The impact of thermal pretreatment on various solid-liquid ratios of Palm Oil Mill Effluent (POME) for enhanced thermophilic anaerobic digestion performance

7 Sabeeha.N.B.A.Khadaroo^a, Paul Grassia^b, Darwin Gouwanda^a, Phaik 8 Eong Poh^{a*}

^a Chemical Engineering Discipline, School of Engineering, Monash University Malaysia, Jalan Lagoon Selatan,47500,
 Bandar Sunway, Selangor, Malaysia

16 *Corresponding Author Email: <u>poh.phaik.eong@monash.edu</u> Tel: + 603 55146272 Fax: + 603 55146207

18 Abstract

19

17

9

An innovative treatment process for the treatment of palm oil mill effluent (POME) was 20 proposed whereby a pretreatment technology and a dewatering device are introduced into the 21 existing treatment process. Thermal pretreatment is a foolproof technique with the ability to 22 23 enhance the rate and ameliorate the biogas production of anaerobic digestion. The dewatering device will infer a means of control on the digesters' load, allowing the removal of microbes 24 and impurities as well as assist in the residual oil removal. The novel treatment process allows 25 the removal of cooling ponds making the treatment process more sustainable in terms of the 26 substantial reduction in the amount of greenhouse gas emission, improved residual oil removal 27 efficiency in the waste stream, and better treated effluent quality. However, to be able to 28 29 implement this innovative treatment method effectively, it is fundamental to know how thermal pretreatment effected the solid content of POME impacts on the anaerobic digestion process 30

31	performance. To undertake the study mentioned above, POME was pretreated at 120°C and
32	was allowed to settle to separate the solid and liquid phases. The chosen method of anaerobic
33	digestion was batch thermophilic anaerobic digestion, which was conducted on various solid:
34	liquid ratios (i.e., 20S:80L, 40S:60L, 50S:50L, 75S:25L, and 100S). It was found that the
35	optimal ratios were 20S:80L and 40S:60L, which generated approximately 9-fold and 6-fold
36	higher methane yield, respectively, in contrast to their untreated counterparts. Thermally
37	pretreated 40S:60L solid loading exhibited a higher removal efficiency in terms of chemical
38	oxygen demand (COD), biological oxygen demand (BOD), total suspended solids (TSS), and
39	oil & grease (O&G), a higher methane yield of 328.73 mL CH ₄ /g COD _{removed} and biogas
40	production of 1886.11±21.63 mL compared to all the other pretreated and untreated ratios.
41	Keywords: Palm oil Mill Effluent, Thermal Pretreatment, Solid loadings, Biogas Production,
42	and Methane Yield.
43	
44	
45	
46	
47	
48	
49	
50	
51	
52	
53	
54	
55	
56	
57	

58 Contents

59	Abstract	1
60	1. Introduction	1
61	2. Materials and Methods	7
62	2.1 Materials	7
63	2.2 Experimental Set up	7
64	2.2.1 Solid-Liquid Separation	7
65	2.2.2 Thermal Pretreatment	3
66	2.2.3 Anaerobic Digestion experimental procedure	Э
67	2.3 Physico-chemical Analysis1	C
68	3. Results and Discussion1	C
69	3.1 Chemical properties analyses in terms of COD, BOD, TSS, and O&G removal1	7
70	3.2 Biogas Production and Total number of anaerobes2	1
71	3.3 Methane Composition in the biogas produced2	3
72	3.4 Methane Yield2	5
73	4. Conclusion2	Э
74	5. Acknowledgment	C
75	Nomenclature	C
76	6. References	C
77		
78 79	Figure 1: Combined cumulative biogas production graphs for untreated POME at different solid: liquid ratios	2
80	Figure 2: Combined cumulative biogas production graphs for thermally treated POME at different	2
81 82 83	Figure 3: Combined daily methane composition graphs for Untreated POME at different solid: liquid ratios	5 5
84 85	Figure 4: Combined daily methane composition graphs for thermally treated solid loadings1 Figure 5: Comparison between the chemical properties of untreated and pretreated solid: liquid	5

90	Table 1: Results obtained for the thermophilic anaerobic digestion of untreated POME at different
91	solid loadings14
92	Table 2: Results obtained for the thermophilic anaerobic digestion of thermally pretreated POME at
93	different solid loadings14
~ •	

97

1. Introduction

98

There are different categories of pretreatment techniques that have been studied over the 99 100 years when it comes to wastewater treatment. Some examples of these pretreatment methods 101 are thermal, biological, chemical, and mechanical pretreatment methods (Carrère et al., 2010). 102 Amongst these pretreatment techniques, some have been commercialized and implemented on 103 the full scale (e.g., thermal pretreatment: conventional heating) whilst others are still at research stages (e.g., chemical pretreatment: advanced oxidation) (Tyagi and Lo, 2011). Pretreatment 104 105 technologies are known to significantly enhance the anaerobic digestion process as they prominently ameliorate the digestion rate, lessen the retention time, and increase the biogas 106 107 production (Khadaroo et al., 2019b). 108 Palm Oil Mill Effluent (POME) is a waste associated with the production of palm oil. It is the 109 primary waste contributor in Malaysia, which is the second-largest palm oil producer worldwide (Lam Man & Lee, 2010). The disposal and waste management of POME is an 110 alarming concern for Malaysia since POME can cause pollution when discharged into 111 watercourses due to the high amount of easily degradable organic matter, thereby referring to 112 the high chemical oxygen demand (COD), biological oxygen demand (BOD) and the total 113 suspended solids (TSS) contents which are of an average concentration of 50,000 mg/L, 25,000 114 mg/L, and 20,000 mg/L respectively (Choong et al., 2018; Iskandar et al., 2018). Another major 115 drawback when it comes to the treatment process of POME is that the chemical characteristics 116 of POME mentioned above typically depend on the efficiency of the palm oil extraction process 117 as well as its chemical characteristics during the high and low crop seasons; owing to these 118 fluctuating parameters, the chemical characteristics of POME tend to vary a great deal thus 119 affecting the efficiency of the treatment process (Poh et al., 2010). Having such high COD and 120 BOD values along with the inconsistency of these parameters, the anaerobic digestion process 121 is subsequently hindered, and a lower anaerobic digestion rate causes the concentration of toxic 122

substances produced in the process to rise (Li et al., 2019). Moreover, owing to the facultative

and the cooling ponds, the current treatment process contributes to significant environmental

- 125 burdens in terms of excessive greenhouse gas emission and water pollution, which is damaging
- 126 to the aquatic ecosystem (Tajuddin et al., 2015).

In the currently implemented POME treatment process, raw POME is directed to a 127 cooling pond where residual oil in POME is scraped off the surface of the cooling ponds to 128 recover crude palm oil (CPO) lost into the waste stream. It is then sent for anaerobic digestion, 129 which is the primary treatment, followed by a secondary treatment process known as the 130 131 aerobic treatment (Bala et al., 2014). The aerobic treatment process is meant to reduce the COD and BOD further; however, the aerobic treatment process consumes an excessively high 132 amount of energy (Singh et al., 2010). The drawback concomitant with the current process is 133 that raw POME discharged at a temperature of 90°C is sent to the cooling ponds which are 134 highly inefficient in terms of the residual oil extraction, and a significant amount of heat is lost 135 to the atmosphere (temperature drops from 90°C to 60°C). POME is then further allowed to 136 cool down to an adequate mesophilic temperature (35°C) before being sent for anaerobic 137 digestion. Furthermore, often the treated effluent quality from the current treatment process is 138 inadequate and does not conform to the stringent environmental standards stipulated by the 139 Malaysian Government, stating that the COD concentration of the treated effluent should be 140 no more than 50mg/L (Chin et al., 2013; Ahmed et al., 2015). 141

