 30\text{ }\mu\text{M}$ on electrode 3). Overall, these re-
 188 sults show that cleaning the electrodes with an IPA dampened wipe after
 189 measuring each sample is effective, based on the measurement repeatability
 190 seen on the three electrodes. Although the magnitude of the current was
 191 consistent across electrodes, it was found that the peak potential was not
 192 stable and varied across measurements (Figure 1 in Supplementary Informa-
 193 tion). The changes in the peak potential are likely to be caused by a number
 194 of issues, including the ratio of oxidized and reduced MB present, the subtle
 195 differences in dissolved species (solutes) between different samples, impuri-
 196 ties on the surface of the reference electrode, and the use of gold reference
 197 electrode on the PCB instead of a stable reference such as Ag/AgCl. The
 198 electrode-to-electrode variation indicates that one cannot rely on a universal
 199 calibration curve when using multiple electrodes, and each electrode must be
 200 calibrated separately before use.

201 *3.2. Longer DNA fragments result in peak current changes proportional to*
 202 *the concentration of DNA present*

203 DPV voltammograms were obtained on PCB electrode for various DNA
 204 concentrations ranging from $10\text{ pg}/\mu\text{l}$ to $200\text{ pg}/\mu\text{l}$ complexed with $10\text{ }\mu\text{M}$

205 MB. When the longer N1-N2 DNA amplicon (943 bp) was measured, we
 206 found that the peak DPV current increases with increasing DNA concen-
 207 tration, whereas the peak current for the shorter N1 fragment (72 bp) does
 208 not show any distinct trend (Figure 3(a)). To explore this further, DPV
 209 measurements were performed with 5 electrodes each for N1-N2 and N1 am-
 210 plicons at different concentrations (Figure 3(b)). Linear regression shows
 211 that the longer N1-N2 fragment has larger sensitivity (0.605 nA/pg/ μ l) and
 212 better linearity ($R^2 = 0.88$) than the shorter N1 fragment (0.27 nA/pg/ μ l,
 213 $R^2 = 0.19$), indicating that the length of the DNA fragment has significant
 214 impact on sensor performance.

215 These results reveal that the amplicon length has a profound impact upon
 216 the linearity and therefore utility of the measurements performed. We believe
 217 these are attributed to intercalation of MB into the DNA and adsorption
 218 of the DNA onto the electrode surface. Other investigators have previously
 219 shown that the size of the oligonucleotide has a large impact on the adsorption
 220 at solid electrodes [48, 49]. The dependence of DNA-MB intercalation on
 221 DNA length and sequence has been reported previously [25, 50–52]. The
 222 intercalation of MB into the DNA base stack results in increase in current
 223 with increasing DNA concentration, since the DNA serves as a scaffold to
 224 bring the redox-active MB molecules into direct contact with the electrode
 225 surface, but does not participate directly in the electron-transfer event [53,
 226 54]. The longer DNA fragments intercalate a greater quantity of MB and
 227 this in turn adsorbs to the unmodified gold electrode surface, resulting in
 228 a greater unit concentration of MB at the electrode surface compared to
 229 shorter DNA fragments.

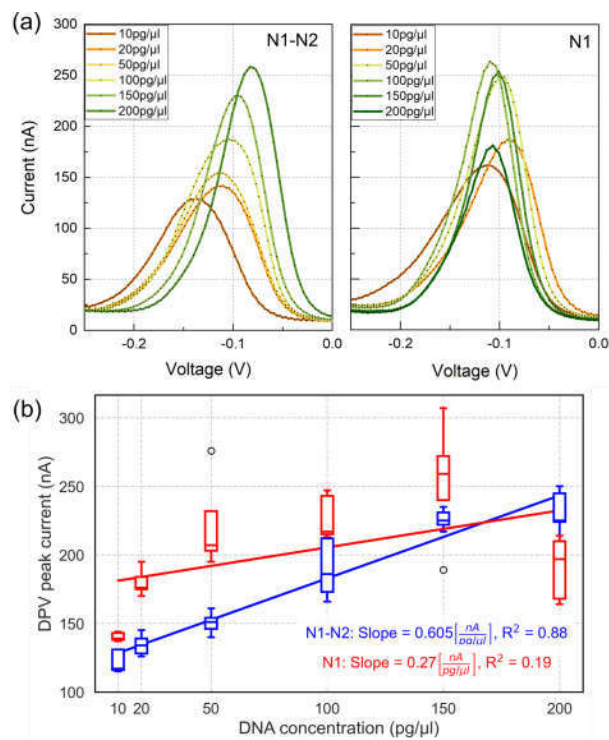


Figure 3: (a) DPV voltammograms obtained on PCB electrode for DNA-MB complex for various DNA concentrations complexed with 10 μ M MB. Left - 943 bp PCR product (N1-N2); Right - 72 bp PCR product. (b) Linear regression of DPV measurements obtained with 5 electrodes each for N1-N2 and N1 indicate more linear sensor response for the longer N1-N2 fragment.

