

Integrated environmental vulnerability to oil spills in sensitive areas

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Abstract

As the typical range of influence of oil spills surrounds urbanised and economically active areas, it is likely that fragile regions may not be part of the most vulnerable zones. This premise is remediated in this paper with the adoption of a vulnerability approach based on the integration of static and dynamic information, such as oil pollution susceptibility. Susceptibility is a poorly consolidated term and is often used as synonym for environmental sensitivity; it is considered here to be the distribution areas of oil slicks. To test the proposed approach, an integrated estimation of environmental vulnerability is carried out for an environmentally sensitive area in the south of Brazil by merging static data inherent to the medium with information of a dynamic nature related to trajectory, behaviour and the fate of oil at sea. Moreover, the oil pollution intensity and environmental sensitivity data in susceptible areas are addressed. Subsequently, the environmental vulnerability is estimated by integrating hazard maps, concentrations and losses of the mass of the oil slick, oil beaching time and the littoral sensitivity index hierarchy. Results will prove to be useful to highlight critical areas for which the highest levels of severity are expected, which can lead to improvements in decision-making processes to support oil-spill prevention, as well as improve response readiness, especially in developing countries that have historically under-protected their sensitive regions.

Keywords: Oil spill modelling, susceptibility, environmentally sensitive, hazard, vulnerability,

1 Main finding

2 This study allowed us to estimate oil spill vulnerability in environmentally sensitive areas, which are the
3 subject of high global concern due to environmental losses and other negative impacts.

4 1. Introduction

5 A vulnerability analysis of marine oil spills remains useful and important, mainly due to the fact that
6 oil exploration and extraction continue to be large contributors in the global market (Nelson and Grubestic,
7 2018). According to classic approaches, oil spill vulnerability represents a measure of the degree of losses for
8 the environmental elements in the face of the consequences of oil spills (Michel et al., 1978; Gundlach and
9 Hayes, 1978; Hanna, 1995; Douglas, 2007; Williams et al., 2008). This information is usually obtained from oil
10 sensitivity mappings that are represented as static variables (Nelson et al., 2015; Nelson and Grubestic, 2018;
11 Amir-Heidari et al., 2019). On the other hand, physical vulnerability is defined as a dynamic variable that
12 results from the integration of intrinsic environment data from other sources (Li et al., 2012; Uzielli et al.,
13 2008). Information such as the fractions of oil evaporated and oil trapped on the coast can also be considered
14 in estimations and classifications of vulnerable areas (Alves et al., 2015). Zones under oil spill risk may have

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15 greater or lesser vulnerability, despite the environmental levels of sensitivities among other conditioning
16 factors (Alves et al., 2014, 2015; Nelson et al., 2015; Goldman et al., 2015; Guo et al., 2019). For example,
17 a high oil-sensitive area, such as a salt marsh, can be less vulnerable than fine sand beaches, considering
18 that the exposure levels of an oil spill are beyond coastal sensitivity levels. This is because oil sensitivity is
19 inherent to the environment, which may be less or more sensitive depending on its own characteristics, while
20 in contrast, physical vulnerability depends on external factors, such as oil covering levels and concentrations
21 and oil beaching time in sensitive areas. Another concept that is sometimes mistakenly used as a synonym for
22 environmental vulnerability is environmental sensitivity. Whereas environmental vulnerability derives from
23 the susceptibility to an event of negative consequences (for Standardization , ISO) and could also be assumed
24 to be a metric of the probability of exposure to external factors (Zacharias and Gregr, 2005), environmental
25 sensitivity is assumed to be an intrinsic property of exposed elements (Guo et al., 2019).

26 Thus, vulnerability mapping comprehends an interdisciplinary evaluation and takes information from
27 distinct scientific domains into account (Alves et al., 2016a,b; Shami et al., 2017). These maps are extremely
28 important to decision-makers when planning for marine protected areas, placing new infrastructure or
29 enacting protocols on handling marine pollution (Alves et al., 2016b; Goldman et al., 2015). The mapping
30 of environmental vulnerability associated with marine oil spills has generally been supported more by
31 information inherent to the environment than by the results of modelling the behaviour of oil in water. This,
32 in turn, may lead to an imprecise evaluation of the extent of damage caused by accidents. Therefore, this
33 study proposes an integrated assessment of environmental vulnerability based on a combination of static
34 and inherent variables in the environment, and those related to the behaviour and dynamics of oil slicks.
35 This work is aimed at estimating the possible reach of oil spills by indicating the most vulnerable areas,
36 and consequently to monitoring expected damage in regions of the highest pollution probabilities. Figure 1
37 presents the main concepts covered by this paper and how they relate to the proposed approach for this
38 integrated estimation of environmental vulnerability.

39 As illustrated Figure 1, vulnerability is understood here to be the expected degree of loss of exposed
40 elements in the face of a potentially harmful event and disaster reduction (UN/ISDR, 2004). This concept not
41 only essentially depends on the resistance of the exposed elements, but also on the potential of an accident
42 (Salter, 1997; Kleissen et al., 2007). Thus, as highlighted in previous research (e.g., Goldman et al. (2015,
43 2017); Nelson et al. (2015); Nelson and Grubescic (2018)), the estimate of the environmental vulnerability to
44 oil spills merges information of different natures, such as probability, intensity or severity of a spill, among
45 other inputs. These considerations allow the determination of the scope of the effects of an oil spill, as well
46 as a notion of the extension of the damage (Afenyo et al., 2016; Xu et al., 2019).

47 Following a well-accepted concept in the literature, a hazard is the probability of occurrence of a
48 potentially damaging phenomenon in a given area and in a given period, and the impact of that event on
49 the environment or society (Larson, 2003; Varnes, 1984). In this study, the hazard was measured as the
50 oil-covering probabilities due to the movement and behaviour of hypothetical oil slicks in alertness areas.
51 In case of accidental marine oil spills, shorelines with high hazard and sensitivity should be prioritised, as
52 they are critical areas for pollution (Alves et al., 2014, 2015). An estimation of oil-susceptible areas can be
53 employed to delimit hazard regions by indicating areas of spreading and diffusion of oil spills.

54 According to this paper's approach, a coast with absolutely no structures for handling fuels or oil tankers
55 could have its integrated environmental vulnerability drastically reduced, even though the susceptibility,
56 threat and intensity of oil spills are only considered from a theoretical point of view. Such an approach is
57 useful in the context of many developing countries like Brazil, where the fact that a coastal zone is sensitive
58 does not ensure that it is protected; to the contrary, many of these areas are the targets of significant damage
59 caused by oil spills (Chun et al., 2020; Guo et al., 2019; Prasad et al., 2019; Oriaku et al., 2017; Salim et al.,
60 2015; Na et al., 2012; Orta-Martínez and Finer, 2010). The case of the oil spill that recently hit several
61 sectors of the Brazilian coast is a prime example of these weaknesses. Oil slicks were first identified on the
62 Brazilian coast on August 30, 2019; they reached about 4,334 km of the coast and covered areas in 11 states
63 of the Northeast and Southeast. The disaster is considered the worst marine oil spill in the history of Brazil
64 and one of the largest ever recorded in the world (de Araújo et al., 2020; Pena et al., 2020).