Subsequently, a substitute treatment process was recommended, which entails eliminating the cooling ponds and introducing a pretreatment technology coupled with a dewatering device such as a thickener prior to the anaerobic digestion process. The pretreatment technology will assist in enhancing the rate-limiting step of anaerobic digestion namely the hydrolysis step while the dewatering device will contribute to making the anaerobic digesters more efficient by inferring a means of control on the digesters' load during the

treatment process of POME by regulating the amount of solids and liquid required in the system 148 to ensure best conditions for the consortium of bacteria to thrive (Li et al., 2019). Another 149 advantage of the dewatering device is that it will aid in controlling the load of the digesters 150 making the anaerobic digestion process comparatively more stable when there is a drastic 151 change in the chemical characteristics of POME as it varies from high crop season to low crop 152 season (Khadaroo et al., 2019a). The pretreated POME would then be directed to the anaerobic 153 154 digesters, which will operate at a higher thermophilic temperature (55°C) since thermophilic anaerobic digestion can achieve a higher solids reduction and biogas production, better 155 156 resistance to foaming, better pathogens destruction and enhancement of the energy balance of the whole treatment process (Dohányos et al., 2004). The treated effluent quality will, 157 therefore, improve considerably, reducing the load on the aerobic treatment, which will, in turn, 158 reduce the energy consumption of the latter (Appels et al., 2011). 159

Khadaroo et al. (2019b) have thoroughly studied the effect of different pretreatment 160 techniques on the anaerobic digestion of POME from various literature sources and assessed 161 which method was most suitable in terms of treated effluent quality, biogas production, 162 environmental encumbrance as well as the technical and economic feasibility. They found that 163 for the treatment of POME, thermal pretreatment was the most adequate option as the energy 164 consumption could be easily compensated by the amount of biogas produced, and the treated 165 effluent quality can be significantly improved. Thermal pretreatment is also known to enhance 166 digestion efficiency, and increase process performance as well as stability (Farhat et al., 2018). 167 Thermal pretreatment can improve dewaterability, which will further assist in the solid: liquid 168 separation (Higgins et al., 2017). These alterations in the POME treatment process will notably 169 enhance the anaerobic digestion process, upsurge the amount of biogas produced, provide 170 better treated effluent quality, all while drastically reducing the greenhouse gas emission to the 171 atmosphere via the ponds. The biogas generated will be captured and utilized as per the mill 172

174 reduction in greenhouse gas emission and waste produced since with a better treated effluent

175 quality, the treated solids can be used as A-grade fertilizers in palm tree plantations.

However, to comprehend how the thermal pretreatment of different solid loadings affects anaerobic digestion of POME and to find out at which solid loading to operate the digesters for optimal anaerobic digestion performance, it is fundamental to methodically understand how these two factors influence anaerobic digestion of the POME. Owing to a distinctly restricted amount of data on the topic, this paper aims to bridge the gap on the influence of the various solid-liquid ratios of POME and the impact of thermal pretreatment of these ratios on the anaerobic digestion performance.

- 183 2. Materials and Methods
- 184

191

185 2.1 Materials

The substrate for anaerobic digestion used is POME. POME was sampled at the Sime
Darby East Oil Mill, Malaysia. The POME collection site lies within coordinates 2.8843° N,
101.4369° E. The temperature of POME at the sampling location was recorded to be 65°C.
Anaerobic seed sludge, which was used as inoculum, was also collected from the same mill.

- 190 2.2 Experimental Set up
 - 2.2.1 Solid-Liquid Separation

To separate the solid from the liquid phase, 5L of POME was carefully poured in a settling column of height 0.7 m with multiple sampling ports to allow the removal of each phase. POME at 65°C was used in the process for the suspension to settle effectively as a study conducted by Khadaroo et al. (2019a) indicated that POME at room temperature does not settle but instead tends to form lumps of solid flocs that float along the settling column. The solid flocs in hot POME were left to settle for approximately 24hrs to ensure the distinct separation of the phases. Samples of settled suspension and clear liquid were hereafter designated ``solid"
and ``liquid" in this study.

200	After the settling of raw POME into solid and liquid phases, they were recombined to
201	achieve the desired ratios by volume for this study (i.e., for a feed volume of 100 ml, the
202	20S:80L ratio consists of 20 mL of settled suspension and 80 mL of clarified liquid). The opted
203	solid: liquid ratios were 20S:80L, 40S:60L, 50S:50L, 75S:25L, and 100S:0L (100S). Anaerobic
204	digestion experiments on both thermally treated and untreated POME of the ratios mentioned
205	above were undertaken to gain an insight into how significant the anaerobic digestion
206	performance was improved across all ratios. These specific ratios were selected to allow the
207	study over the full range of solids ratios for a complete picture of the performance of anaerobic
208	digester.
209	2.2.2 Thermal Pretreatment
210	Prior to the anaerobic digestion process for the pretreated experiments, the solids from
211	POME suspension was thermally treated after the separation of the solids from the liquid phase,
212	as described above. The thermal treatment was undertaken by heating the solids at 120°C for
213	an hour in an oven. Thermal pretreatment temperatures can range from 100 to 190°C, and
214	thermal pretreatment effected at temperatures below 150 °C tends to be a more cost-effective
215	process performance as compared to other pretreatment methods (Ariunbaatar et al., 2014).
216	However, 120°C was chosen to ensure the compensation of energy input through the expected
217	increase in biogas production, which can be used for heating purposes in the mill. Since the

temperature is required to be raised by 30°C only, the boost in the amount of biogas produced

can easily compensate for the energy input into the system. To ensure homogenous heating,
the medium was stirred several times during the thermal treatment. Once the solid was
pretreated, the appropriate volume of the liquid phase was added to the treated solid. The pH
of the system was brought to 7.20 by dosing with 1M NaHCO₃. After pretreatment and

recombining of the specific ratios, the samples were stored in a refrigerator below 4 °C (Wood
et al., 2009).

225 2.2.3 Anaerobic Digestion experimental procedure Thermophilic batch anaerobic digestion was the chosen mode of anaerobic digestion. 226 The requisite ratio was placed in a 250ml Schott bottle with two outlets. The working volume 227 of the system was set as 100mL. The anaerobic seed sludge (inoculum) was acquired from the 228 anaerobic treatment ponds at the mill site. The inoculum volume was taken and maintained as 229 20% of the working volume throughout the experiments. The inoculum was cultivated and 230 231 acclimatized in a reactor heated at 55°C under anaerobic conditions for one month to allow the bacteria to be accustomed to the conditions under which the anaerobic digestions experiments 232 were undertaken as described in a study by Poh and Chong. (2010). The inoculum volume was 233 234 kept constant in all the tested conditions to be able to gauge how thermal pretreatment affects the anaerobic digestion performance of the different solid: liquid ratios, particularly as 235 inoculum to substrate ratio, has proven to affect the anaerobic digestion performance (Cappai 236 et al., 2015). The digesters were heated and continuously stirred using a hot plate magnetic 237 stirrer for appropriate homogenization of the medium. 238

239 Each digester was connected to a water displacement column via a pipe to enable the biogas generation volume measurement. As per ASTM D5511, the pH of the water in the 240 displacement column was adjusted to 2.0 using 1M H₂SO₄ to prevent carbon dioxide from 241 dissolving into the water. The latter ensures that the biogas measurement via the water 242 displacement method is carried out more accurately (Müller et al., 2004). The pH of the 243 medium consisting of the mixture of the solids and the liquid was kept between 6.8 to 8.0 by 244 dosing 1M NaHCO₃ to ensure optimum conditions for the methanogenic bacteria to convert 245 the available substrate into methane gas. To ensure that the pH was within the mentioned range, 246 247 the pH was tested daily. The composition of biogas was measured using the Binder COMBIMASS Gas Analyzer. The experiments were brought to an end when the Binder
COMBIMASS Gas Analyzer detected no more methane in the digesters (Khadaroo et al.,
2020).