230 3.3. CV measurements indicate adsorption of DNA-MB complex at the elec- 231 trode surface

232 Focusing upon the N1-N2 amplicon, we analyzed the electrochemical re-
233 sponse of different samples to establish the role of DNA adsorption to the
234 electrode surface. Firstly, we performed CV measurements at different scan
235 rates for 50 μ M MB (no DNA), and 100 pg/ μ l DNA (N1-N2) complexed with
236 50 μ M MB (DNA+MB) (Figure 4(a)). As described above, the cathodic peak
237 potential varies for both samples due to the use of gold reference electrode
238 on the PCB instead of a stable reference such as Ag/AgCl. The cathodic
239 and anodic peak currents increase with increasing scan rate as expected.
240 When the logarithm of the cathodic peak current (expressed in nA) is ex-
241 amined with respect to the logarithm of the scan rate (expressed in mV/s),
242 we observe a slope of approximately 1 (1.07 and 0.966 for MB and DNA-
243 MB complex respectively). The linear trend is also clearly established with
244 $R^2 \geq 0.98$ for MB and DNA-MB complex (Figure 4(b)). Electrochemically
245 reversible electron transfer involving freely diffusing redox species correspond
246 to $i_p \propto \nu^{0.5}$ (i.e. slope = 0.5 in $\log(i_p) - \log(\nu)$ plot), where i_p and ν denote
247 cathodic peak current and scan rate respectively. In contrast, a slope of 1
248 (i.e. $i_p \propto \nu$) indicates that the current response is due to electrode-adsorbed
249 species [55, 56], and thus points to adsorption of MB and DNA-MB complex
250 on the gold electrodes. Adsorption of DNA-MB complex on the electrode
251 leads to higher DPV current for higher DNA and MB concentrations.

252 3.4. Higher MB concentrations increase electrode sensitivity

253 To fully optimize the sensor performance, we explored the role MB con-
254 centration has upon the peak DPV current in response to changes in the