65 Thus, the proposed integrated vulnerability approach allows representation of different levels of potential
66 harm caused by the interaction between oil and the environment. Furthermore, it can improve the decision-

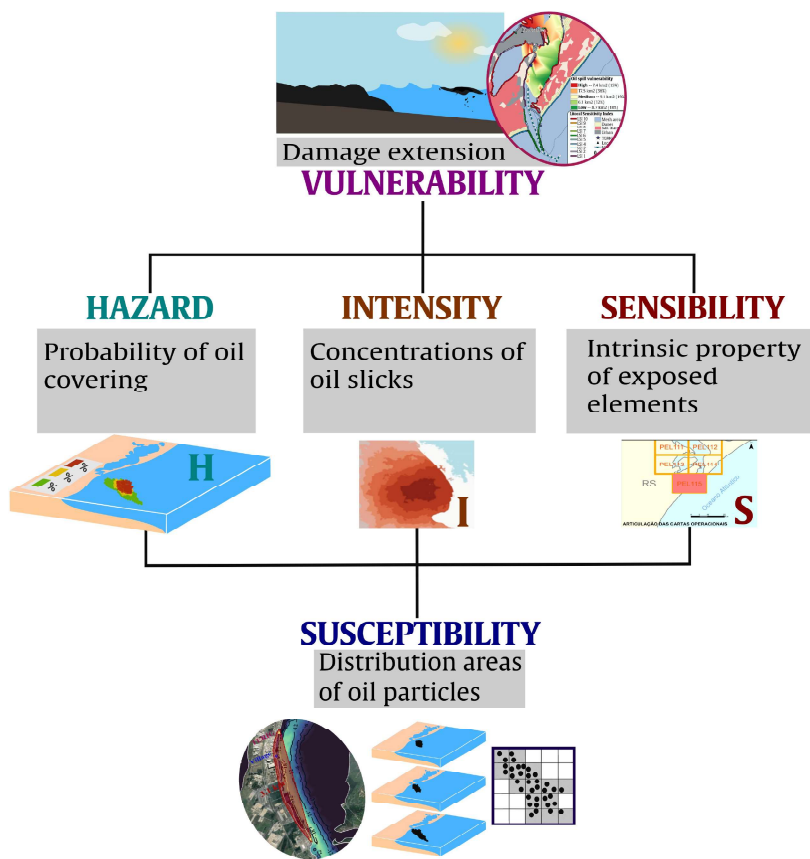


Figure 1: Proposed methodology involving the main extension of damages concepts to estimate integrated environmental vulnerability to oil spills.

67 making process for operators, responders and stakeholders to support oil-spill prevention and response
68 readiness (Nelson et al., 2015). Additionally, this vulnerability estimating is useful to assess a range of
69 conditions and scenarios and to provide a better understanding of the potential risks related to marine oil
70 spills.

71 2. Materials and Methods

72 Numerical simulations of oil spill events were performed in order to estimate the probability of the studied
73 areas to be polluted by oil in case of accidental spills. A total of 17.382 different scenarios were simulated,
74 each characterised by the spill event's proximity to different possible pollution sources, such as shipping
75 routes, pipelines and refinery and petrochemical industries. The oil covering probabilities were estimated
76 using numerical modelling data. The oil spill simulations were carried out by coupling two numerical models,
77 TELEMAC-3D and Easy Coupling Oil System (ECOS).

78 The methodological steps of this section can be summarised as follows:

- 79 I. Bathymetric, shoreline, oceanographic, meteorological and river discharge data were processed and
80 interpolated to the computational grid, since they are key for TELEMAC-3D to generate the hydrodynamic
81 results used in the oil-spill model.
- 82 II. The movement and behaviour of the oil spill and the susceptible areas were obtained from the ECOS
83 model. For the weathering, it was used the time series of evaporation and emulsion mass in the first
84 48 h after the accidents, as well as for mass loss and oil beaching time data.
- 85 III. Oil-spill-susceptible areas were analysed and the probability of pollution was mapped using *TerraView*¹.
- 86 IV. The modelled data, the oil spill probabilities and the environmental oil sensibility data and spill
87 behaviour were processed and integrated into a GIS environment using QGIS², and the Analytic
88 Hierarchy Process (AHP) devised by Saaty (1990) was used to finally obtain the vulnerability areas.

89 A more thorough description of the dataset, computational grid independence tests, calibration of
90 hydrodynamic parameters and validation of the simulation setup were previously performed and presented
91 by (Monteiro et al., 2019a,b).

92 2.1. Study area

93 The Southern Brazilian Shelf (SBS, Fig. 2a) was selected as the study area due to historical records of
94 accidental oil spills in its estuarine and coastal environments. Extending from latitude 34° S to 22° S, it is a
95 wide continental shelf with intense coastal currents and large mesoscale variability generated by bathymetric
96 and meteorological factors, such as local winds, micro-tidal behaviour and the constant influence of weather
97 fronts, which are stronger during the winter (Stech and Lorenzetti, 1992; Möller Jr et al., 2001; Dereczynski
98 and Menezes, 2015). In terms of oil sensitivity, which is represented by the Littoral Sensitivity Index (LSI),
99 this region presents rocky shores, which are classified as features of low coastal sensitivity to oil (LSI 1 and 2),
100 sandy beaches (LSI 4 to 6) and sandy or muddy plains with vegetated margins classified according to specific
101 characteristics (LSI of 8–10) (Marinho¹ and Nicolodi, 2019; Nicolodi, 2016; MMA, 2007). Most northern
102 areas present a medium sensitivity to the influence of oil, except for some high- LSI areas around coastal
103 lagoons. In the south, LSI values are more diverse, and more severe oil sensitivity levels can be found.

104 In the southern zone of the study area, the Patos Lagoon estuary is an economically important area for
105 the southern Brazilian states due to maritime, industrial and tourism activities (Fig. 2b). This area is also
106 threatened by pollution effects from the Rio Grande port and industrial zones, because of the activities of
107 the oil refinery and the port terminal, fertiliser-production plants, fisheries and urban places in Rio Grande
108 (Mirlean et al., 2003a,b; Lopes et al., 2019; Leal et al., 2019). At the northern region, the vicinities of

¹<http://www.obt.inpe.br/OBT/assuntos/projetos/terralib-terraview>

²<https://www.qgis.org>

109 Tramandaí and Imbé are subjected to accidental spills at the oil collector buoy MN-602 (Fig. 2c), which
110 connects the Almirante Soares Dutra Oceanic Terminal (TEDUT) to the Alberto Pasqualini Refinery (Refap)
111 in Porto Alegre (Stringari, 2013; Monteiro et al., 2017; Marques et al., 2017). This locale poses an additional
112 oil response challenge because oil spills can occur anywhere along the vessel routes, which makes it difficult
113 to monitor the impact thereof.