Three assays for each ratio (20S:80L, 40S:60L, 50S:50L, 75S:25L, and 100S) were undertaken for which each test consisted of three batch anaerobic digesters. Triplicates were also used to conduct physico-chemical analyses. This set-up produced a total of 9 sets of data for each condition, which presented sufficient data for an efficient experimental analysis for this study.

256

2.3 Physico-chemical Analysis

The chemical characterization tests such as the COD, BOD, TSS, and O&G experiments 257 were carried out before and post anaerobic digestion to determine the quality of the effluent 258 discharged from the anaerobic digesters. The COD, BOD, and TSS tests were performed as per 259 260 the HACH Standard Methods 8000, 8043, and 8006, respectively, while the O&G test was done as per the ASTM method D7066-04. The most probable number (MPN) was carried out 261 as per ASTM STP695. Total solids test was undertaken as per the standard methods for the 262 263 examination of water and wastewater to measure the solid content present in each tested condition (APHA, 1998). All the tests mentioned above are approved by the United States 264 Environmental Protection Agency (USEPA), which is in accordance with ASTM standards as 265 well as the APHA. The methane yield was calculated using the volume of methane produced 266 and the mass of COD removed (Heidrich et al., 2011; Jingura and Kamusoko, 2017). The mass of 267 268 COD_{removed} was used for this study for better accuracy instead of Volatile Solids removed (VS_{removed}) as often used in other literature; since POME consists of 95-96% of water with only 269 4-5% of total solids (Iskandar et al., 2018; Khadaroo et al., 2019b). 270

271 3. Results and Discussion

The results obtained for the anaerobic digestion of untreated and thermally pretreated 273 POME at 20S:80L, 40S:60L, 50S:50L, 75S:25L, and 100S solid loadings are presented below. 274 The error bars in Figures 1,2,3, and 4 were calculated based on the standard deviations from 275 the 9 sets of data obtained. Figure 1 illustrates the daily cumulative biogas production for 276 untreated POME, while Figure 2 shows the daily cumulative biogas production for the 277 thermally pretreated conditions. It can be observed that the anaerobic digestion durations of the 278 thermally treated samples are shorter than that of the untreated conditions, as shown in Tables 279 1 and 3. An explanation for this occurrence may be due to the thermal pretreatment of the 280 281 effluent, the microbial cell walls holding the intracellular matter rupture during pretreatment releasing the intracellular substances in the liquid phase which allowed the microorganisms to 282 break down the latter faster as it is more easily and readily available (Pilli et al., 2015a, 2015b). 283 284 This phenomenon, in turn, noticeably ameliorates the digestion rate, lessens the retention time, and augments the biogas production (Appels et al., 2008, 2011). 285



288 Figure 1: Combined cumulative biogas production graphs for untreated POME at different solid: liquid289 ratios



292 Figure 2: Combined cumulative biogas production graphs for thermally treated POME at different

293 solid: liquid ratios

296 Table 1: Results obtained for the thermophilic anaerobic digestion of untreated POME at different solid loadings

Ratios	Dry Solids Content/ %TS	Initial pH	Final pH	Cumulative Biogas Production/ mL	Maximum Methane Composition /%	Minimum Methane Composition /%	H2S compositi on/ mg/L	Total Anaerobe s/ 100mL	Duration of Experime nt /days	COD removal /%	BOD removal /%	TSS removal /%	O&G removal /%
20S:80L	3.29	7.23 ±0.05	7.52 ±0.03	539.44 ±10.29	73.83 ±2.42	50.36 ±2.07	204±4	1.2x10 ⁶	≈25	62.53 ±1.14	58.07 ±1.76	55.44 ±3.46	26.20 ±0.46
40S:60L	4.02	$\begin{array}{c} 7.21 \\ \pm \ 0.02 \end{array}$	7.74 ±0.09	1431.67 ±17.56	77.33 ± 1.20	57.80 ±2.67	341±16	2.1x10 ⁶	≈20	48.89 ±1.12	32.64 ±1.66	39.57 ±2.26	25.05 ±0.62
50S:50L	5.25	7.22 ±0.12	7.56 ±0.04	1322.78 ±13.62	64.26 ±2.71	54.13 ±0.87	560±21	7.5x10 ⁵	≈18	33.26 ±0.71	23.97 ±1.30	29.40 ±1.55	19.24 ±0.45
75S:25L	6.4	7.27 ±0.10	7.44 ±0.04	1122.70 ±9.94	57.73 ±1.62	40.67 ±0.58	1823±23	1.5x10 ⁵	≈17	23.11 ±0.41	24.10 ±0.52	20.76 ±0.57	26.57 ±0.18
100S	7.86	7.23 ±0.07	7.75 ±0.07	833.88 ±17.11	33.17 ±1.30	20.87 ±3.04	2968±52	9.3x10 ⁴	≈17	7.67 ±1.05	21.78 ±1.84	19.06 ±2.85	31.16 ±0.62

297

298 Table 2: Results obtained for the thermophilic anaerobic digestion of thermally pretreated POME at different solid loadings

Ratios	Dry Solids Content/ %TS	Initial pH	Final pH	Cumulative Biogas Production/ mL	Maximum Methane Composition/ %	Minimum Methane Composition / %	H₂S compositi on/ mg/L	Total Anaerobes/ 100mL	Duration of Experime nt /days	COD removal /%	BOD removal /%	TSS removal /%	O&G removal /%
20S:80L	2.88	7.23 ±0.04	7.53 ±0.02	1471.10 ±15.23	79.23 ±1.34	71.30 ±2.71	176±8	1.5x10 ⁶	≈ 22	84.50 ±1.01	84.41 ±0.15	83.03 ±0.91	82.88 ±0.34
40S:60L	3.98	7.27 ±0.02	7.57 ±0.07	1886.11 ±21.63	83.40 ±0.31	78.83 ±1.31	256±21	4.6x10 ⁶	≈ 19	81.63 ±0.46	81.01 ±1.16	80.72 ±0.16	80.02 ±0.11
508:50L	5.14	7.30 ±0.05	7.48 ±0.02	1509.43 ±4.43	76.97 ±0.73	71.40 ±0.79	392±29	1.1x10 ⁶	≈ 17	65.38 ±0.04	62.72 ±0.36	67.20 ±0.75	64.81 ±0.40
758:25L	6.32	7.20 ±0.02	7.71 ±0.06	1326.13 ±4.74	51.97 ±2.03	43.00 ±2.35	1503±35	2.4x10 ⁵	≈ 15	51.51 ±1.62	50.88 ±0.56	53.32 ±0.36	50.56 ±1.12
100S:0L	7.29	7.22 ±0.01	7.58 ±0.06	970.00 ±2.89	38.20 ±0.75	23.24 ± 1.25	2745±68	1.5x10 ⁵	≈ 15	41.40 ±1.39	40.12 ±2.16	40.59 ±1.84	50.65 ±1.54

300 Figures 3 and 4 depict the methane composition of the biogas produced during the untreated and the thermally treated anaerobic digestion of POME at different solid loadings, 301 respectively. In Figure 4 below compared to Figure 3, it can be noted that the methane 302 composition in the biogas produced is at a higher percentage from day 1 in the treated sample 303 304 whereas it takes longer time for methane conversion to take place in the untreated sample except for 100S solid loading which generated 15.3% of the methane from day 1. An 305 306 explanation for the faster methane conversion which occurs in the treated ratios is owing to the organic matter being readily available for the methanogens to break down and convert to 307 308 methane.