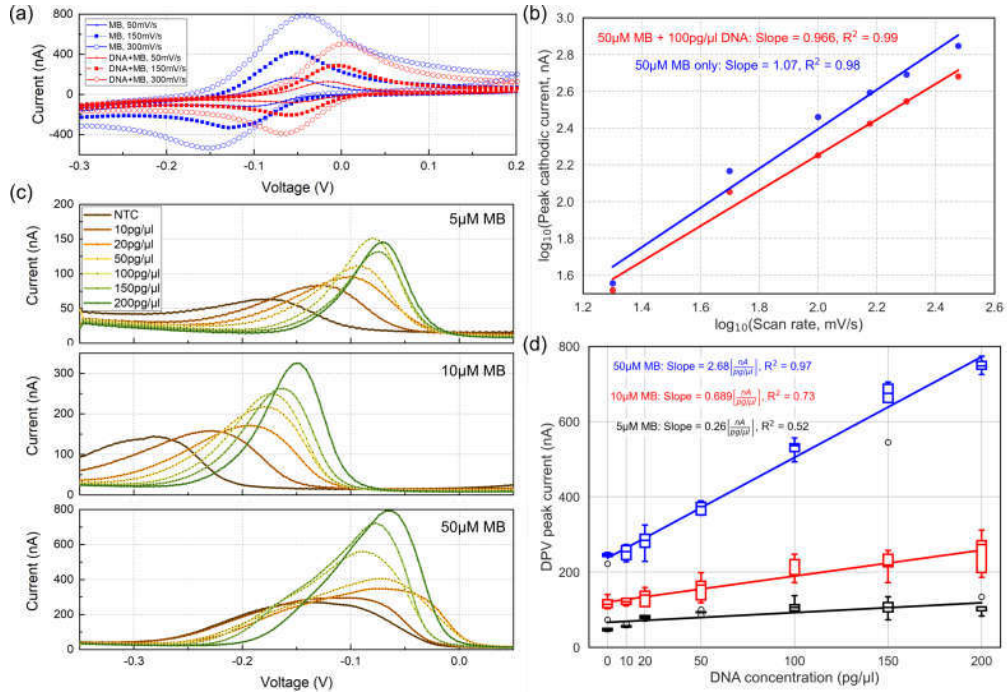


Figure 4: Electrochemical response of the sensor as a function of MB concentration. (a) Example CV measurements obtained at different scan rates for 50 μM MB only, and 100 $\text{pg}/\mu\text{l}$ DNA (N1-N2) complex with 50 μM MB (DNA+MB). (b) Peak cathodic current variation with scan rate for MB only and DNA-MB complex displayed in log-log plot. The logarithm of peak cathodic current shows linear trend of variation with logarithm of scan rate, with slope ≈ 1 , indicating that the current response is due to adsorption of MB and DNA-MB complex on the electrode surface. (c) DPV measurements obtained on PCB electrode for DNA(N1-N2)-MB complex for various concentrations of MB (5 μM , 10 μM , 50 μM) and DNA (ranging from 10 $\text{pg}/\mu\text{l}$ to 200 $\text{pg}/\mu\text{l}$). (d) Linear regression of DPV measurements obtained with 5 electrodes each for 5 μM , 10 μM and 50 μM MB indicate that the sensor response is more linear for higher concentration of MB.

255 concentration of the N1-N2 amplicon. We performed DPV measurements
 256 on DNA-MB complexes with different concentrations of MB (5 μ M, 10 μ M,
 257 50 μ M) and DNA. The DNA concentration was varied from 10 pg/ μ l (approx-
 258 imately 1.7 fM, equivalent to 10^{12} copies/ μ l) to 200 pg/ μ l (Figure 4(c)). We
 259 also performed linear regression using DPV peak currents obtained with 5
 260 electrodes each for 5 μ M, 10 μ M and 50 μ M MB intercalated with DNA (Fig-
 261 ure 4(d)). These results show that the sensitivity and linearity improve with
 262 increasing concentrations of MB (slope [nA/pg/ μ l] = 2.68, 0.689 and 0.26
 263 for 50 μ M, 10 μ M and 5 μ M MB respectively; R^2 = 0.97, 0.73 and 0.52 for
 264 50 μ M, 10 μ M and 5 μ M MB respectively). These results show that for our
 265 sensor design, the linearity and sensitivity are affected by the MB concentra-
 266 tion. This has important implications for future studies and suggests that
 267 wide range of MB concentrations should be explored during the optimization
 268 stage of sensor development.

269 *3.5. Long fragment PCR product is detectable after 10 cycles with PCB elec-* 270 *trodes*

271 Amplified products obtained using various template concentrations were
 272 visualized using agarose gel electrophoresis (Figure 5(a)). We found that
 273 the N1-N2 fragment (943 bp) is only visible for template concentrations of
 274 10^4 and 10^5 copies/ μ l. After optimizing the PCB electrodes with respect
 275 to DNA length and MB concentration, we then explored the performance
 276 of the electrodes during PCR cycles. PCR product was visualized using
 277 agarose gel electrophoresis after 5, 10, 20, 30 and 40 cycles (Figure 5(b)).
 278 For intermediate PCR amplicons obtained with 10^5 copies/ μ l template, the
 279 amplified product is visible in agarose gel only beyond 20 cycles of PCR. In