114 2.2. Coupling between ECOS and TELEMAC-3D

115 The TELEMAC-3D hydrodynamic model was used to solve the hydrostatic Navier–Stokes equations under
116 Boussinesq approximation, along with transport equations for scalar quantities of salinity and temperature.
117 The finite element method was employed to solve the hydrodynamic equations, using sigma coordinate system
118 for the vertical discretisation to represent both sea surface height and bed (Hervouet, 2007). Results from
119 the three-dimensional hydrodynamic model, such as currents, salinity and temperature fields, were used by
120 the ECOS oil-spill model to calculate the weathering processes and the transport of oil. ECOS represents an
121 oil slick as a set of discrete particles, the displacement and properties of which are updated at each time
122 step and changes in oil properties and hydrodynamic conditions are considered (Mello et al., 2011; Stringari,
123 2013; Marques et al., 2017). ECOS computes the oil slick evaporation using the algorithms of Stiver and
124 Mackay (1984), Fingas (1995, 1999), with empirical coefficients obtained from Lehr et al. (2002). For the
125 emulsification process, ECOS uses a first-order differential equation from Mackay et al. (1980) to consider
126 the maximum emulsification for a light type of oil and other empirical constants according to Mackay et al.
127 (1982). Mass changes in the oil slicks are computed based on emulsification and evaporation. Previous studies
128 have verified that ECOS’s results are in agreement with technical reports that monitor real oil spill incidents
129 in the same region (Monteiro et al., 2019b) and with similar studies using the same setup (Marques et al.,
130 2017; Monteiro et al., 2017).

131 2.3. Oil spill scenarios

132 To examine the potential spatial distribution of marine oil spills on the Southern Brazilian Shelf (SBS), the
133 probability of the oil spill pollution was estimated by assuming three points for the start of the hypothetical
134 oil spills modelled for a total of 17,382 daily events over nine years of simulations (2007–2015). **The initial
135 volume (20 m³) of spills was chosen based on cases that have already occurred in the region. Each spill event
136 was simulated separately as a scenario in which the oil load is instantly released from a certain point and
137 then drifts for 48 h; results were recorded every 30 min.** For the initialisation of the oil spills, two points
138 were selected in the south (Fig. 2b): one was located in the Rio Grande Port manoeuvring zone (ZP-19—
139 32°14'2" S and 51°58'5" W), and the other was at Transpetro’s Waterway Terminal (TERIG—32°3'21" S
140 and 52°3'12" W) in the estuary. In the north zone, the coordinates of the MN-602 buoy (30°1'36" S and
141 50°5'12" W) were used (Fig. 2c) as the initialisation point. After sensitivity simulations that used different
142 values for oil specific gravity in the study area (Monteiro et al., 2019a), it was decided not to vary the oil
143 specific mass values, and to use a single API gravity of 33.8 (856 kg/m³), which is the typical diesel fuel that
144 is frequently handled and transported in the study region.

145 2.4. Oil hazard

146 Mapping of the oil covering was based on the delimitation of areas susceptible to simulated oil slicks in the
147 nine years of analysis (2007–2015) and was properly calibrated and validated for hydrodynamic conditions
148 (Monteiro et al., 2019a,b). Therefore, oil particle distribution areas for the 17,382 scenarios allowed an
149 approximation of the spatio-temporal occurrence of these events. Each oil spill simulation used a set of
150 10,000 particles to discretise the oil slicks; this number of particles ensures that the oil pollution estimation
151 depends on the patterns of physical force that dictates oil movement in water, rather than the volume of
152 oil in each particle. The estimation of the probability of oil covering $P(x, y)$ was centred on the longitude
153 x and latitude y of each element of the computational grid. As was suggested by Goldman et al. (2017),
154 contamination is confirmed by the presence of at least one particle in the batch of 10,000 particles released.
155 Thus, hazard maps were generated for all points in the computational domain that were hit by at least one
156 particle at any simulation time step, divided by the total number of particles for each simulated scenario. **A**