309

Figure 3: Combined daily methane composition graphs for Untreated POME at different solid: liquidratios



313

Figure 4: Combined daily methane composition graphs for thermally treated solid loadings315

After thermal pretreatment, it can be observed that the initial COD, BOD, TSS, and 316 O&G for all the conditions are less compared to the non-treated assays indicating that thermal 317 pretreatment by itself can successfully reduce the COD, BOD, TSS and O&G content of 318 POME. For example, the initial COD, BOD, TSS, and O&G for the untreated 20S:80L solid 319 320 loading were recorded to be 37600±190, 16350±152, 9500±100, and 268±23 mg/L, respectively whereas the initial COD, BOD, TSS, and O&G for the thermally treated 20S:80L 321 solid loading was found to be 30300±125, 12740±190, 11000±200 and 198±20 mg/L, 322 respectively which is comparatively lower than the untreated sample (Pilli et al., 2015a, 2015b). 323

3.1 Chemical properties analyses in terms of COD, BOD, TSS, and O&G removal 324 The chemical properties experiments were undertaken, as described in section 2.3. Nine sets 325 of data were used to evaluate the chemical characteristics discussed below. The chemical 326 properties of the untreated 20S:80L condition prior to anaerobic digestion in terms of COD, 327 BOD, TSS, and O&G were 37600±190, 16350±152, 9500±100, and 268±23 mg/L, 328 respectively; while, the COD, BOD, TSS, and O&G of the untreated 20S:80L after anaerobic 329 digestion was found to be 14088±430, 6855±287, 4233±328 and 197±1 mg/L. Alternatively, 330 after anaerobic digestion of the thermally treated 20S:80L ratio, the COD, BOD, TSS, and 331 O&G significantly declined to 4696±305, 1986±150, 1866±100 and 34±1 resulting in 332 prominent percentage removal of 84.50±1.01, 84.41±0.15, 83.03±0.91 and 82.88±0.31 % of 333 COD, BOD, TSS, and O&G respectively. The treated 20S:80L has 21.97, 26.34,27.59, and 334 56.68 % higher removal efficiencies of COD, BOD, TSS, and O&G than the nontreated 335 20S:80L ratio congruently. 336

337 The COD, BOD, TSS, and O&G evaluated prior to anaerobic digestion for the untreated 40S:60L condition were 47700±120, 20740±138, 16200±113, and 279±13 mg/L, respectively 338 and after anaerobic digestion of the untreated 40S:60L, the COD, BOD, TSS, and O&G was 339 reduced to 24377±537, 13696±345, 9788±366 and 209±2 mg/L. Contrarywise, the initial 340 COD, BOD, TSS, and O&G for pretreated 40S:60L solid loading were 40800±100, 341 342 20090±130, 16000±150 and 208±12 mg/L respectively. Post anaerobic digestion the COD, BOD, TSS, and O&G drastically decreased to 8155±44, 4015±67, 3162±65 and 42±1 343 achieving a notable removal percentage of 80.63±0.46, 81.01±1.16, 80.72±0.16 and 344 80.02±0.11 % of COD, BOD, TSS, and O&G respectively. The treated 40S:60L gave rise to 345 32.74, 48.37, 41.15, and 54.97 higher removal efficiency of COD, BOD, TSS, and O&G 346 compared to the untreated 40S:60L samples. Though the COD, BOD, TSS, and O&G removal 347 348 efficiency is similar to the treated 20S:80L solid loading, the increased biogas production and

methane yield from the treated 40S:60L solid loading is highly advantageous. It was perceived 349 that the anaerobic digestion experiment of the treated 40S:60L solid loading was even more 350 stable than the treated 20S:80L solid loading which is indicated by the smaller calculated 351 percentage deviation as shown in Table 2 in addition to all 9 runs exhibiting similar results in 352 terms of chemical characteristics on top of biogas production. Figure 5 shows the comparison 353 of the chemical properties between the untreated and pretreated solid: liquid ratios after 354 anaerobic digestion. It can be seen in Figure 5 that the COD, BOD, TSS, and O&G after 355 anaerobic digestion of the untreated assays are significantly higher than that of the pretreated 356 ratios indicating that the effluent quality of the pretreated ratios is notably ameliorated. 357



Figure 5: Comparison between the chemical properties of untreated and pretreated solid: liquid ratiosafter anaerobic digestion

361

As for the chemical properties analyses for the 50S:50L assays, before anaerobic digestion, the COD, BOD, TSS, and O&G for the nontreated 50S:50L ratio were 49500±210, 22098±195, 15800±115, and 294±25 mg/L, respectively while post anaerobic digestion, the

recorded COD, BOD, TSS, and O&G were 33033±352, 16799±287, 11155±244 and 237±1 365 mg/L respectively. Conversely, the initial COD, BOD, TSS, and O&G for the pretreated 366 50S:50L solid loading were 48500±200, 21280±125, 20800±135, and 245±15 mg/L, 367 respectively. While after anaerobic digestion the COD, BOD, TSS, and O&G was reduced to 368 16789±22, 7932±77, 6822±156 and 86±1 mg/L. The pretreated 50S:50L sample achieved a 369 percentage removal of 65.38±0.04, 62.72±0.36, 67.20±0.75, and 64.81±0.40% of COD, BOD, 370 371 TSS, and O&G, congruently. Compared to the untreated 50S:50L ratio, the pretreated counterpart attained 32.12, 38.75, 37.80, 45.57% better COD, BOD, TSS, and O&G removal. 372 373 Although the removal efficiency of COD, BOD, TSS, and O&G decreased compared to treated 20S:80L and 40S:60L solid loadings, the removal efficiency of COD, BOD, TSS and O&G for 374 treated 50S:50L solid loading is still superior in contrast to all the untreated tested conditions. 375 It was established that the COD, BOD, TSS, and O&G removal for all the thermally pretreated 376 conditions were much higher than that of the non-treated ones. 377

378 For the untreated 75S:25L chemical analyses, prior to anaerobic digestion, the COD, BOD, TSS, and O&G condition were 52400±145, 24950±230, 22000±162, and 322±35 mg/L, 379 respectively; after anaerobic digestion, the COD, BOD, TSS, and O&G was reduced to 380 40289±212, 18959±129, 17433±126 and 236±1 mg/L. However, the initial COD, BOD, TSS, 381 and O&G the pretreated 75S:25L ratio were 50600±210, 21327±125, 20500±105, and 305±10 382 383 mg/L, respectively. Whereas after anaerobic digestion the COD, BOD, TSS, and O&G declined to 24533±128, 10475±118, 8569±74 and 150±13 mg/L. The pretreated 75S:25L ratio brought 384 about removal efficiencies of 51.51±1.62, 50.88±0.56, 53.32±0.36, and 50.56±1.12 % of COD, 385 BOD, TSS, and O&G respectively. Consequently, the removal efficiencies of COD, BOD, 386 TSS, and O&G for the treated 75S:25L assays were 28.40, 26.78, 32.56, 23.99% greater than 387 the untreated corresponding ratio. 388

An important observation made in the nontreated samples for 75S:25L and the 100S 389 loadings is that the O&G removal was higher than the 40S:60L and the 50S:50L. This 390 occurrence can be explained owing to the formation of a layer of scum at the top of the 75S:25L 391 and the 100S loadings where some of the O&G may have been lost in the scum layer (Khadaroo 392 et al., 2020; Soares et al., 2019). The formation of the scum layer usually occurs in high fat, oil 393 and grease content substracts whereby long-chain fatty acids cause digester instability; as a 394 result, it is more challenging for the bacteria consortium to break down the lipids in the medium 395 (Long et al., 2012; Martínez et al., 2011). On the other hand, for the 75S:25L and the 100S 396 397 loadings pretreated samples, no scum layer was observed as the O&G was broken down by the thermal pretreatment before anaerobic digestion (Pilli et al., 2015a, 2015b). This incidence is 398 seen in the O&G removal trend, whereby it steadily decreases from pretreated 20S:80L to 100S 399 400 solid loadings.