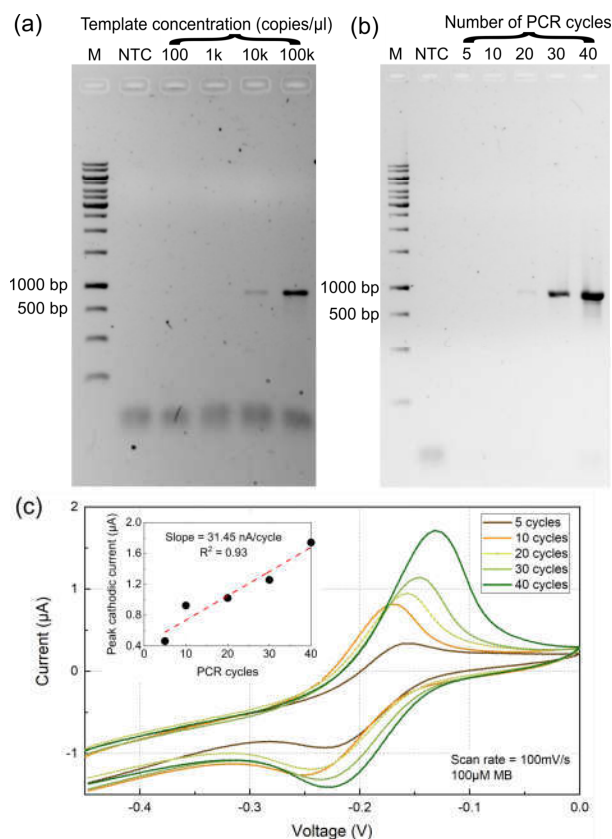


Figure 5: PCB electrode performance during intermediate PCR cycles. (a) Agarose gel showing the 2019-nCoV_N_Positive Control after 40 PCR cycles using the N1-N2 primers. The N1-N2 fragment is only visible with starting template concentrations of 10^4 and 10^5 copies/μl (M - Marker, NTC - Negative Control). (b) Agarose gel of PCR product following amplification of 10^5 copies template after fewer cycles. The amplified product is visible in agarose gel only beyond 20 cycles of PCR. (c) CV voltammograms obtained at 100 mV/s scan rate after adding 100 μM MB to the PCR products, showing that current response is proportional to the number of cycles.

contrast, the change in electrochemical sensor signal is clearly seen in the CV voltammogram performed at 100 mV/s scan rate after adding 100 μ M MB to the PCR products (Figure 5(c)). Furthermore these results show a linear trend for variation in peak cathodic current with number of PCR cycles. The PCB electrodes are thus capable of discerning intermediate PCR products for as low as 10 PCR cycles (i.e. approximately 10^8 copies/ μ l), that are not visible in the agarose gel.

3.6. SARS-CoV-2 RNA can be isolated from wastewater and electrochemically detected using a low cost PCR machine and PCB electrode

We explored the utility of the system developed here with wastewater samples spiked with ATCCTM SARS-CoV-2 genomic RNA. While longer fragment amplicons appear to be better for MB intercalation and electrochemical detection, amplification efficiency of long fragments of SARS-CoV-2 RNA isolated from heterogeneous samples such as wastewater is challenging because RNA is prone to shearing and degradation during isolation. Importantly, we found that products longer than 500 bp could not be amplified efficiently. Considering these practical challenges associated with RNA isolation from heterogeneous samples, PCR amplification and electrochemical sensing, it emerges that a multi-parameter optimization study is required to engineer a robust testing protocol. Though attempts were made to amplify the 943 bp N1-N2 fragment, we opted to focus upon the shorter 72 bp N1 region as this can be amplified more efficiently from wastewater (Figure 6(a)). It should be noted that when we work with heterogeneous environmental samples such as wastewater, it is possible that there may be non-specific amplification [57]. The band at approximately 20 bp in lanes NTC (no template

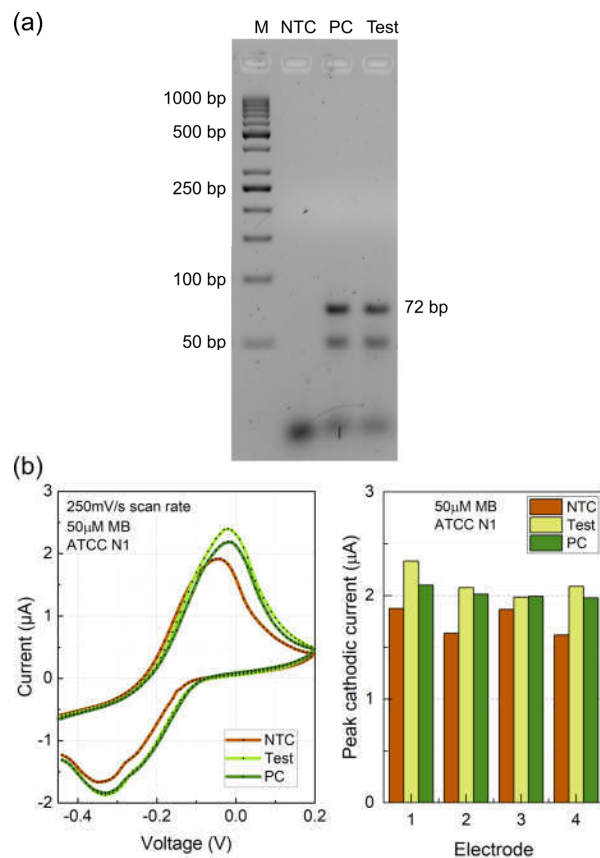


Figure 6: Detection of SARS-CoV-2 RNA spiked into wastewater using PCB electrodes. (a) Agarose gel showing PCR amplicons from a simulated wastewater sample containing SARS-CoV-2 RNA (M - Marker, NTC - Negative Control, PC - Positive Control, Test - Test Sample, SARS-CoV-2 RNA negative wastewater spiked with ATCC™ SARS-CoV-2 RNA as detailed in section 2.1.3). In addition to the expected band at 72 bp, we also found a band at approximately 50 bp in the positive control and sample. (b) (Left) Example CV voltammograms obtained on electrode 1, with 50 μM MB added to RT-PCR products. (Right) CV cathodic peak currents measured on 4 separate PCB electrodes. While there is large electrode-to-electrode variation, the PCB electrodes produce larger peak currents for test sample and PC than NTC.