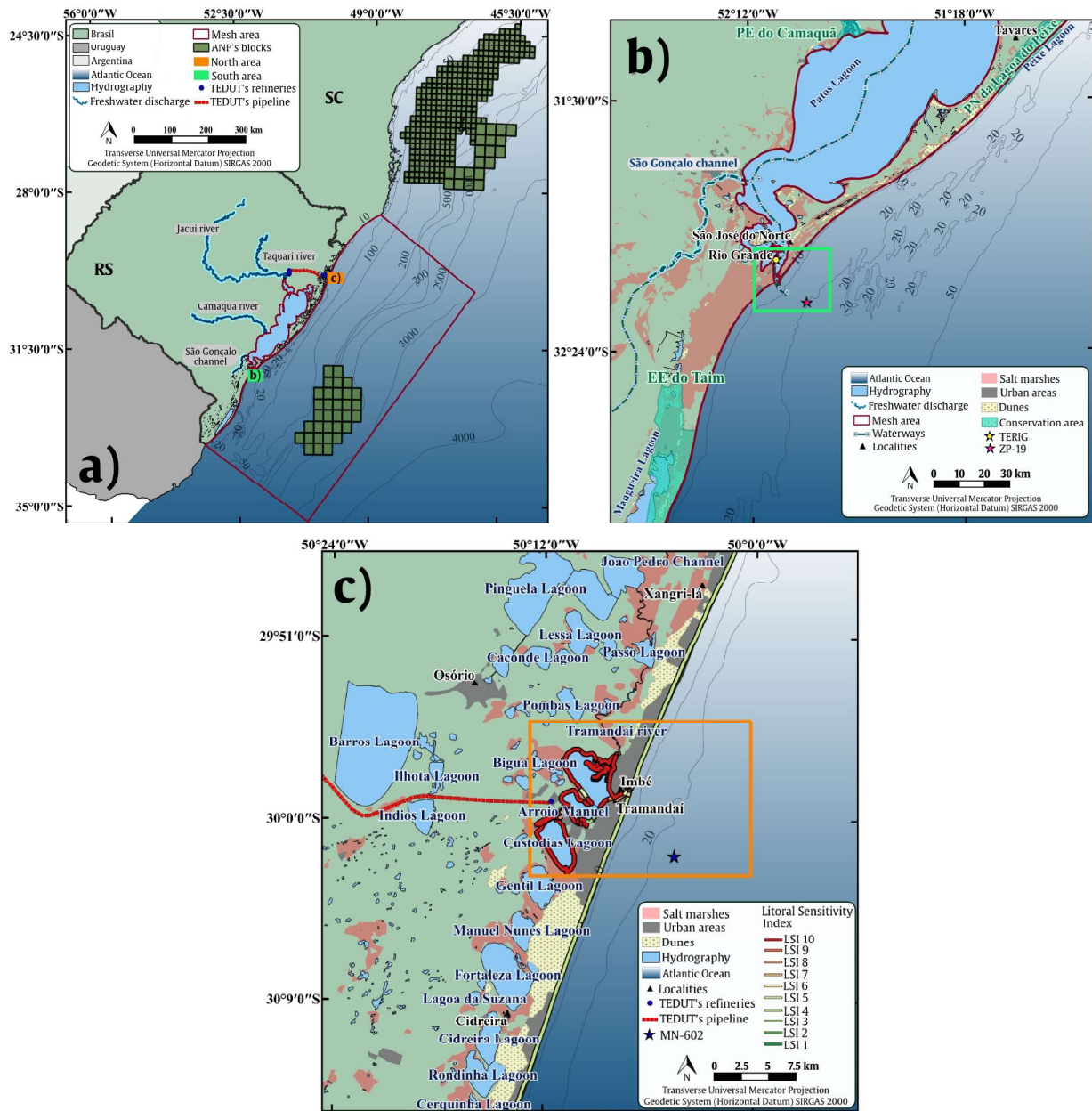


Figure 2: (a) The Southern Brazilian Shelf (SBS) and detail of the (b) south region of the study area: green box and star highlight estuarine and manoeuvring zone near the lagoon access channel respectively; and the (c) north region of the study area: orange box, with blue star highlighting the buoy near the coast.

157 limitation of the methodology is that the probabilities of oil covering were not estimated by categorising the
158 volume of oil accumulated in the studied area. Therefore, this form of analysis may have overestimated the
159 vulnerability in some regions with a high probability of oil covering, even under low concentrations.

160 The oil spilt at Patos Lagoon estuarine area is mainly driven by the strong estuarine currents, the
161 dynamics of which essentially depend on wind and freshwater discharge levels (Möller Jr et al., 1996, 2001;
162 Fernandes et al., 2001; Tavora et al., 2019), which leads to two main scenarios, flood and ebb currents. Thus,
163 to better represent this dual flow, the analyses were conducted to consider separate oil spills under flood or
164 ebb currents. To create better spatial orientation in the estuarine zone, the main critical oil sensitivity sectors
165 were highlighted. It is worth pointing out that, even though these sectors had been recognised as ultra-
166 sensitive areas, they did not receive any additional weight when estimating their integrated environmental
167 vulnerability. The environmental sensitivity of these ultra-sensitive areas was weighted in the same way as
168 all other areas assessed in this study (i.e., according to their ISL). The Kernel Probability Map concept
169 (Mahmoodi and Sayedi, 2015) was used to measure the oil pollution probability for the study region as a
170 matrix with a fixed number of rows and columns. The rows are related to the number of kernels in a window,
171 the columns are related to the number of bins for each kernel, and each pixel is related to one kernel. In this
172 study, a 5,000 × 5,000 window was assumed; thus, 25,000,000 kernels were possible. A quartic function and
173 adaptive radius were used with SIG techniques and kernel maps in *TerraView* (Alves et al., 2019; Ferreira
174 and Nascimento, 2019; Gehlen et al., 2019). These parameters were selected based on the works of Alves et al.
175 (2019); Ferreira and Nascimento (2019); Gehlen et al. (2019) and were adapted according to the susceptibility
176 area dimensions that resulted from the oil spill simulations. A probability function was estimated for each
177 kernel, which allowed the calculation of each of those. Values of the probability for oil pollution, which are
178 specified between 0 (0% probability) and 1 (representing the maximum oil coverage), were considered for all
179 17,382 simulated oil spills.

180 2.5. Littoral Sensitivity Index Information

181 The LSI from the Brazilian Oil Sensitivity Charts (SAO charts) for Pelotas Basin (MMA, 2007; Nicolodi,
182 2016) were used to obtain oil sensitivity information. The LSI classifies the coastlines according to an
183 environmental sensitivity scale that ranges from 1 as the lowest sensitivity to 10 as the highest sensitivity.
184 This index considers the levels of exposure to wave and tidal energy, the slope of the coast and the type of
185 substrate (Jensen et al., 1998; Petersen et al., 2002; Wieczorek et al., 2007). The study area was classified
186 based on LSI at 1:50,000 scale (PEL101, PEL109–PEL115³). Then the study area was hierarchised in five
187 levels of severity; the weight in zones with the highest LSI values, such as environmental conservation areas,
188 urban zones next to fisheries and touristic areas, were increased.

189 2.6. Vulnerability classification

190 Hazard maps that contained the probability of oil covering, the LSI information and marine oil spill
191 intensity levels were combined and integrated using GIS techniques in QGIS to estimate vulnerability.

192 Based on previous works (e.g., Zafirakou et al. (2018); Ha (2018); Nelson and Grubescic (2018); Alves et al.
193 (2016a); Nelson et al. (2015)), after selecting six criteria related to oil vulnerability, pairwise comparisons
194 were used to calculate priorities using Analytic Hierarchy Process (AHP) analysis, following the methodology
195 of Saaty (1990). This generated 15 pairwise comparisons (Tab. 1) according to the matrix with the AHP
196 priorities, and ranked which criterion was more important on a scale from 1–9. This pairwise matrix was
197 reliable and presented a Consistency Ratio (CR) of 0.002 as defined by Saaty (1990). The highest weight
198 values were assigned to important infrastructures, touristic sites, and highly sensitive coastal regions, due to
199 their morphology and environmental structure.

200 For visualisation purposes, the vulnerability scores were classified into five categories that corresponded
201 to different vulnerability levels: high (two levels: > 75 % and 50–75 %), medium (25–50 %) and low (two
202 levels: 15–25 % and < 15 %); these were adapted from Nelson et al. (2015).

³www.mma.gov.br/seguranca-quimica/cartas-de-sensibilidade-ao-oleo/base-de-dados.html

Table 1: Factors and pairwise comparisons used to calculate priorities of the AHP. AHP Scale: 1*–Equal Importance, 3–Moderate importance, 5–Strong importance, 7–Very strong importance, 9–Extreme importance (2, 4, 6, 8: values in-between).

| Factors | Priority (%) | Rank | 1 | 2 | 3 | 4 | 5 | 6 | |
|-----------------------------|--------------|------|----------|------|------|------|----|----|----|
| Probability of oil covering | 28 | 1 | 1 | 1* | 1 | 1 | 5 | 6 | 5 |
| Oil slick concentration | 28 | 1 | 2 | 1 | 1* | 1 | 5 | 6 | 5 |
| LSI levels | 28 | 1 | 3 | 1 | 1 | 1* | 5 | 6 | 5 |
| Evaporation rates | 5.4 | 4 | 4 | 0.2 | 0.2 | 0.2 | 1* | 1 | 1 |
| Emulsion mass | 5.0 | 6 | 5 | 0.17 | 0.17 | 0.17 | 1 | 1* | 1 |
| Oil beaching time | 5.4 | 4 | 6 | 0.2 | 0.2 | 0.2 | 1 | 1 | 1* |

3. Results and Discussions

An integrated oil spill vulnerability could be useful to mitigate oil spreading within few hours of its onset, before wind and currents disperse the slick (Alves et al., 2016b). Such data can also guide response protocols to prioritise the protection of critical areas. The simulated spills in estuarine region presented similar behaviour to that found in the results of previous studies for this area (e.g., Janeiro et al. (2008); Monteiro et al. (2006); Stringari et al. (2012); Lopes et al. (2019); Leal et al. (2019)). Hence, existing data of salinity, current velocity and wind direction were analysed along cross-sections in the estuarine region, which allowed the distinction of oil spills occurring under ebb or flood flows throughout the navigation channel and the inner shelf next to the starting point of spills (Fig. 3).