401 Furthermore, the COD, BOD, TSS, and O&G recorded for the nontreated 100S solid loading before anaerobic digestion were 54000±250, 25975±125, 23100±190, and 649±32 402 mg/L, respectively; after anaerobic digestion, the COD, BOD, TSS, and O&G declined to 403 49856±568, 20315±478, 18711±658 and 446±4 mg/L. Alternatively, the initial COD, BOD, 404 405 TSS, and O&G for thermally treated 100S solid loading were 48500±200, 21280±125, 20800±135, and 558±15 mg/L, respectively. While after anaerobic digestion the COD, BOD, 406 407 TSS, and O&G decreased to 30766±692, 14906±788, 13722±425 and 275±20 mg/L. The pretreated 100S resulted in removal efficiencies of 41.40 ± 1.39 , 40.12 ± 2.16 , 40.59 ± 1.84 , and 408 50.65±1.54 % for COD, BOD, TSS, and O&G respectively. Accordingly, the removal 409 efficiencies of COD, BOD, TSS, and O&G for the treated 100S was 33.73, 18.34, 21.53, and 410 19.40% enhanced in contrast to the untreated equivalent solid loading. 411

412 3.2 Biogas Production and Total number of anaerobes

The pretreated 20S:80L samples achieved a maximum biogas production of 1471.10 mL, 413 which was 960 mL more than the untreated sample. A comparison in the performance of the 414 thermally treated assays and the nontreated ones in terms of cumulative biogas production can 415 416 be seen in Figure 6. Based on the results for the individual runs for pretreated 20S:80L solid loading, it was noted that the anaerobic digestion process was more stable than the untreated 417 418 20S:80L solid loading when comparing the percentage error as indicated in Table 1 and 2. The noticeable increase in biogas production is due to more intracellular matter available for the 419 420 methanogens to convert to biogas as well as good substrate/microbes contact in the thermally treated samples. In contrast to the limited amount of solids present in the untreated 20S:20L 421 solid loading medium, which are confined by the cell walls and making it further challenging 422 for the bacteria to access the intracellular organic matter. An indication of the statement 423 mentioned above whereby the microorganisms cannot thrive in such conditions is designated 424 425 by the total number of anaerobes estimated in the untreated and thermally treated experiments 426 as displayed in Tables 1 and 2 (Cappai et al., 2015; Pilli et al., 2015b). It can be clearly seen that the total number of anaerobes in the untreated sample is much less than that of the 427 thermally pretreated assay. The longer anaerobic digestion duration can be explained by less 428 solid and a smaller number of anaerobes present in the 20S:20L medium (Mishra et al., 2019; 429 Rouches et al., 2019). 430

Comparable to the untreated 40S:60L solid loading, the pretreated 40S:60L solid loading brought forth the best results in comparison to the other pretreated conditions in regard to the total biogas production, the methane purity in the biogas produced, methane yield as well as the treated effluent quality. From Figure 2, it can be observed that the total biogas produced by the pretreated 40S:60L was 1886.11±21.63 mL, which accounts for 415.01 mL more biogas produced than pretreated 20S:80L solid loading. Contrarywise to the untreated 40S:60L ratio,

the treated counterpart generated 454.44 mL more biogas. When comparing the 40S:60L of the 437 non-treated POME to that of the treated one as shown in Figure 6, it can be noted that the 438 439 amount of biogas produced in the pretreated experiments is substantially higher accounting for 456 mL more biogas generated than the untreated 40S:60L solid loading. The latter occurs 440 owing to readily accessible organic matter since the cell membrane holding the latter is 441 disintegrated when thermally pretreated. This occurrence in turns allows more bacteria to grow 442 443 and thrive, the MPN values obtained for pretreated 40S:60L solid loading was evaluated to be 4.6x10⁶ total anaerobes/100mL while the bacteria count for the untreated 40S:60L solid loading 444 was 2.1×10^6 total anaerobes/100mL which is slightly less than half the number of bacteria 445 present in the pretreated 40S:60L medium. 446

The anaerobic digestion duration for treated 50S:50L solid loading was determined to 447 be around 19 days, which is even shorter than both the treated 20S:80L and 40S:60L solid 448 loading along with all the untreated conditions. It can be observed that the maximum biogas 449 produced by the pretreated 50S:50L solid loading was 1509.43±4.43 mL, which was 450 approximately 186.71 mL more biogas produced than the untreated 50S:50L ratio and 377 mL 451 less than the pretreated 40S:60L solid loading. Subsequently, the anaerobic digestion of the 452 treated 50S:50L solid loading showed the removal efficiencies of COD, BOD, TSS, and O&G 453 decreased compared to that of the pretreated 40S:60L solid loading. The total number of 454 455 anaerobes present after anaerobic digestion in all the thermally treated conditions was higher 456 than that of the untreated ones. The reduced time for anaerobic digestion can be due to the implementation of the thermal treatment, which coincides with numerous other studies (Carrere 457 et al., 2016; Carrère et al., 2010). It can be noted that treated 20S:80L and 50S:50L solid 458 loadings attained particularly close total biogas production but not regarding the percentage 459 removal of COD, BOD, TSS, and O&G. Either condition can be made beneficial in terms of 460 either biogas production or treated effluent quality, depending on the mills' prerequisites. 461

The anaerobic digestion duration for pretreated 75S:25L solid loading was observed to 462 be shorter than the treated 50S:50L solid loading, which was around 15 days. The pH of the 463 system ranged from 7.20±0.02 to 7.71±0.06. The cumulative biogas achieved from pretreated 464 75S:25L solid loading was 1326.13±4.74 mL; this is less biogas produced compared to 465 pretreated 20S:80L, 40S: 60L and 50S:50L solid loadings but approximately 204 mL more 466 biogas produced in contrast to the untreated 75S:25L solid loading. As for the total number of 467 468 anaerobes, it was found that the untreated 75S:25L assay contained approximately 1.5×10^5 total thermally pretreated sample yielded 2.4×10^5 total anaerobes/100mL while the 469 470 anaerobes/100mL. However, the recorded values of total anaerobes are significantly lower than those present in both treated and untreated 20S:80L, 40S:60L, and 50S:50L solid loadings. 471

The data obtained for 100S exhibited a similar trend to that of 75S:25L solid loading. The total volume of biogas produced by the treated 100S solid loading was obtained as 970.00±2.89 mL⁻ which results in 137 mL more biogas than the untreated 100S solid loading. Another explanation for the reduction in biogas production at higher solid loadings is attributable to the decrease in water content and the associated least effective transport and mass transfer conditions whereby the microorganisms cannot thrive in an environment with limited soluble substrates, thus making the anaerobic digestion process less efficient (Le Hyaric et al., 2012;

479 Xu et al., 2014; Zhang et al., 2018).

480

3.3 Methane Composition in the biogas produced

For the 20S:80L solid loading, it is worthy to note that compared to the untreated 20S:80L solid loading, which produced the maximum methane at day 15, as shown in Figure 3, the thermally treated assays produced maximum methane by day 6, which is much earlier than the former as shown in Figure 4. The incidence above is considerably more favorable as the amount of biogas produced in the earlier days is distinctly higher, thus generating more methane compared to after day 10. The result obtained from the anaerobic digestion of pretreated 20S:80L solid loading was remarkable in contrast to its non-treated counterpart. The maximum methane
purity of the biogas in the pretreated samples reached up to 79.23±1.34%, which is slightly
more than the untreated 20S:80L solid loading.

The purity of methane in the biogas produced by the thermally treated 40S:60L 490 improved to 83.40±0.31% while the methane composition recorded for the untreated assay was 491 492 $77.33 \pm 1.20\%$ when comparing the maximum methane purity. The purity of methane produced in the pretreated 40S:60L solid loading was approximately 13% higher than that of the 493 untreated 40S:60L solid loading. It was also observed that at 24 hours after the setup of the 494 experiment, the percentage of methane produced was twice the amount of that produced by the 495 untreated sample. The rate of methane production was faster and more sustained for the 496 thermally pretreated 40S:60L experiments in comparison to the untreated one. 497

Though the methane purity for treated 50S:50L solid loading was found to be 76.97 \pm 0.73%, the untreated counterpart methane composition was measured to be 64.26 \pm 2.70%. It was observed that the methane composition for pretreated 20S:80L and 50S:50L are especially close, which further shows that the anaerobic digestion process for POME can be customized as per the requirements of the mill.