control), PC (positive control), and Test corresponds to excess primers. After confirmation of PCR amplicons, we performed CV measurements using the PCB electrodes and found that a change in the peak current could be observed when the amplicons were present in the sample (Figure 6(b)). The negative control (NTC) also induces electrochemical sensor response due to adsorption of MB on the electrode, and therefore each electrode must be separately calibrated with NTC and PC samples in order to distinguish false positives from true positives. These results suggest that, with further development, the electrode system could be used in the field for determination of SARS-CoV-2 amplicons.

4. Conclusion

In this paper, we characterize a low-cost detection system for the identification of SARS-CoV-2 nucleic acid in wastewater. Our approach takes advantage of existing primer sets, and uses a commercially available low-cost thermocycler to reverse transcribe and amplify key regions. Using this sensor system, we report that it is possible to identify the presence of SARS-CoV-2 in wastewater. We showed that the peak DPV current is proportional to the DNA concentration due to the adsorption of DNA-MB complex on the electrode. We also show that the electrochemical sensitivity is strongly determined by concentration of MB and the length of the DNA fragment. The platform presented in this work is a cost-effective electrochemical sensing solution, with manufacturing costing as low as USD \$0.55 (INR ₹40) per electrode for quantities of 100. Unlike real-time PCR (qPCR) techniques, this platform does not require expensive reagents or instrumentation, thus

329 further boosting the cost-effectiveness of this solution. The electrodes are also
330 easily cleaned using a lint-free wipe dampened with IPA and are reusable.
331 Furthermore, the electrodes do not undergo any surface modification and
332 thus have long shelf-life. Our future work will focus upon identifying optimal
333 primers for both PCR amplification and electrochemical sensing, integrating
334 this assay with on-board thermocycling and electrochemical measurement,
335 and exploring methods for enhanced stability of the reference electrode to
336 achieve redox peaks at consistent potentials.

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376 Declaration of interest

377 The authors declare that they have no known competing financial inter-
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380 References

- 381 [1] E. Janik, M. Bartos, M. Niemcewicz, L. Gorniak, M. Bijak, SARS-CoV-
382 2: Outline, Prevention, and Decontamination, *Pathogens* 10 (2021) 114.
- 383 [2] M. Mofijur, I. R. Fattah, M. A. Alam, A. S. Islam, H. C. Ong, S. A.
384 Rahman, G. Najafi, S. Ahmed, M. A. Uddin, T. Mahlia, Impact of
385 covid-19 on the social, economic, environmental and energy domains:
386 Lessons learnt from a global pandemic, *Sustainable production and*
387 *consumption* 26 (2021) 343–359.
- 388 [3] I. L. Organization, ILO Monitor: Covid-19 and the world of work, 2020.
- 389 [4] S. Alpdagtas, E. Ilhan, E. Uysal, M. Sengor, C. B. Ustundag, O. Gun-
390 duz, Evaluation of current diagnostic methods for COVID-19, *APL*
391 *bioengineering* 4 (2020) 041506.
- 392 [5] J. Watson, P. F. Whiting, J. E. Brush, Interpreting a covid-19 test
393 result, *Bmj* 369 (2020).
- 394 [6] W. Lodder, A. M. de Roda Husman, SARS-CoV-2 in wastewater: po-
395 tential health risk, but also data source, *The lancet Gastroenterology &*
396 *hepatology* 5 (2020) 533–534.