Under dominant ebb conditions, the oil has a longitudinal dispersion and reaches larger areas, is both deeper and shallower (Fig. 3b, d) and spread towards both inland regions and entrance of the access channel. On the other side, oil spills under flood conditions (Fig. 3a, c) led to a higher probability distribution of oil pollution over inland estuarine areas. For these two hydrodynamic conditions, the oil spill scenarios agreed with the results of previous works (e.g., Tomazelli (1993); Möller Jr et al. (1996); Möller and Castaing (1999); Möller Jr et al. (2001); Tavora et al. (2019)).

Due to the action of northeastern (NE) winds (Fig. 3b) in higher frequencies, ebb flows promote lagoon flux in a maritime direction (Fig. 3d), which is associated with high river discharges in situations where NE winds dominate, especially during winter and spring. Such settings favour the movement of oil slicks along the navigation channel. As presented in Fig. 3a and c, beyond the contribution of NE winds, contributions of strong southeast (SE) and southwest (SW) winds also increase the oil dispersion into shallower and estuarine areas under flood conditions. Here, the oil covering probabilities can reach such localities as the Franceses Bridge (P2) and Mangueira Bay (P3), both of which have significant environmental and economic importance to the region.

Different uses of water and resources in these areas can be affected. As an example, oil spills can jeopardise the artisanal fishery in this region, which has existed since the end of the 19th century and is still the main source of income for some communities (Pereira and D’Incao, 2012), supporting about 6,000 families throughout the coastline of Rio Grande do Sul state (Kalikoski et al., 2002). The studied Patos Lagoon estuary is an important environment for the life cycle of many fish, mollusc and crustacean species, notably because of its role as a sheltered area that allows their development from early larval stages (Castello, 1985; Muelbert et al., 2008; Vieira et al., 2008)) and offers an abundance of food (e.g., zooplankton) and protection from predators (Martins et al., 2007). These sheltered areas are also frequently used for fishing (D’Incao, 1991) and for the growth of estuarine juvenile pink shrimp and fishes (Vianna and D’Incao, 2006; Oliveira and Bemvenuti, 2006; Benedet et al., 2010).

According to Fig. 3, even though the radius of pollution can be extended to further areas, such as Torotama Island (P7), the highest oil pollution probabilities are near the Franceses Bridge (P2), Mangueira Bay (P3), New Harbor (P4), Marinheiros Island (P6) and the city of São José do Norte (P8).

As described by Alves et al. (2016a), these oil distribution probability scenarios can trigger more damages, since sheltered areas from the open sea often trap large quantities of oil, which worsens pollutions scenarios. Besides, salt-marshes and other highly vegetative locations with high LSI values are not only critical areas

241 because they are the most productive aquatic environments, but also because they are difficult to clean
 242 up and oil particles can persist in these places for years (Gundlach and Hayes, 1978; Heflin and Wallace,
 243 2017; Marinho¹ and Nicolodi, 2019). Diversely, under prevailing NE and SW local winds, high continental
 244 freshwater supply (Fig. 3b and d) of predominant ebb conditions cause the highest pollution probabilities to
 245 remain in the middle of the estuary.

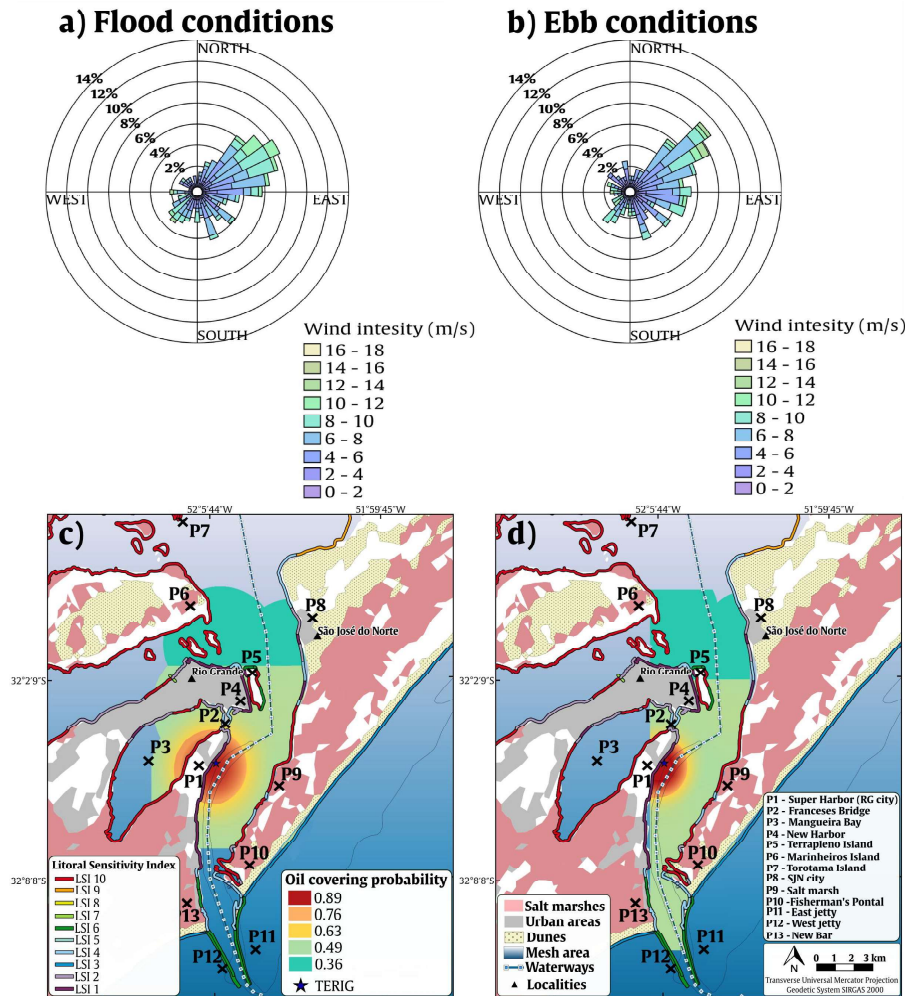


Figure 3: Probabilities for the estuarine area: under flood (a: wind velocity; c: hazard) and ebb (b: wind velocity; d: hazard).