503 The methane composition for the untreated 75S:25L sample was measured to be 57.73±1.62%, while the treated sample had a lower methane composition of 51.97±2.03%. A 504 505 noteworthy observation distinguished in the pretreated 75S:25L solid loading was that the methane purity was moderately low, and the amount of hydrogen sulfide gas produced was 506 excessively elevated, ranging from 1503 to 1823 mg/L for the treated and untreated samples as 507 508 shown in Tables 1 and 2. This occurrence can be explained by the dissimilatory sulfate reduction, which is associated with the growth of sulfate-reducing bacteria (Choi et al., 2018). 509 The sulfate-reducing bacteria utilizes organic matters and sulfate as electron donors and an 510

electron acceptor, respectively. In anaerobic treatment, the methanogenesis can be
monumentally influenced by the sulfide produced from sulfate reduction (Yan et al., 2018). A
high concentration of free-hydrogen sulfide can lead to the inhibition of methanogenesis and,
eventually, a deficient anaerobic digestion process (Mizuno et al., 1998).

The higher production of hydrogen sulfide gas and the decline in methane composition were also observed with the 100S solid loading. The H₂S composition was noted to be around 2745 and 2968 mg/L for the treated and untreated samples. The maximum methane purity obtained was further inhibited in both the treated and non-treated assays, the pretreated 100S sample was recorded to be $38.20\pm0.75\%$ while that of the untreated counterpart was measured to be $33.17\pm1.30\%$.

521 3.4 Methane Yield

The methane yield is defined as the amount of methane generated for a specific amount of 522 523 organic matter removed, and this is the consequence of the activity of the anaerobic consortium of bacteria. The methane yield is constant during steady-state conditions for a particular carbon 524 substrate in anaerobic digestion conditions and is subjected to the fraction of the biodegradable 525 526 matter and the characteristics of the compounds present in the medium. In other words, the methane yield is constant when the anaerobic ecosystem uses carbon for growth and 527 conservation only. Theoretically, the methane yield of an anaerobic consortium is 350 mL 528 CH₄/g COD_{removed} when the anaerobic digestion is undertaken at room temperature. Reaching 529 this value would indicate that the bacteria would have used all the carbon available for 530 531 anaerobic digestion and growth (Michaud et al., 2002). However, the theoretical methane yield can be higher at thermophilic temperatures. 532

The methane yield calculated for the untreated and pretreated 20S:80L ratios were 36.20 and 313.18 mL CH_4/g $COD_{removed}$, correspondingly, which brings about a substantial 9-fold increase in the methane generated. While the methane yield evaluated for the non-treated and

the thermally pretreated 40S:60L ratios were 58.40 and 328.73 mL CH₄/g COD_{removed}, which is approximately 6-fold higher than the untreated assays. From Figure 3 and 4, it can be discerned that the maximum methane yield occurred between day 6 to day 8 for both the untreated and treated assays. It can be noted that the methane yield achieved for thermally pretreated 20S:80L and 40S:60L is relatively close to the theoretical value.

The methane yield for the treated assays was approximately 2-folds higher than the untreated 50S:50L assays. The total methane yield was evaluated to be 40.06 and 89.88 mL CH₄/g COD_{removed} for the untreated and the thermally treated 50S:50L ratios, respectively. The methane yield for both the treated and the untreated 50S:50L solid loadings are, however, lower than that obtained for the pretreated 40S:60L solid loading.

The methane yield for the untreated and thermally pretreated 75S:25L ratios were 27.84 and 54.06 mL CH₄/g COD_{removed}, correspondingly, which is also 2-fold superior to the untreated 75S:25L solid loading. Whereas, the methane yield was computed to be 16.69 and $31.52 \text{ mL CH}_4/\text{g COD}_{\text{removed}}$ for the untreated and the treated 100S assays, correspondingly. An important observation is that the methane yield for thermally pretreated 50S:50L, 75S:25L, and 100S all produced approximately 2-folds more methane than their untreated counterparts.

The difference in the anaerobic digestion performance between the thermally treated 552 and non-treated samples was prominent. It can be ascertained that the amount of biogas 553 554 produced, and the methane yield was notably enhanced in comparison to the non-treated 555 samples. The thermally pretreated conditions had more than twice better removal efficiencies when it comes to the COD, BOD, TSS, and O&G across all the different solids contents tested. 556 557 Another consistent occurrence across all the thermally pretreated conditions was that the anaerobic digestion process of POME was more stable when pretreated compared to those of 558 the untreated solid loadings. Based on Table 2, it can be seen that 40S:60L and 50S:50L ratios 559

achieved a more stable anaerobic digestion process as the percentage error in the cumulative biogas is significantly lower than 20S:80L, 75S:25L, and 100S solid loadings. In other words, the biogas production for all 9 assays for the ratios mentioned above is comparatively consistent. The above-described observation is in agreement with Farhat et al. (2018) study on thermally pretreated municipal sewage sludge and Toutian et al. (2020) investigation on the effect of temperature on the biogas yield of thermally hydrolyzed waste activated sludge.

The present study allows a means of control not only on the anaerobic digesters' load 566 but also on the results sought to be achieved. For example, if the industry aims to produce more 567 biogas with a high methane purity for steady electricity generation, in which case, it is more 568 favorable to use lower solid loadings. However, if the mill prefers to achieve a better-treated 569 effluent quality such that the treated solids can be used as fertilizers in the plantations, then a 570 higher solid loading is more suitable. Whereas for the treated liquid to conform to the 571 environmental standards to be discharged in watercourses, lower solid loadings are more 572 573 favorable. In a context where all the scenarios mentioned above need to be fulfilled, then using a solid loading close to 40S:60L will aid in attaining the desired outcomes. 574

With approximately 80% removal of COD, BOD, TSS, and O&G, the treated effluent 575 will require less time in the secondary treatment, which is the aeration treatment to sufficiently 576 remove the remaining COD, BOD, TSS, and O&G from the effluent. As the aeration treatment 577 578 is known to be a high energy consumption process, the lesser the time for which the effluent needs to be aerated, the higher the savings for the electrical energy required for the continuous 579 supply of oxygen. The substantial reduction in energy consumption is a primary advantage 580 581 when it comes to utility cost and hygienisation (Ahmad and Krimly, 2014; Khadaroo et al., 2019b). 582



585 Figure 6: Comparison between thermally pretreated and untreated POME cumulative biogas production graphs

586 587

4. Conclusion

588 Thermally pretreated POME achieved better biogas production, a higher purity along with methane yield, and the removal efficiencies for COD, BOD, TSS, and O&G are extensively 589 superior compared to the non-treated POME. The best performing condition was established 590 591 to be the thermally treated 40S:60L solid loading, which produced a maximum biogas production of 1886.11±21.63 mL, whereby the methane composition was measured to be 592 593 83.40±0.31%. The methane yield calculated for the thermally treated 40S:60L solid loading 594 was 328.73 mL CH₄/g COD_{removed}. As for the thermally treated 40S:60L, removal efficiencies in terms of COD, BOD, TSS, and O&G achieved was 80.63±0.46, 81.01±1.16, 80.72±0.16, 595 and 80.02±0.11% respectively. The pretreated POME provides a better treated effluent quality 596 such that when the treated effluent is sent for the secondary treatment, the latter can effectively 597 further reduce the COD, BOD, TSS, and O&G so that the treated effluent conforms to the 598 stipulated environmental standards. This study on pretreated solid loadings demonstrates that 599 the complex anaerobic digestion process can be made more versatile and customizable based 600 on the desired attainable results by allowing a means of control on the digesters' load. 601 Implementing the thermal pretreatment and a solid-liquid separation step in the process shows 602 remarkable potential in improving the treatment process of POME, all while producing a 603 significantly higher amount of biogas with an enhanced methane composition and improving 604 the treated effluent quality making the process sustainable and cleaner since it also notably 605 reduces the environmental encumbrance by allowing the use of the treated effluent as 606 fertilizers. 607

608

610 5. Acknowledgment

- 611
- The authors acknowledge financial support from Monash University Malaysia and the
 Royal Academy of Engineering under the Newton Research Collaboration Programme (Project
 NRCP1516/4/34).