- 397 [7] D. A. Larsen, K. R. Wigginton, Tracking COVID-19 with wastewater,
398 Nature Biotechnology 38 (2020) 1151–1153.
- 399 [8] D. Barceló, Wastewater-Based Epidemiology to monitor COVID-19 out-
400 break: Present and future diagnostic methods to be in your radar, Case
401 Studies in Chemical and Environmental Engineering 2 (2020) 100042.
- 402 [9] P. Foladori, F. Cutrupi, N. Segata, S. Manara, F. Pinto, F. Malpei,
403 L. Bruni, G. La Rosa, SARS-CoV-2 from faeces to wastewater treat-
404 ment: what do we know? A review, Science of the Total Environment
405 743 (2020) 140444.
- 406 [10] D. Polo, M. Quintela-Baluja, A. Corbishley, D. L. Jones, A. C. Singer,
407 D. W. Graham, J. L. Romalde, Making waves: Wastewater-based epi-
408 demiology for COVID-19—approaches and challenges for surveillance and
409 prediction, Water Research 186 (2020) 116404.
- 410 [11] T. Zuo, Q. Liu, F. Zhang, G. C.-Y. Lui, E. Y. Tso, Y. K. Yeoh, Z. Chen,
411 S. S. Boon, F. K. Chan, P. K. Chan, et al., Depicting SARS-CoV-2 faecal
412 viral activity in association with gut microbiota composition in patients
413 with COVID-19, Gut 70 (2021) 276–284.
- 414 [12] W. Lin, Z. Xie, Y. Li, L. Li, C. Wen, Y. Cao, X. Chen, X. Ou, F. Hu,
415 F. Li, et al., Association between detectable SARS-COV-2 RNA in anal
416 swabs and disease severity in patients with coronavirus disease 2019,
417 Journal of medical virology 93 (2021) 794–802.
- 418 [13] J. Liu, Y. Xiao, Y. Shen, C. Shi, Y. Chen, P. Shi, Y. Gao, Y. Wang,
419 B. Lu, Detection of SARS-CoV-2 by RT-PCR in anal from patients

420 who have recovered from coronavirus disease 2019, *Journal of medical*
421 *virology* 92 (2020) 1769–1771.

422 [14] Y. Tian, L. Rong, W. Nian, Y. He, Gastrointestinal features in COVID-
423 19 and the possibility of faecal transmission, *Alimentary pharmacology*
424 *& therapeutics* 51 (2020) 843–851.

425 [15] S. P. Sherchan, S. Shahin, L. M. Ward, S. Tandukar, T. G. Aw,
426 B. Schmitz, W. Ahmed, M. Kitajima, First detection of SARS-CoV-
427 2 RNA in wastewater in North America: a study in Louisiana, USA,
428 *Science of The Total Environment* 743 (2020) 140621.

429 [16] E. Haramoto, B. Malla, O. Thakali, M. Kitajima, First environmental
430 surveillance for the presence of SARS-CoV-2 RNA in wastewater and
431 river water in Japan, *Science of The Total Environment* 737 (2020)
432 140405.

433 [17] S. G. Rimoldi, F. Stefani, A. Gigantiello, S. Polesello, F. Comandatore,
434 D. Mileto, M. Maresca, C. Longobardi, A. Mancon, F. Romeri, et al.,
435 Presence and infectivity of SARS-CoV-2 virus in wastewaters and rivers,
436 *Science of the Total Environment* 744 (2020) 140911.

437 [18] M. Kumar, A. K. Patel, A. V. Shah, J. Raval, N. Rajpara, M. Joshi,
438 C. G. Joshi, First proof of the capability of wastewater surveillance for
439 COVID-19 in India through detection of genetic material of SARS-CoV-
440 2, *Science of The Total Environment* 746 (2020) 141326.

441 [19] E. Palecek, M. Bartosik, *Electrochemistry of nucleic acids*, *Chemical*
442 *Reviews* 112 (2012) 3427–3481.

- 443 [20] A. Tani, A. J. Thomson, J. N. Butt, Methylene blue as an electrochem-
444 ical discriminator of single-and double-stranded oligonucleotides immo-
445 bilised on gold substrates, *Analyst* 126 (2001) 1756–1759.
- 446 [21] E. L. Wong, P. Erohkin, J. J. Gooding, A comparison of cationic and
447 anionic intercalators for the electrochemical transduction of DNA hy-
448 bridization via long range electron transfer, *Electrochemistry commu-
449 nications* 6 (2004) 648–654.
- 450 [22] E. L. Wong, J. J. Gooding, Charge transfer through DNA: a selective
451 electrochemical DNA biosensor, *Analytical chemistry* 78 (2006) 2138–
452 2144.
- 453 [23] T. H. Fang, N. Ramalingam, D. Xian-Dui, T. S. Ngin, Z. Xianting,
454 A. T. L. Kuan, E. Y. P. Huat, G. Hai-Qing, Real-time pcr microflu-
455 idic devices with concurrent electrochemical detection, *Biosensors and
456 Bioelectronics* 24 (2009) 2131–2136.
- 457 [24] B. Y. Won, S. Shin, S. Baek, Y. L. Jung, T. Li, S. C. Shin, D.-Y.
458 Cho, S. B. Lee, H. G. Park, Investigation of the signaling mechanism
459 and verification of the performance of an electrochemical real-time PCR
460 system based on the interaction of methylene blue with DNA, *Analyst*
461 136 (2011) 1573–1579.
- 462 [25] E. Farjami, L. Clima, K. V. Gothelf, E. E. Ferapontova, DNA interac-
463 tions with a methylene blue redox indicator depend on the DNA length
464 and are sequence specific, *Analyst* 135 (2010) 1443–1448.