246 Under flood flows, winds from the SW that are associated to east (E) and northwest (NW) quadrant
 247 currents contribute to the oil displacement over a large area. In this case, the hazard probabilities were also
 248 verified, both in the estuarine interior and in the portions closest to the lagoon inlet. Oil reached areas with
 249 varying levels of oil sensitivity, namely the Franceses Bridge (P2), Mangureira Bay (P3), New Harbor (P4),
 250 Marinheiros Island (P6), São José do Norte (P8), Torotama Island (P7) and vicinities of the lagoon entrance
 251 channel (P10–P13). In these cases, the most susceptible areas are those next to the Rio Grande port (P1) in
 252 a region near the oil release point due to the fast oil arrival time of an average of one hour to the nearby
 253 coastal structures.

254 In both scenarios, areas such as Fisherman's Pontal (P10) and the jetties (P11, P12) — especially the
 255 east jetty (P12) — were under hazard due to an increased interest in the conservation of marine species, such
 256 as pinnipeds (Kinas et al., 2005). Despite being distant from the oil release point, these locations were also
 257 susceptible to being covered by oil, with a probability of 0.49. Furthermore, these areas presented features,

258 such as salt marshes, coastal lagoons, bays and natural reserves, which cause them to be more susceptible to
 259 oil trapping.

260 For oil spills that started at the nearby continental shelf, a set of susceptibility areas to oil effects resulted
 261 in a greater spreading of areas likely to be contaminated (Fig. 4a). The highest hazard probabilities (i.e.,
 262 values between 0.89–0.72) were confined to the smallest area in the vicinities of the Patos Lagoon inlet, and
 263 also overlapped important beaches in the region (Cassino & Barra de São José do Norte beaches) when
 264 winds and currents disperse them. Critical areas as the east jetty (P12) presented the highest oil-coverage
 265 probability (i.e., 0.89).

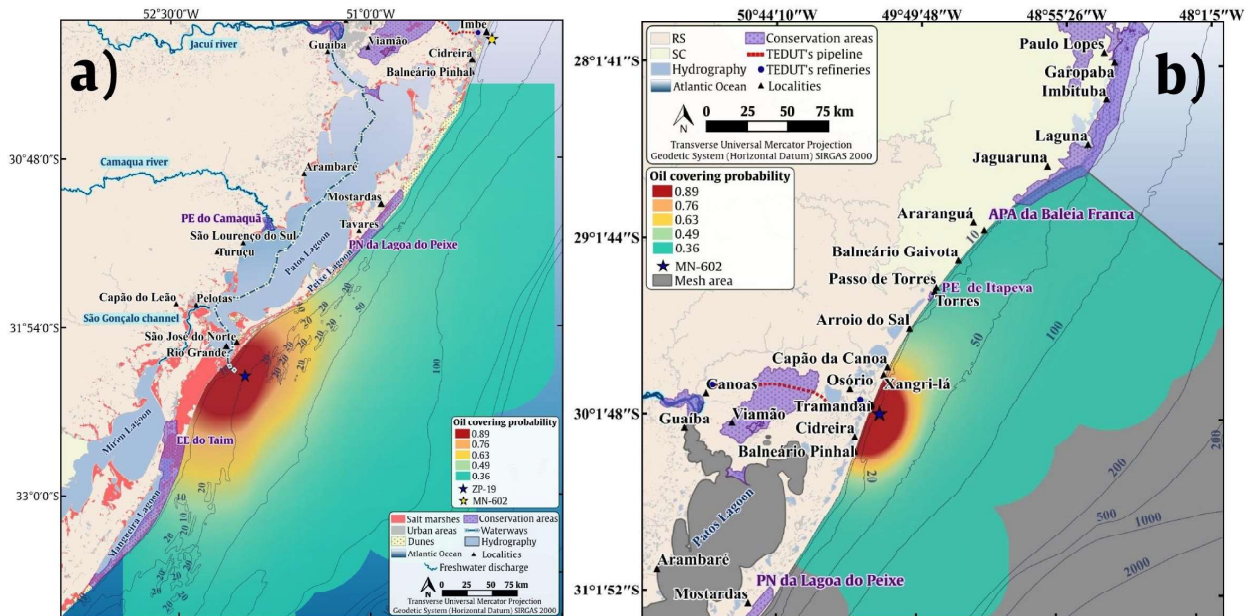


Figure 4: Oil covering hazard for (a) the manoeuvring zone near the lagoon access channel (ZP-19) and (b) the oil collector buoy (MN-602).

266 Despite this, the inherent danger of oil dispersal in this shore face has resulted in several threats over
 267 a wide area. Even though there were relatively low pollution probabilities (from 0.49–0.36), oil particles
 268 still reached environmental conservation sites, such as the Taim Ecological Reserve and the Lagoa do Peixe
 269 National Park. Meanwhile, this location is important to freshwater distribution that comes from Patos
 270 Lagoon and has environmental relevance because it forms an oceanfront mass of low-dynamic waters; this
 271 low-dynamic flow area was discussed in previous works (e.g., Soares et al. (2007a,b)). These results show that
 272 in the vicinities of Patos Lagoon plume, there is a tendency for high stratification at the lagoon discharge
 273 point. A low stratification area is then formed around it, and finally, high stratification values are found at
 274 the offshore border of the plume. This configuration helps the formation of a low-dynamic flow area in the
 275 inner shelf of the Lagoon plume, which plays an important ecological role (Soares et al., 2007a).

276 As presented in Fig. 4a, the highest probabilities of oil covering (i.e., 0.89–0.63) reached localities with
 277 relevant ecological functions and touristic importance. Beaches and fishing points can be jeopardised by oil
 278 spills in these regions (Terceiro, 2017, 2018), which threaten coastal lagoons, beaches, dunes and salt marshes.
 279 Although under lower pollution probabilities (i.e., 0.49), the effects of spill cases for the north point can hit
 280 long extensions along the SBS and reach conservation areas like Peixe Lagoon National Park and the whales
 281 environmental protection area (i.e., APA Baleia Franca) at the northern end of the computational grid.

282 For the three starting points studied, the potential hazards of these accidents were based on the calculation
 283 of a variable capable of predicting the presence and absence of oil particles within a given mapping unit.
 284 Thus, different features, such as strips of beaches, shorelines, coastal lagoons and salt marshes, were analysed.
 285 Concerning the vulnerability mapping, the challenge of the analysis was to identify regions with different

286 potential harm when faced with the impact of an oil spill. This step was preceded by calculations of oil
 287 dispersion areas, which also indicated how oil slicks spread and behave within different affected environments.