615 Nomenclature

617	APHA	American Public Health Association
618	ASTM	American Society for Testing Material
619	BOD	Biological Oxygen Demand
620	COD	Chemical Oxygen Demand
621	СРО	Crude Palm Oil
622	L	Liquid
623	MPN	Most Probable Number
624	O&G	Oil and Grease
625	POME	Palm Oil Mill Effluent
626	S	Solid
627	TS	Total Solids
628	TSS	Total Suspended Solids
629	USEPA	United States Environmental Protection Agency
630	VS	Volatile Solids

- 631 6. References
- 632
- Ahmad, A., Krimly, M.Z., 2014. Palm Oil Mill Effluent Treatment Process Evaluation and
 Fate of Priority Components in an Open and Closed Digestion System. Curr. World
 Environ. 9, 321–330. https://doi.org/10.12944/CWE.9.2.12
- Ahmed, Y., Yaakob, Z., Akhtar, P., Sopian, K., 2015. Production of biogas and performance
 evaluation of existing treatment processes in palm oil mill effluent (POME). Renew.
 Sustain. Energy Rev. https://doi.org/10.1016/j.rser.2014.10.073
- APHA, 1998. Standard methods for the examination of water and wastewater. American
 Public Health Association. Inc., Washington, DC.
- 641 Appels, L., Baeyens, J., Degrève, J., Dewil, R., 2008. Principles and potential of the

- anaerobic digestion of waste-activated sludge. Prog. Energy Combust. Sci. 34, 755–781.
 https://doi.org/10.1016/j.pecs.2008.06.002
- Appels, L., Lauwers, J., Degrve, J., Helsen, L., Lievens, B., Willems, K., Van Impe, J.,
 Dewil, R., 2011. Anaerobic digestion in global bio-energy production: Potential and
 research challenges. Renew. Sustain. Energy Rev. 15, 4295–4301.
 https://doi.org/10.1016/j.rser.2011.07.121
- Ariunbaatar, J., Panico, A., Esposito, G., Pirozzi, F., Lens, P.N.L., 2014. Pretreatment
 methods to enhance anaerobic digestion of organic solid waste. Appl. Energy 123, 143–
 156. https://doi.org/10.1016/j.apenergy.2014.02.035
- Bala, J.D., Lalung, J., Ismail, N., 2014. Palm Oil Mill Effluent (POME) Treatment
 "Microbial Communities in an Anaerobic Digester": A Review. Int. J. Sci. Res. Publ.
 4, 1–24.
- Cappai, G., Gioaniis, G., Muntoni, A., Polettini, A., Pomi, R., Spiga, D., 2015. Effect of
 Inoculum To Substrate Ratio (Isr) on Hydrogen Production Through Dark Fermentation
 of Food Waste. Fifteenth Int. Waste Manag. Landfill Symp. S.
 https://doi.org/10.1007/s10329-014-0411-9
- Carrere, H., Antonopoulou, G., Affes, R., Passos, F., Battimelli, A., Lyberatos, G., Ferrer, I.,
 2016. Review of feedstock pretreatment strategies for improved anaerobic digestion:
 From lab-scale research to full-scale application. Bioresour. Technol. 199, 386–397.
 https://doi.org/10.1016/j.biortech.2015.09.007
- Carrère, H., Dumas, C., Battimelli, A., Batstone, D.J., Delgenès, J.P., Steyer, J.P., Ferrer, I.,
 2010. Pretreatment methods to improve sludge anaerobic degradability: A review. J.
 Hazard. Mater. 183, 1–15. https://doi.org/10.1016/j.jhazmat.2010.06.129
- Chin, M.J., Poh, P.E., Tey, B.T., Chan, E.S., Chin, K.L., 2013. Biogas from palm oil mill
 effluent (POME): Opportunities and challenges from Malaysia's perspective. Renew.
 Sustain. Energy Rev. 26, 717–726. https://doi.org/10.1016/j.rser.2013.06.008
- Choi, G., Kim, J., Lee, S., Lee, C., 2018. Anaerobic co-digestion of high-strength organic
 wastes pretreated by thermal hydrolysis. Bioresour. Technol. 257, 238–248.
 https://doi.org/https://doi.org/10.1016/j.biortech.2018.02.090
- Choong, Y.Y., Chou, K.W., Norli, I., 2018. Strategies for improving biogas production of
 palm oil mill effluent (POME) anaerobic digestion: A critical review. Renew. Sustain.
 Energy Rev. 82, 2993–3006. https://doi.org/10.1016/j.rser.2017.10.036
- Dohányos, M., Zábranská, J., Kutil, J., Jeníček, P., 2004. Improvement of anaerobic digestion
 of sludge. Water Sci. Technol. 49, 89–96.
- Farhat, A., Asses, N., Ennouri, H., Hamdi, M., Bouallagui, H., 2018. Combined effects of
 thermal pretreatment and increasing organic loading by co-substrate addition for
 enhancing municipal sewage sludge anaerobic digestion and energy production. Process
 Saf. Environ. Prot. 119, 14–22.
- 680 https://doi.org/https://doi.org/10.1016/j.psep.2018.07.013
- Heidrich, E.S., Curtis, T.P., Dolfing, J., 2011. Determination of the Internal Chemical Energy
 of Wastewater. Environ. Sci. Technol. 45, 827–832. https://doi.org/10.1021/es103058w
- Higgins, M.J., Beightol, S., Mandahar, U., Suzuki, R., Xiao, S., Lu, H.-W., Le, T., Mah, J.,
 Pathak, B., DeClippeleir, H., Novak, J.T., Al-Omari, A., Murthy, S.N., 2017.