- [26] H.-Y. Tseng, V. Adamik, J. Parsons, S.-S. Lan, S. Malfesi, J. Lum, L. Shannon, B. Gray, Development of an electrochemical biosensor array for quantitative polymerase chain reaction utilizing three-metal printed circuit board technology, *Sensors and Actuators B: Chemical* 204 (2014) 459–466.
- [27] L. Lu, J. C. Si, Z. F. Gao, Y. Zhang, J. L. Lei, H. Q. Luo, N. B. Li, Highly selective and sensitive electrochemical biosensor for ATP based on the dual strategy integrating the cofactor-dependent enzymatic ligation reaction with self-cleaving DNzyme-amplified electrochemical detection, *Biosensors and Bioelectronics* 63 (2015) 14–20.
- [28] S. Cinti, E. Proietti, F. Casotto, D. Moscone, F. Arduini, Paper-based strips for the electrochemical detection of single and double stranded DNA, *Analytical chemistry* 90 (2018) 13680–13686.
- [29] T. Deféver, M. Druet, D. Evrard, D. Marchal, B. Limoges, Real-time electrochemical PCR with a DNA intercalating redox probe, *Analytical chemistry* 83 (2011) 1815–1821.
- [30] F.-T. Zhang, J. Nie, D.-W. Zhang, J.-T. Chen, Y.-L. Zhou, X.-X. Zhang, Methylene blue as a G-quadruplex binding probe for label-free homogeneous electrochemical biosensing, *Analytical chemistry* 86 (2014) 9489–9495.
- [31] F.-T. Zhang, L.-Y. Cai, Y.-L. Zhou, X.-X. Zhang, Immobilization-free DNA-based homogeneous electrochemical biosensors, *TrAC Trends in Analytical Chemistry* 85 (2016) 17–32.

- 488 [32] A. Martin, K. B. Grant, F. Stressmann, J.-M. Ghigo, D. Marchal,
489 B. Limoges, Ultimate single-copy DNA detection using real-time elec-
490 trochemical LAMP, *Acs Sensors* 1 (2016) 904–912.
- 491 [33] E. Nunez-Bajo, A. S. P. Collins, M. Kasimatis, Y. Cotur, T. Asfour,
492 U. Tanriverdi, M. Grell, M. Kaisti, G. Senesi, K. Stevenson, et al., Dis-
493 posable silicon-based all-in-one micro-qPCR for rapid on-site detection
494 of pathogens, *Nature communications* 11 (2020) 1–10.
- 495 [34] T. M. Herne, M. J. Tarlov, Characterization of DNA probes immobilized
496 on gold surfaces, *Journal of the American Chemical Society* 119 (1997)
497 8916–8920.
- 498 [35] D. Moschou, A. Tserepi, The lab-on-PCB approach: tackling the μ TAS
499 commercial upscaling bottleneck, *Lab on a Chip* 17 (2017) 1388–1405.
- 500 [36] B. B. Narakathu, S. G. R. Avuthu, A. Eshkeiti, S. Emamian, M. Z.
501 Atashbar, Development of a microfluidic sensing platform by integrating
502 PCB technology and inkjet printing process, *IEEE Sensors Journal* 15
503 (2015) 6374–6380.
- 504 [37] F. Güth, P. Arki, T. Löher, A. Ostmann, Y. Joseph, Electrochemical
505 sensors based on printed circuit board technologies, *Procedia Engineer-*
506 *ing* 168 (2016) 452–455.
- 507 [38] F. T. Moreira, M. J. M. Ferreira, J. R. Puga, M. G. F. Sales, Screen-
508 printed electrode produced by printed-circuit board technology: Appli-
509 cation to cancer biomarker detection by means of plastic antibody as