288 Different weights were attributed to LSI values so that the highest oil sensitive levels received the highest
 289 weights because they portrayed more critical situations. This weighting increased the vulnerability level of
 290 areas that had high concentrations of surface oil, while the weight was reduced for coastal and shelter areas
 291 with low LSI. Finally, as it relates to oil-covering probability, the LSI values and the spill intensity received
 292 equally high weights (i.e., 28 %—Tab. 1). All weighted factors were proportionally relevant to classify the
 293 study area in terms of vulnerability. This justifies, for example, the reasons why the areas with the highest
 294 oil-covering probability do not necessarily correspond to the most vulnerable areas.

295 For the estuarine oil spill starting points, two conditions of vulnerability were estimated; one was
 296 predominantly in flood conditions (Fig. 5a) and another was under ebb flow (Fig. 5b). When observing the
 297 differences between the vulnerability mapping for the same area, it was a possible to consider the relevance
 298 of physical forces, such as weather and oceanographic conditions, that work as a background to condition the
 299 movement, fate and behaviour of oil spill accidents.

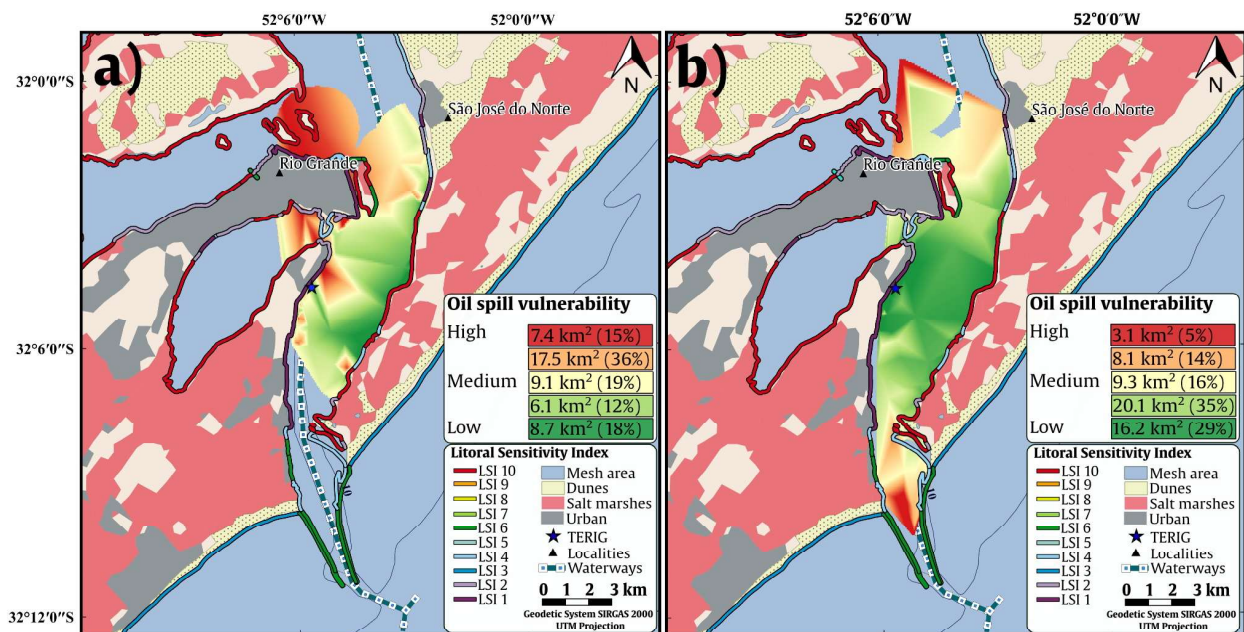


Figure 5: Vulnerability of oil spills started on the estuarine area under (a) flood and (b) ebb flow conditions.

300 Under more frequent autumn and winter winds (S and SW) that are associated with lower river discharges,
 301 flood flows predominate and favour saline wedge entrance from the inlet towards the innermost portions of the
 302 Patos Lagoon estuary. This flux also forces the superficial current velocities that conduct oil slick movement
 303 together with flood currents to become more prominent for these extreme hydrodynamic conditions (Fig. 5a).
 304 The scenario presented in Fig. 5a also showed areas with high vulnerability levels that represent about 51 %
 305 of the total area mapped (24.9 km²), with high oil-sensitive areas in internal sections of the lagoon, such as
 306 Terrapleno, Marinheiros and Torotama Island.

307 Under ebb flows (Fig. 5b), the oil slicks spread along wider areas, and only about 19 % (11.2 km²) of them
 308 present the highest vulnerability score, followed by another portion with a medium level of vulnerability (16 %—
 309 9.3 km²). Hence, both of these areas are the most critical because of their greater vulnerability. Thus, a possible
 310 strategy to provide more urgent responses is to concentrate civil protection teams in these critical areas.

311 With an opposite effect, Fig. 5a shows the flood conditions, in which oil vulnerability area distribution is
 312 mainly ruled by estuarine currents forced by remote and local wind effects (i.e., the NE quadrant with higher
 313 frequencies) and high river discharges. These areas occur towards the inlet zone and generate oceanographic

314 fronts that are freshwater plumes over the heavier salty waters of the shoreface. Here, the vulnerable areas
 315 tend to be limited to salt marshes, bays and other sensitive areas for the biologic reproduction of local species
 316 and fishing activities. Under this flow condition, the highly vulnerable areas were less frequent (5%) due to
 317 the major oil slick dispersion, despite the total area of 56.71 km² of vulnerable regions.

318 For the nearby continental shelf, the vulnerability areas were spread out, representing an area of 35,937 km²
 319 (Fig. 6), the largest among the modelled scenarios. This spreading is probably due to a stronger action of
 320 currents and local wind effects that condition the local hydrodynamics. About 23% (8,323 km²) of the total
 321 vulnerable areas correspond to the highest scores; most are located next to the most sensitive areas with
 322 severe conditions of oil-covering probability.

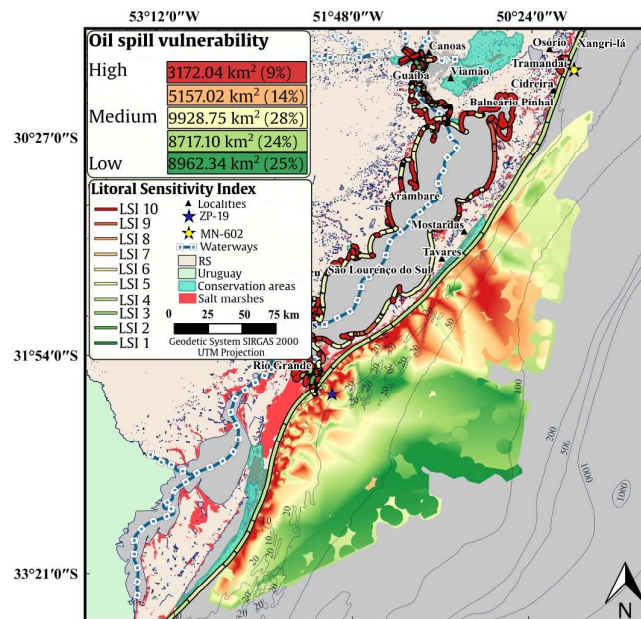


Figure 6: Vulnerability to oil spills started in the manoeuvring zone near the lagoon access channel (ZP-19).