685 686 687	Pretreatment of a primary and secondary sludge blend at different thermal hydrolysis temperatures: Impacts on anaerobic digestion, dewatering and filtrate characteristics. Water Res. 122, 557–569. https://doi.org/https://doi.org/10.1016/j.watres.2017.06.016
688	Iskandar, M.J., Baharum, A., Anuar, F.H., Othaman, R., 2018. Palm oil industry in South
689	East Asia and the effluent treatment technology—A review. Environ. Technol. Innov. 9,
690	169–185. https://doi.org/https://doi.org/10.1016/j.eti.2017.11.003
691 692 693	Jingura, R.M., Kamusoko, R., 2017. Methods for determination of biomethane potential of feedstocks: a review. Biofuel Res. J. 4, 573–586. https://doi.org/10.18331/BRJ2017.4.2.3
694 695 696 697	Khadaroo, S.N.B.A., Grassia, P., Gouwanda, D., Poh, P.E., 2020. The influence of different solid-liquid ratios on the thermophilic anaerobic digestion performance of palm oil mill effluent (POME). J. Environ. Manage. 257, 109996. https://doi.org/https://doi.org/10.1016/j.jenvman.2019.109996
698	Khadaroo, S.N.B.A., Grassia, P., Gouwanda, D., Poh, P.E., 2019a. Is the dewatering of Palm
699	Oil Mill Effluent (POME) feasible? Effect of temperature on POME's rheological
700	properties and compressive behavior. Chem. Eng. Sci. 202, 519–528.
701	https://doi.org/https://doi.org/10.1016/j.ces.2019.03.051
702 703 704 705	 Khadaroo, S.N.B.A., Poh, P.E., Gouwanda, D., Grassia, P., 2019b. Applicability of various pretreatment techniques to enhance the anaerobic digestion of Palm oil Mill effluent (POME): A review. J. Environ. Chem. Eng. 7, 103310. https://doi.org/https://doi.org/10.1016/j.jece.2019.103310
706	Lam Man & Lee, K.T., 2010. Renewable and sustainable bioenergies production from palm
707	oil mill effluent (POME): Win-win strategies toward better environmental protection.
708	Biotechnol. Adv. 29, 124–41. https://doi.org/10.1016/j.biotechadv.2010.10.001.
709	Le Hyaric, R., Benbelkacem, H., Bollon, J., Bayard, R., Escudié, R., Buffière, P., 2012.
710	Influence of moisture content on the specific methanogenic activity of dry mesophilic
711	municipal solid waste digestate. J. Chem. Technol. Biotechnol. 87, 1032–1035.
712	https://doi.org/10.1002/jctb.2722
713	Li, Y., Chen, Y., Wu, J., 2019. Enhancement of methane production in anaerobic digestion
714	process: A review. Appl. Energy 240, 120–137.
715	https://doi.org/https://doi.org/10.1016/j.apenergy.2019.01.243
716	Long, J.H., Aziz, T.N., Reyes, F.L.D.L., Ducoste, J.J., 2012. Anaerobic co-digestion of fat,
717	oil, and grease (FOG): A review of gas production and process limitations. Process Saf.
718	Environ. Prot. 90, 231–245. https://doi.org/10.1016/j.psep.2011.10.001
719	Martínez, E.J., Redondas, V., Fierro, J., Gómez, X., Morán, A., 2011. Anaerobic Digestion of
720	High Lipid Content Wastes:FOG Co-digestion and Milk Processing FAT Digestion. J.
721	Residuals Sci. Technol. 8, 53–60.
722 723 724	Michaud, S., Bernet, N., Buffière, P., Roustan, M., Moletta, R., 2002. Methane yield as a monitoring parameter for the start-up of anaerobic fixed film reactors. Water Res. 36, 1385–1391. https://doi.org/https://doi.org/10.1016/S0043-1354(01)00338-4
725 726 727	Mishra, P., Ameen, F., Zaid, R.M., Singh, L., Wahid, Z.A., Islam, M.A., Gupta, A., Nadhari, S. Al, 2019. Relative effectiveness of substrate-inoculum ratio and initial pH on hydrogen production from palm oil mill effluent: Kinetics and statistical optimization. J.

Clean. Prod. 228, 276–283. https://doi.org/https://doi.org/10.1016/j.jclepro.2019.04.317 728 729 Mizuno, O., Li, Y.Y., Noike, T., 1998. The behavior of sulfate-reducing bacteria in acidogenic phase of anaerobic digestion. Water Res. 32, 1626–1634. 730 https://doi.org/https://doi.org/10.1016/S0043-1354(97)00372-2 731 732 Müller, W.-R., Frommert, I., Jörg, R., 2004. Standardized methods for anaerobic biodegradability testing. Re/Views Environ. Sci. Bio/Technology 3, 141–158. 733 https://doi.org/10.1007/s11157-004-4350-6 734 Pilli, S., More, T., Yan, S., Tyagi, R.D., Surampalli, R.Y., 2015a. Anaerobic digestion of 735 thermal pre-treated sludge at different solids concentrations - Computation of mass-736 energy balance and greenhouse gas emissions. J. Environ. Manage. 157, 250–261. 737 https://doi.org/https://doi.org/10.1016/j.jenvman.2015.04.023 738 Pilli, S., Yan, S., Tyagi, R.D., Surampalli, R.Y., 2015b. Thermal Pretreatment of Sewage 739 Sludge to Enhance Anaerobic Digestion: A Review. Crit. Rev. Environ. Sci. Technol. 740 45, 669–702. https://doi.org/10.1080/10643389.2013.876527 741 Poh, P.E., Chong, M.F., 2010. Biomethanation of Palm Oil Mill Effluent (POME) with a 742 743 thermophilic mixed culture cultivated using POME as a substrate. Chem. Eng. J. 164, 146–154. https://doi.org/https://doi.org/10.1016/j.cej.2010.08.044 744 Poh, P.E., Yong, W., Chong, M.F., 2010. Palm Oil Mill Effluent (POME) Characteristic in 745 High Crop Season and the Applicability of High-Rate Anaerobic Bioreactors for the 746 Treatment of POME. Ind. Eng. Chem. Res. 49, 11732-11740. 747 748 https://doi.org/10.1021/ie101486w Rouches, E., Escudié, R., Latrille, E., Carrère, H., 2019. Solid-state anaerobic digestion of 749 750 wheat straw: Impact of S/I ratio and pilot-scale fungal pretreatment. Waste Manag. 85, 751 464-476. https://doi.org/10.1016/j.wasman.2019.01.006 Singh, R.P., Ibrahim, M.H., Esa, N., Iliyana, M.S., 2010. Composting of waste from palm oil 752 mill: A sustainable waste management practice. Rev. Environ. Sci. Biotechnol. 9, 331-753 344. https://doi.org/10.1007/s11157-010-9199-2 754 Soares, J.L., Cammarota, M.C., Gutarra, M.L.E., Volschan, I., 2019. Reduction of scum 755 accumulation through the addition of low-cost enzymatic extract in the feeding of high-756 757 rate anaerobic reactor. Water Sci. Technol. 80, 67-74. 758 https://doi.org/10.2166/wst.2019.247 Tajuddin, H.A., Abdullah, L.C., Idris, A., Choong, T.S.Y., 2015. Effluent Quality of 759 Anaerobic Palm Oil Mill Effluent (POME) Wastewater Using Organic Coagulant 4, 760 761 667-677. Toutian, V., Barjenbruch, M., Unger, T., Loderer, C., Remy, C., 2020. Effect of temperature 762 763 on biogas yield increase and formation of refractory COD during thermal hydrolysis of 764 waste activated sludge. Water Res. 171, 115383. https://doi.org/https://doi.org/10.1016/j.watres.2019.115383 765 Tyagi, V.K., Lo, S.L., 2011. Application of physico-chemical pretreatment methods to 766 enhance the sludge disintegration and subsequent anaerobic digestion: An up to date 767 768 review. Rev. Environ. Sci. Biotechnol. 10, 215-242. https://doi.org/10.1007/s11157-011-9244-9 769 Wood, N., Tran, H., Master, E., 2009. Pretreatment of pulp mill secondary sludge for high-770

- rate anaerobic conversion to biogas. Bioresour. Technol. 100, 5729–5735.
- 772 https://doi.org/10.1016/j.biortech.2009.06.062
- Xu, F., Wang, Z.-W., Tang, L., Li, Y., 2014. A mass diffusion-based interpretation of the
 effect of total solids content on solid-state anaerobic digestion of cellulosic biomass.
 Bioresour. Technol. 167, 178–185.
- 776 https://doi.org/https://doi.org/10.1016/j.biortech.2014.05.114
- Yan, L., Ye, J., Zhang, P., Xu, D., Wu, Y., Liu, J., Zhang, H., Fang, W., Wang, B., Zeng, G.,
 2018. Hydrogen sulfide formation control and microbial competition in batch anaerobic
 digestion of slaughterhouse wastewater sludge: Effect of initial sludge pH. Bioresour.
 Technol. 259, 67–74. https://doi.org/https://doi.org/10.1016/j.biortech.2018.03.011
- Zhang, E., Li, J., Zhang, K., Wang, F., Yang, H., Zhi, S., Liu, G., 2018. Anaerobic digestion
 performance of sweet potato vine and animal manure under wet, semi-dry, and dry
 conditions. AMB Express 8. https://doi.org/10.1186/s13568-018-0572-9