- sensing material, *Sensors and Actuators B: Chemical* 223 (2016) 927–935.
- [39] D. Evans, K. I. Papadimitriou, N. Vasilakis, P. Pantelidis, P. Kelleher, H. Morgan, T. Prodromakis, A novel microfluidic point-of-care biosensor system on printed circuit board for cytokine detection, *Sensors* 18 (2018) 4011.
- [40] K. Tsougeni, A. Kastania, G. Kaprou, M. Eck, G. Jobst, P. Petrou, S. Kakabakos, D. Mastellos, E. Gogolides, A. Tserepi, A modular integrated lab-on-a-chip platform for fast and highly efficient sample preparation for foodborne pathogen screening, *Sensors and Actuators B: Chemical* 288 (2019) 171–179.
- [41] H. Shamkhalichenar, C. J. Bueche, J.-W. Choi, Printed Circuit Board (PCB) Technology for Electrochemical Sensors and Sensing Platforms, *Biosensors* 10 (2020) 159.
- [42] H. Zhu, Z. Föhlerová, J. Pekárek, E. Basova, P. Neužil, Recent advances in lab-on-a-chip technologies for viral diagnosis, *Biosensors and Bioelectronics* 153 (2020) 112041.
- [43] G. Dutta, A. Regoutz, D. Moschou, Commercially fabricated printed circuit board sensing electrodes for biomarker electrochemical detection: The importance of electrode surface characteristics in sensor performance, in: *Multidisciplinary Digital Publishing Institute Proceedings*, volume 2, 2018, p. 741.

- [44] D. Moschou, T. Trantidou, A. Regoutz, D. Carta, H. Morgan, T. Prodromakis, Surface and electrical characterization of Ag/AgCl pseudo-reference electrodes manufactured with commercially available PCB technologies, *Sensors* 15 (2015) 18102–18113.
- [45] N. Bhalla, Y. Pan, Z. Yang, A. F. Payam, Opportunities and challenges for biosensors and nanoscale analytical tools for pandemics: COVID-19, *ACS Nano* 14 (2020) 7783–7807.
- [46] W. Randazzo, P. Truchado, E. Cuevas-Ferrando, P. Simón, A. Allende, G. Sánchez, SARS-CoV-2 RNA in wastewater anticipated COVID-19 occurrence in a low prevalence area, *Water research* 181 (2020) 115942.
- [47] L. M. Fischer, M. Tenje, A. R. Heiskanen, N. Masuda, J. Castillo, A. Bentien, J. Émneus, M. H. Jakobsen, A. Boisen, Gold cleaning methods for electrochemical detection applications, *Microelectronic engineering* 86 (2009) 1282–1285.
- [48] A. Steel, R. Levicky, T. Herne, M. J. Tarlov, Immobilization of nucleic acids at solid surfaces: effect of oligonucleotide length on layer assembly, *Biophysical journal* 79 (2000) 975–981.
- [49] M. Pedano, G. Rivas, Immobilization of dna on glassy carbon electrodes for the development of affinity biosensors, *Biosensors and Bioelectronics* 18 (2003) 269–277.
- [50] R. García-González, A. Costa-García, M. T. Fernández-Abedul, Methylene blue covalently attached to single stranded DNA as electroactive

- 554 label for potential bioassays, *Sensors and Actuators B: Chemical* 191
555 (2014) 784–790.
- 556 [51] C. Li, X. Chen, N. Wang, B. Zhang, An ultrasensitive and label-free
557 electrochemical DNA biosensor for detection of DNase I activity, *RSC*
558 *advances* 7 (2017) 21666–21670.
- 559 [52] W. Yang, M. Ozsoz, D. B. Hibbert, J. J. Gooding, Evidence for the di-
560 rect interaction between methylene blue and guanine bases using DNA-
561 modified carbon paste electrodes, *Electroanalysis: An International*
562 *Journal Devoted to Fundamental and Practical Aspects of Electroanal-*
563 *ysis* 14 (2002) 1299–1302.
- 564 [53] C. G. Pheeney, J. K. Barton, DNA electrochemistry with tethered
565 methylene blue, *Langmuir* 28 (2012) 7063–7070.
- 566 [54] A. L. Furst, M. G. Hill, J. K. Barton, Electrocatalysis in DNA sensors,
567 *Polyhedron* 84 (2014) 150–159.
- 568 [55] N. Elgrishi, K. J. Rountree, B. D. McCarthy, E. S. Rountree, T. T.
569 Eisenhart, J. L. Dempsey, A practical beginner’s guide to cyclic voltam-
570 metry, *Journal of chemical education* 95 (2018) 197–206.
- 571 [56] A. J. Bard, L. R. Faulkner, *Electrochemical methods: Fundamentals*
572 *and applications*, Wiley New York, 2001.
- 573 [57] K. Megha, S. B. Lad, K. Kondabagil, Detection of SARS-CoV-2 in
574 wastewater: Less than what meets the eye?, *Communicated* (2021).