323 Additionally, previous studies in this region reported the accumulation of spawning fish and plankton next
 324 to the lagoon access channel, which emphasises that the low dynamic in this area is an important feature for
 325 fish spawning and larval development (Soares et al., 2007a). For these reasons, accidental oil spills are a
 326 severe potential threat in these areas, especially in regions with oceanographic fronts like the studied region,
 327 due to the Patos Lagoon plume also being capable of dragging oil slicks, thereby jeopardising different species
 328 of local fauna and local activities of economic and social importance.

329 For the northern scenarios, about 28% of these vulnerable areas (5,123 km²) correspond to high levels
 330 of severity to oil slick exposures (Fig. 7). In addition to beaches, coastal lagoons located within about 30%
 331 of the littoral zones (Nicolodi, 2016) also showed high vulnerability to oil spills. Coastal lagoons in this area
 332 have a depth ranging from 1–2 m, which allows sunlight to penetrate the entire water column. This situation
 333 is favourable for the development of algae, which is the food base for fish and crustaceans, making this region
 334 a natural nursery for these species (Cotrim and Miguel, 2007). Therefore, environmental hazards that arise
 335 from oil spills in this region can lead to both ecological and economic losses (Terceiro, 2017), similar to what
 336 happened in a recent and severe oil spill in shallow areas on the Brazilian coast (de Araújo et al., 2020; Pena
 337 et al., 2020). As was verified by Terceiro (2017, 2018), the region's fishing activities are in conflict with the
 338 oil industry. According to the author, the vulnerability of fishing resources and the fisherman economy is
 339 high with respect to oil enterprises in which the buoy considered by the study is located. Considering the
 340 mapping of the types of fishing in the region and the results of the present study, areas of greater vulnerabil-
 341 ity (from high to medium), which extend over about 41% of the mapped area (7,441.63 km²), are confronted
 342 with sport, boat and net fishing.

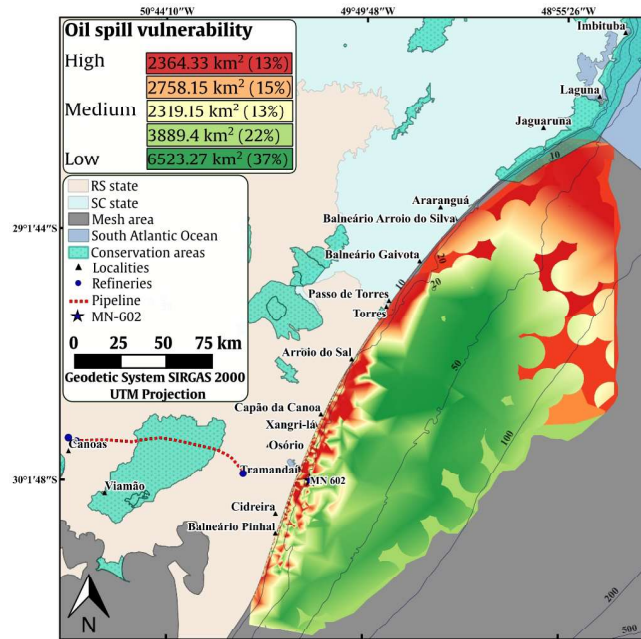


Figure 7: Vulnerability to oil under spills started on the buoy MN-602 point, near the Tramandaí coastal zone.

343 Moreover, the present vulnerability mapping can predict the pollution of fragile areas about which Alves
 344 et al. (2015) warned, and salt marshes and coastal lakes can draw water from foreshore areas polluted with
 345 oil, thereby contaminating sub-surface aquifers and soil for a long time after an oil spill. These findings
 346 can improve oil spill response in sensitive areas, based both on static mapping information for expeditious
 347 measurements, such as SAO chart work, and actions that consider dynamic variables integrated with the
 348 characterisation of the environment involved in oil spill cases. Furthermore, the classification of areas based
 349 on the vulnerability levels thereof can optimise civil protection management by promptly directing efforts
 350 into critical areas (Alves et al., 2016a).

351 In addition, the results agree such classical works as (Gundlach and Hayes, 1978) and (Hanna, 1995),
 352 since the vulnerability concept allows determination of which coastal environmental would be more seriously
 353 damaged by oil contamination so that they can receive priority protection. Urgent responses in high-
 354 vulnerability areas are important to anticipate the costs of oil spill responses, as was presented by Etkin (2000),
 355 because costs are estimated in terms of the affected area and the degree of shoreline fragility, among other
 356 variables. Thus, as the resources and equipment for oil contingency actions are usually limited, these results
 357 can help to mitigate major oil spill accidents and the undesirable effects thereof on marine environment.

358 4. Conclusions

359 The aim of this paper was to improve the body of knowledge concerning an integrated vulnerability
 360 assessment for oil spills in sensitive areas by incorporating modelled data, oil covering hazard probability and
 361 littoral sensitivity information, among other data related to marine oil spill behaviour. The modelling system
 362 utilised consisted of coupling a 3D hydrodynamic model (i.e., TELEMAC-3D) with an oil-spill model (i.e., ECOS)
 363 and employing the results from global meteorological and oceanographic models as boundary conditions. A
 364 set of 17,382 oil spills performed along nine years of simulations (2007–2015), together with information on
 365 environmentally sensitive elements of oil effects, made the estimation of areas susceptible to oil spills possible.

366 The results showed that circulation patterns have a strong influence on the oil-covering probability
 367 and vulnerability estimates. The enhanced methodology allowed: (1) accurate base maps to be consulted
 368 whenever a marine oil spill case occurred in an environmentally sensitive region; (2) a detailed estimation of

369 susceptible and vulnerable areas; (3) the use of a relational GIS database to store and integrate information
370 on oil-sensitive areas, shorelines, bays, coastal lagoons and estuarine features; and (4) the identification of
371 oil-pollution-critical areas when the origin of the simulated spill is close to shipping routes, pipelines, gas
372 wells, single-buoy moorings and other oil-spill-risk sources.

373 In addition, the outputs of this study can inform the process of systematic planning in the event of oil
374 spills. The identification of priority areas for response actions can maximise the achievement of conservation
375 goals while minimising the expected threats and losses. Furthermore, the vulnerability scenarios can be used
376 as a background to recommend possible cleaning and containment methods. Finally, these findings can help
377 guide pollution-control actions in areas under greater vulnerability, thereby prioritising timely interventions
378 in sensitive areas in the aftermath of oil-spill accidents.

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